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Late spring phytoplankton bloom in the Lower St. Lawrence Estuary: the flushing hypothesis revisited

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> ABSTRACT: In the Lower St. Lawrence Estuary (LSLE), environmental conditions (stratification, surface light and nutrients) are favorable for phytoplankton growth starting in May, but the spring phytoplankton bloom typically does not occur until early summer (late June-July). Possible explanations for the late onset of the phytoplankton bloom include flushing of the surface layer due to the spring freshwater runoff, loss of phytoplankton cells from the thin euphotic layer through sinking and mixing, and temperature limitation of phytoplankton growth rates. We use 1- and 2-D time-dependent models of phytoplankton dynamics to explore these hypotheses. In particular, we illustrate the role of (1) phytoplankton cell sinking versus vertical turbulent mixing and (2) flushing of freshwater runoff on primary production in the LSLE. Results of the 1-D simulations show the dramatic effect of phytoplankton cell sinking in a thin euphotic zone, while at the same time high vertical turbulent mixing may act to maintain these sinking phytoplankton cells in the euphotic layer. Nevertheless, the 1-D analysis cannot account for spatio-temporal patterns in the development of the phytoplankton bloom observed during a high resolution physical, chemical and biological sampling field experiment performed in the summer of 1990 in the LSLE. 2-D simulations, run with seaward advective velocities in the range 0.15 to 0.3 m s⁻¹, close to observed values, generate downstream patterns of phytoplankton biomass that resemble these observed patterns. Comparison with observations helps to specify the range of sinking and advective velocities that operate in concert to control the timing and spatial location of the bloom.

> KEY WORDS: Phytoplankton bloom · Physical-biological coupling · Vertical turbulent mixing · Sinking · Flushing rates · St. Lawrence Estuary

INTRODUCTION

The Lower St. Lawrence Estuary (LSLE; Fig. 1) is a wide (30 to 50 km) and deep (300 m) estuary at the maritime end of a large (400 km long) and complex hydrodynamic system known as the St. Lawrence Estuary. The LSLE's main feature is the 300 m deep Laurentian Trough intruding from the Atlantic Ocean and ending abruptly at its head near Tadoussac, where the depth reduces from 300 to 50 m over less than

20 km. The high freshwater discharges from the St. Lawrence and Saguenay rivers (18 to $24 \times 10^3 \, \text{m}^3 \, \text{s}^{-1}$), coupled with active mixing at the confluence of these rivers at the head of the Laurentian Channel, lead to the continuous creation of a light water mass. Due to the width of the LSLE, which is several local internal Rossby deformation radii, this light water mass is subjected to geostrophic adjustment and the estuarine induced seaward flow takes the form of a jet-like structure that follows the south shore and amplifies downstream as the Gaspé current (Mertz et al. 1988). Physical conditions are also strongly dependent on atmospheric events at synoptic or seasonal time scales (Koutitonsky & Bugden 1991, Mertz et al. 1992) but it is

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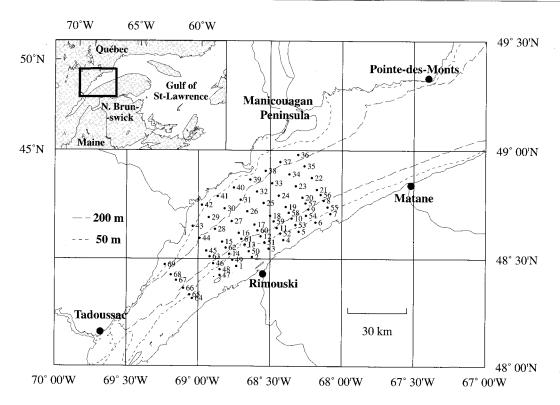


Fig. 1. Geographical location and detailed map of the Lower St. Lawrence Estuary. Dashed lines: 50 and 200 m isobaths, showing the location of the Laurentian Channel and extension of the shelf. Also shown are the locations of the physical and biological sampling stations of the COUPPB-90 experiment

now assumed that this buoyancy-induced circulation dominates the physics of the LSLE (Ingram & El-Sabh 1990, Koutitonsky & Bugden 1991).

These hydrodynamic conditions show a strong spatio-temporal variability that leads to marked variability in primary and secondary production (e.g. Therriault et al. 1990, de Lafontaine et al. 1991, Vézina 1994). Sporadic blooms of toxic algae (*Alexandrium* sp.) in the LSLE or in the Gaspé current may be due to particular hydrodynamic conditions (Therriault et al. 1985). Higher phytoplankton productivity or local zooplankton aggregations (Runge & Simard 1990) occur in relation to several types of oceanic fronts (Ingram & El-Sabh 1990). Studies of the coupling between physical and biological processes in the estuary and gulf are needed in order to understand the dynamics of harmful algae blooms and the carbon cycle, as well as the constraints on secondary and fish production in the region.

The spring phytoplankton bloom, a general feature in high and moderate latitudes, typically occurs in early summer (late June-July) in the LSLE even though environmental conditions (surface stratification, high light and nutrient concentrations) are favorable for phytoplankton growth in May, 4 to 8 wk earlier (e.g. Levasseur et al. 1984). The factors responsible for this delay are not well understood but are likely to be hydrodynamic. Many authors believe that the major features of the temporal (seasonal as well as synoptic time scales) distribution of phytoplankton biomass and production in estuaries and shelves is predominantly

related to physical processes (e.g. Pingree 1978, Malone et al. 1980, Sinclair et al. 1981, Powell et al. 1989, Lucas et al. 1998).

In this study, we use 1- and 2-D time-dependent models of phytoplankton dynamics to evaluate the relative roles of (1) vertical turbulent mixing versus phytoplankton cell sinking and (2) seaward advective transport associated with the freshwater runoff in determining the timing of this late spring phytoplankton bloom. The formulation and initialization of the models are defined with regards to physical, chemical and biological data obtained during the COUPPB-90 (COUPlage entre les processus Physiques et Biogeochimiques) mesoscale field experiment which covered the early summer 1990 bloom period (Savenkoff et al. 1997). By comparison with the COUPPB-90 data set, we show here that advective velocities in the LSLE and low sinking rates are consistent with the observed spatio-temporal patterns in the development of the phytoplankton bloom. This allows us to specify the range of sinking and advective velocities that operate in concert to control the timing and spatial location of the bloom.

FORMULATION OF THE MODELS

Background and working hypothesis. The numerical tool presented here is a process model which uses mean biological parameters found in the literature and focuses on the effect of physical processes on phyto-

plankton production. The simulations do not attempt to reproduce any particular event in the LSLE but attempt to capture the general conditions that affect bloom formation in the LSLE in order to explain its delay. The impact of short-term (less than 1 d) variability in the environmental conditions (semi-diurnal tidal mixing, diurnal light cycle) on the phytoplankton dynamics is not considered in this study. We assume that population development is more likely to depend on ecological conditions averaged over time scales greater than typical doubling times of phytoplankton cells (i.e. more than 1 d, see Zakardjian & Prieur 1994).

Likewise, the simulations do not take into account self-shading processes, which are not believed to impede the first stage of the bloom, hence the timing of the bloom. Assuming stationary physical forcing light, vertical turbulent mixing and seaward advection—the time-dependent vertical distribution of phytoplankton biomass (B), nitrate (NO₃) and ammonium (NH₄) concentrations are computed from the set of differential equations given in Table 1. Physical forcing processes (vertical turbulent mixing and advection) were defined from observations made during the COUPPB-90 mesoscale field experiments (see next section). Phytoplankton growth rate (μ) is formulated with a Liebig-type formulation (Eq. 7), assuming a hyperbolic saturation curve for photosynthesis (Eq. 9a) and using the O'Neill et al. (1989) substitutable model for nitrate and ammonium assimilation (Eq. 8b). Respiration is not explicitely included in the model; instead it can be considered to be a fraction of μ_{max} (e.g. Prieur & Legendre 1988), as it may be of little importance in initiating bloom development (Smetacek & Passow 1990).

The dynamics of micro- and mesozooplankton populations is not believed to affect the timing or the development of the bloom. Plourde & Runge (1993) have shown that the timing of final maturation and the spawning event of Calanus finmarchicus, a dominant copepod species in the LSLE, is strongly linked with the timing of the bloom. Grazing pressure from this population increases after the beginning of the spawning event, which proceeds throughout the summer, concurrent with relatively high phytoplankton biomass in the LSLE (e.g. Levasseur et al. 1984, Plourde & Runge 1993). While ciliated protozoans may play a

Table 1. System of differential equations used to simulate the spatial and temporal evolution of the state variables. The symbols are defined in Table 2

$$\frac{\partial B}{\partial t} + \nabla (\vec{U} \cdot B) = \nabla (K_{D} \cdot \nabla B) + \mu B - \text{Graz.}(B) + \frac{\partial}{\partial z} (w_{s}B)$$
(1)

$$\frac{\partial NO_3}{\partial t} + \nabla (\vec{U} \cdot NO_3) = \nabla (K_D \cdot \nabla NO_3) - S_{NO_3} \mu B$$
 (2)

$$\frac{\partial NH_4}{\partial t} + \nabla(\vec{U} \cdot NH_4) = \nabla(K_D \cdot \nabla NH_4) - S_{NH_4} \mu B + \text{Reg.Graz.}(B)$$
(3)

Exp. =
$$\int_{z=0}^{zb} (1 - \text{Reg.}) \, \text{Graz.}(B) \delta z$$
 (4) $S_{\text{NH}_4} + S_{\text{NO}_3} = 1$ (5)

Graz.(B) =
$$\overline{Z}G_{\max}\{1 - \exp[\beta(B - S_b)]\}H(B, S_b)$$
 (6) $\mu = \frac{\ln(2)}{\max[t_c, t_s]}$ (7)

$$t_{\rm s} = \frac{t_{\rm d\,min}}{f[N]} \qquad \qquad f[N] = \frac{{\rm NO_3}K_{\rm NH_4} + {\rm NH_4}K_{\rm NO_3}}{K_{\rm NO_3}K_{\rm NH_4} + {\rm NH_4}K_{\rm NO_3}}$$
(8b)

$$S_{\text{NO}_3} = \frac{\text{NO}_3 K_{\text{NH}_4}}{\text{NO}_3 K_{\text{NH}_4} + \text{NH}_4 K_{\text{NO}_3}}$$
(8c)
$$S_{\text{NH}_4} = \frac{\text{NH}_4 K_{\text{NO}_3}}{\text{NO}_3 K_{\text{NH}_4} + \text{NH}_4 K_{\text{NO}_3}}$$
(8d)
$$t_c = t_{\text{d} \min} \left(1 + \frac{K_E}{E} \right)$$
(9a)
$$E = E_0 \exp(-\chi z)$$
(9b)
$$K_z = 0.25 \frac{\varepsilon}{N^2}$$
(10)

$$t_{\rm c} = t_{\rm d\,min} \left(1 + \frac{K_{\rm E}}{E} \right) \tag{9a}$$

$$E = E_0 \exp(-\chi z)$$

$$K_z = 0.25 \frac{\varepsilon}{N^2} \tag{10}$$

Table 2. List of symbol and values of constants used in the simulations (Var. denotes variables calculated by the model as opposed to imposed variables and parameters). PAR: photosynthetically active radiation

Symbol	Value/units	Definition
В	Var. mmol N m ⁻³	Phytoplankton biomass in terms of nitrogen
β	$0.35 \; \mathrm{mmol} \; \mathrm{N}^{-1} \; \mathrm{m}^{3}$	Ingestion coefficient for the Ivlev-type grazing function
χ	$0.3 \; \mathrm{m}^{-1}$	Light extinction coefficient
δt	1.0 h	Time step of the simulation
δx	5 km	Horizontal increment of the 2-D Eulerian grid
δz	1.0 m	Vertical increment of the Eulerian grid
E	Ein $m^{-2} d^{-1}$	Mean daily available PAR at each depth
E_0	$30.0 \text{ Ein m}^{-2} d^{-1}$	Mean daily surface irradiance (PAR)
Exp.	Var. mmol N $m^{-2} d^{-1}$	Biologically exported organic matter
2	$0.5 - 5.0 \ 10^{-7} \ m^2 \ s^{-3}$	Dissipation rate of turbulent kinetic energy
$G_{ ext{max}}$	$0.5 \ d^{-1}$	Maximum grazing rate for the Ivlev-type function ^a
Graz.(B)	Var. mmol N $m^{-3} d^{-1}$	Ivlev-type grazing function
$H(B, \hat{S_{b}})$	Var., dimensionless	Heavyside function associated with the feeding threshold
$K_{\rm D}$	$\mathrm{m^2d^{-1}}$	General 3-D diffusivity tensor
X_z	$\mathrm{m^2~d^{-1}}$	Vertical turbulent diffusivity coefficient
$\tilde{K_{\rm E}}$	10.0 Ein m ⁻² d ⁻¹	Half saturation constant for the light versus photosynthesis relationship
K _{NH}	0.1 mmol m^{-3}	Half saturation constant for the Michaelis-Menten kinetics of NH ₄ uptake
K_{NO_3}	0.5 mmol m ⁻³	Half saturation constant for the Michaelis-Menten kinetics of NO ₃ uptake
1 103	Var. d ⁻¹	Phytoplankton growth rate
V	s^{-1}	Brunt-Väisälä frequency
NH₄	Var. mmol N m ⁻³	Ammonium concentration
NO_3	Var. mmol N m ⁻³	Nitrate concentration
Reg.	0.4, dimensionless	Fraction of grazed nitrogen which is excreted by zooplankton as NH ₄
$S_{\rm b}$	0.1 mmol N m ⁻³	
$S_{\rm NH_4}$	Var., dimensionless	Feeding threshold for the Ivlev-type grazing function
NO ₃	Var., dimensionless	Fraction of phytoplankton growth sustained by NH ₄
'NO ₃	Var. d	Fraction of phytoplankton growth sustained by NO ₃
c crit	8.4 d	Phytoplankton doubling time in carbon biomass
crit d min	0.5 d	Maximum phytoplankton doubling time: if $t_c(z)$ or $t_s(z,t) > t_{crit}$, then $\mu(z,t) = t_{crit}$
a min	Var. d	Minimum phytoplankton doubling time
j J	m s ⁻¹	Phytoplankton doubling time in nitrogen biomass
$U_{\rm mes}$	$m s^{-1}$	General 3-D velocity vector
$J_{\rm s}$ $U_{ m mes}$	m s ⁻¹	Measured along-estuary current velocity
	0-3 m d ⁻¹	Surface current velocity
₩ _s	1.2 mmol N m ⁻³	Sinking rate of phytoplankton cells Mean grazer zooplankton biomass ^b

^bConsidering higher values in the range of ciliate and copepod abundances in the LSLE and Eastern Canadian waters in late spring/early summer, e.g. Runge & Simard (1990), de Lafontaine et al. (1991), Sime-Ngando et al. (1995)

major role in the trophic links between phytoplankton and mesozooplankton in the LSLE, it is unlikely that they impede the development of the bloom (e.g. Sime-Ngando et al. 1995). In the model, grazing pressure (Graz.) is simply defined by an Ivlev-type function (Eq. 6), assuming a mean constant grazer biomass (meso- and microzooplankton) taken from the literature (see references at the bottom of Table 2) for LSLE and Eastern Canadian waters in late spring/early summer. This means zooplankton biomass was unable to prevent the diatom bloom, or a 2-fold higher biomass, as described later.

To take into account nitrogen regeneration, a constant fraction (0.4) of the grazed phytoplankton nitrogen is regenerated in terms of ammonium (e.g. Frost 1980, Kiørboe 1989). The fraction of grazed and nonregenerated nitrogen is taken as exported production

(Exp. in Eq. 4, Table 1). Nutrients and phytoplankton biomass are computed in mmol N m⁻³. Phytoplankton biomass and production terms are then converted to mg chl a m⁻³ and mg C m⁻³ d⁻¹, respectively, to facilitate comparisons with field data. Conversion in terms of carbon is made through the Redfield molar ratio 106/16 for C/N and for chl a using the weight ratio 1/80 for chl a/C, which would be representative of the diatom population in the pre- and early bloom period (mainly *Nitzschia* sp., see Sime-Ngando et al. 1995).

The range of sinking rates tested in the simulations is 0 to 3 m d⁻¹, typical of individual sinking diatoms (Smayda 1970). Diatom chains may sink faster and flocculation of diatom blooms can lead to sinking of aggregated growing cells at speeds of several tens of meters per day (Alldredge & Jackson 1995). We assumed, however, that the coupled flocculation and aggregation processes are of little importance in the early stage of the bloom, hence for its timing, but may play a major role in the later stages of the bloom (e.g. Jackson & Lochman 1992).

Physical processes: the COUPPB-90 data set. The COUPPB-90 experiment consisted of 3 consecutive cruises (Legs G2, G3 and G4) that covered the 50 station grid shown in Fig. 1 each day for 3 d. The experiment was conducted between 30 June and 9 July and covered half a neap-spring tidal cycle (from neap to spring). At each station, sampling included temperature, salinity, oxygen and turbidity profiles (surface to 5 to 10 m off bottom) from an Applied Microsystem STD-12; vertical profiles of horizontal currents from a 150 kHz Acoustic Doppler Current Profiler (ADCP); and 0 to 30 m continuous profiles of photosynthetic pigments (in vivo fluorescence of chl a) and nutrients (nitrate and silicate). Several additional biogeochemical and physical measurements were conducted at 3 stations located in the Laurentian Trough. In addition, Aanderaa current meters were moored at 7 locations between May and September. Details of these experiments can be found in Savenkoff et al. (1997) and Marsden & Gratton (1998).

Briefly, this study showed that, in 1990, the onset of the late spring phytoplankton bloom in the LSLE is related to the formation of transverse currents generated by geostrophic adjustments during the passage of freshwater pulses originating from the head of the Laurentian Channel. Conditions prior to the bloom featured a well-defined, along-shore alignment of thermohaline properties and a prominent seaward residual current. The switch from this prevailing situation to the transverse current configuration took place in a short time, i.e. 3 to 5 d. The circulation patterns after the passage of these pulses showed complex features at various scales (small to mesoscale) that superseded the conspicuous seaward residual currents previously observed. The stratification of the euphotic layer (12 m depth) increased from Legs G2 to G4 as the result of a decrease in salinity in the upper layer due to the passage of the freshwater pulse through the study area (Savenkoff et al. 1997). Objective analysis has shown that the observed physical and biological structures were unrelated to the phases of the semi-diurnal tidal cycle. Detailed presentation and discussion of these results can be found in Savenkoff et al. (1997).

In each simulation, the vertical diffusion coefficient K_z was estimated from the mean Brunt-Väisälä frequency profile (N) from Leg G2 (Fig. 2a,b) using the Osborn (1980) formulation (Eq. 10, Table 1). The mean N profile was calculated as a Chebyshev interpolation of the complete Leg G2 data set using the trend1d routine of the GMT package (Generic Mapping Tools, Wessel & Smith 1991), as for all other Chebyshev inter-

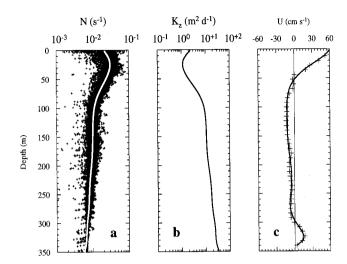


Fig. 2. Physical conditions used in the simulations: (a) vertical distributions of observed Brunt-Väisälä frequencies (N) during Leg G2 and the Chebyshev polynomial used in the simulations to define the mean stratification (solid line); (b) eddy diffusion coefficient (K_z) calculated from the mean stratification with a turbulent kinetic energy (TKE) dissipation rate (ε) of 5.0×10^{-8} m³ s⁻¹; (c) Acoustic Doppler Current Profiler measured vertical profile of the along-channel baroclinic velocity (U) obtained at Stn 23 during Leg G2 (+) and the fitted Chebyshev polynomial used in the simulations (solid line)

polations in this study. This mean N profile is considered to be representative of the pre-bloom stratification conditions.

In the LSLE, the head of the Laurentian Channel is an active zone of mixing where high turbulent mixing rates are due to the large shear associated with the generation of internal tides or travelling solitons, both being detectable as far as 100 km downstream of the generation area (e.g. Forrester 1974, Galbraith 1992). While these high turbulent mixing rates are typical of the particular hydrodynamic conditions at the head of the Laurentian Channel, we can expect similar, but attenuated, mixing processes over a large part of the LSLE (Gratton et al. 1988). We used different values for the turbulent kinetic energy (TKE) dissipation rate $(\varepsilon = 5.0 \times 10^{-8} \text{ to } 5.0 \times 10^{-7} \text{ m}^2 \text{ s}^{-3})$ to define several vertical turbulent mixing regimes for this single N profile. These are test simulations that do not consider the response of stratification to the imposed mixing.

Current velocities used in the 2-D model are derived from a Chebyshev interpolation of the baroclinic velocities (ADCP measured minus vertical mean) at Stn 23 (Fig. 2c) during Leg G2. We selected this station because it showed the highest seaward surface velocity (0.6 m s⁻¹) and surface baroclinicity (0.6 m s⁻¹ over 50 m). Stn 23 was visited during high tide slack water during Leg G2. Strong baroclinicity in the euphotic layer may have a marked effect on the vertical distributions of chl *a*, nutrients and dissolved oxygen (Savenkoff et al.

1997). Note that below 50 m, the current reverses and flows slowly upstream. Analogous to the classical estuarine circulation, this 2-layer circulation in the Laurentian Trough is typical of the LSLE (e.g. El-Sabh 1979, Koutitonsky & Bugden 1991).

Nevertheless, the high surface baroclinic velocities at Stn 23 are the result of a point measurement and may not be representative of the mean residual seaward advection of the buoyancy-induced circulation. They represent rather an upper bound to the associated surface current. The seaward, buoyancy-induced circulation in the LSLE in late spring shows residual currents of the order 0.3 m s⁻¹ (e.g. El-Sabh 1979, Mertz et al. 1988). Other 2-D simulations made with values one-half to a quarter of the values in the measured baroclinic velocity profile at Stn 23 were assumed to be more typical of the seaward residual component of the springtime buoyancy-induced circulation in the LSLE.

Numerical techniques. Numerical techniques are already described in Zakardjian & Prieur (1994) and are briefly reviewed here. Both models extend vertically to 150 m and the 2-D version has a horizontal domain of 200 km (approximately from the head of the Laurentian Channel, near Tadoussac, to Pointe-des-Monts; see Fig. 1). Except for phytoplankton nitrogen. initial conditions are defined as the Chebyshev interpolations of vertical distributions observed during the COUPPB-90 experiment. For nitrate and ammonium we used the complete sets of vertical distributions (Stns 23, 26 and 67 and Legs G2 to G4) from bottle casts (Fig. 3). The initial phytoplankton nitrogen vertical distribution represents a constant low phytoplankton biomass (0.6 mmol N m⁻³) over the first 30 m, decreasing exponentially below this.

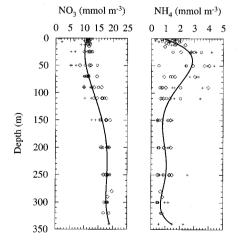


Fig. 3. Vertical profiles used as initial and lateral boundary conditions for nitrate (NO₃) and ammonium (NH₄). Solid lines: Chebyshev polynomials fitted to observed values for the 3 legs, G2 to 4 (\$\displaystyle \text{: Stn 67, +: Stn 23, o: Stn 26})

Vertical boundary conditions are: no fluxes at the sea surface (Neuman condition with vertical derivatives set to zero for all variables) and constant concentrations (Dirichlet condition), for the bottom boundary. Constant concentrations at the bottom are calculated from the Chebyshev polynomials at 150 m depth. Upstream and downstream boundaries in the 2-D model are open (Dirichlet conditions), with constant vertical distributions as the initial conditions. Observations made during the COUPPB-90 experiment were consistent with time-independent upstream conditions. The downstream boundary conditions have no impact on the simulations because the flow in the surface layer is downstream.

We solved the set of differential equations Eqs. (1) to (4) (Table 1) in a discrete form (semi-implicit Crank-Nicolson scheme derived from the Control Volume Approach, see Roach 1972) with Choleski's double scanning method (also called Thomas algorithm in Roach 1972) in an iterative form (because of the non-linear grazing and phytoplankton growth terms) and with the method of alternate directions in the 2-D model.

RESULTS

The standard simulation as an illustration of a classical spring bloom

The standard 60 d simulation using parameters found in the literature (see Table 2), non-sinking phytoplankton and a low value for the TKE dissipation rate ($\varepsilon = 5.0 \times 10^{-8} \,\mathrm{m}^2 \,\mathrm{s}^{-3}$) led to a strong bloom that appears first as a rapid increase in depth-integrated phytoplankton biomass and primary production (Fig. 4a) within the first 5 d. Maximum values of ca $2.2 \text{ g C m}^{-2} \text{ d}^{-1}$ for primary production and 110 mg chl a m⁻² for phytoplankton biomass were reached at Days 5 and 13 respectively. In the surface layer, chl a concentrations reached maximum values of ca 9 to 11 mg m⁻³ during the bloom and the phytoplankton exhausted surface nitrogen nutrient (nitrates and ammonium in the first 5 m) after less than 5 d. Later, grazing acted to decrease phytoplankton nitrogen and the system slowly tended to a steady state with a slight deep chlorophyll maximum (DCM) which developed from the deepening nutricline down to the bottom of the euphotic zone. After 60 d, the depth-integrated phytoplankton nitrogen and primary production were low, ca 18.0 mg chl a m⁻² and 150 mg C m^{-2} d^{-1} respectively (Fig. 4a). These levels fall within the range of values for the spring phytoplankton bloom and post-bloom conditions in the LSLE (Levasseur et al. 1984, Therriault & Levasseur 1985, Vézina 1994, Vézina et al. 1995, Savenkoff et al. 1997).

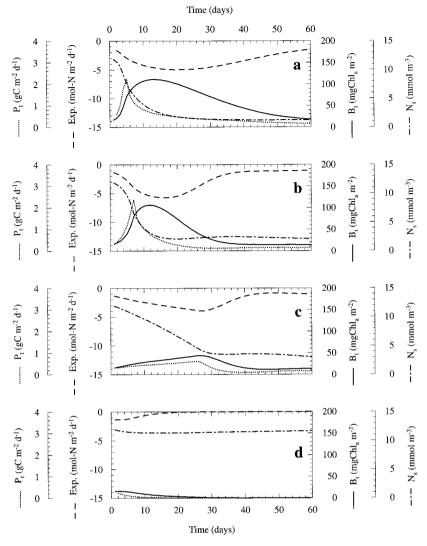


Fig. 4. Temporal evolution of depth-integrated primary production (P_n 0–15 m), chl a biomass (0–15 m) and exported production (Exp., 0–150 m) and surface nutrients ($N_{\rm s}={\rm NO_3+NH_4}$ averaged over the euphotic zone) for the 60 d simulations made with a low TKE dissipation rate ($\epsilon=5.0\times10^{-8}$ m² s⁻³) and phytoplankton sinking rates of (a) 0, (b) 1, (c) 2 and (d) 3 m d⁻¹

The lag between the primary production and biomass maxima (Fig. 4a) is related in part to the imposition of constant grazing pressure, which uncouples changes in phytoplankton and grazer biomass. The lag is also related to regeneration processes that acted to retain some of the previously assimilated nitrogen in the surface layer. Note that the time scale of mesozooplankton population dynamics (typically 2 to 3 wk) is greater than the 1 wk lag between the higher primary production and higher phytoplankton biomass. Thus, except for the final stage, which did not include a response of grazer populations, this temporal sequence is typical of the train of events during a phytoplankton bloom. This first simulation shows that standard nutri-

ent, light and stratification conditions prevailing before and during the COUPPB-90 experiment are favorable for a rapid bloom which would affect the whole LSLE.

Vertical turbulent mixing versus sinking

Application of higher values for the TKE dissipation rate ($\epsilon \approx 1$ to 5×10^{-7} m² s⁻³) did not significantly change the temporal evolution of the phytoplankton bloom. The bloom developed more slowly (less than 10 d) but yielded higher maximum and steady-state primary production and phytoplankton biomass (Fig. 5a). The computed depth-integrated phytoplankton biomass and production remained typical of a phytoplankton coastal or estuarine bloom (170 mg chl a m⁻² and 3.5 g C m^{-2} d^{-1}). The surface layer was rapidly and permanently depleted of nutrients, while the vertical distribution of phytoplankton biomass did not show subsurface maxima as in the preceding simulation (Fig. 6). This indicates that vertical turbulent mixing, under the stratification conditions imposed in the simulation, is not a limiting factor for phytoplankton growth, i.e. there is not sufficient turbulent leakage of phytoplankton cells out of the euphotic zone in the sense described by Lucas et al. (1998).

In contrast, phytoplankton cell sinking has a more pronounced effect on the dynamics of phytoplankton. With a sinking rate of 1 m d⁻¹, the bloom was delayed slightly and yielded maximum values for depth-integrated new production

and phytoplankton biomass similar to those in the preceding simulation run with the same ϵ value. Phytoplankton biomass decreased faster after the bloom and was lower in the steady-state solution (Fig. 4b). Fig. 6 shows the rapid sinking of the DCM formed during the bloom. With a sinking rate of 2 m d⁻¹, the temporal changes in integrated phytoplankton biomass, primary production and export were greatly reduced. Maximum values of ca 45 mg chl a m⁻² and 610 mg C m⁻² d⁻¹ for phytoplankton biomass and primary production were reached after 26 d (Fig. 4c). Nutrients in the surface layer slowly decreased during the first half of the simulation but were not totally depleted after 60 d. Under low vertical turbulent mixing conditions, a 3 m

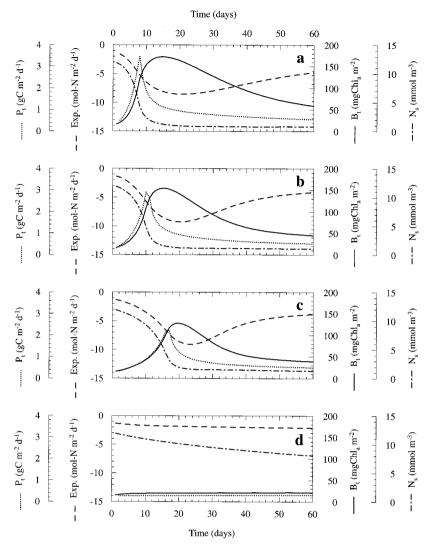


Fig. 5. Temporal evolution of depth-integrated primary production (P_n , 0–15 m), chl a biomass (0–15 m) and exported production (Exp., 0–150 m) and surface nutrients ($N_{\rm s}={\rm NO_3+NH_4}$ averaged over the euphotic zone) for the 60 d simulations made with a high TKE dissipation rate ($\epsilon=5.0\times10^{-7}$ m² s⁻³) and phytoplankton sinking rates of (a) 0, (b) 1, (c) 2 and (d) 3 m d⁻¹

 d^{-1} sinking rate removed all the phytoplankton biomass in less than 30 d (Fig. 4d).

A stronger vertical turbulent regime (larger ϵ) restored the bloom at a sinking rate of 2 m d⁻¹ (Fig. 5c) and led to constant low phytoplankton biomass and primary production at higher sinking rates (Fig. 5d). This is because vertical turbulent mixing can partially oppose sinking by redistributing phytoplankton cells in the euphotic layer (see Fig. 6). This result agrees with the Lagrangian study of Lande & Wood (1987) and the shallow estuary case described by Lucas et al. (1998). Nevertheless, values of phytoplankton biomass, primary production and nitrogen nutrients were not representative of a phytoplankton bloom when the

sinking rate was greater than 2 m d⁻¹, i.e. no bloom was simulated under such conditions. Note the long delay (more than 15 d) before the onset of the phytoplankton bloom under conditions of high vertical turbulence and a 2 m d⁻¹ sinking rate (Fig. 5c). Lastly, the lag between primary production and phytoplankton biomass maxima was reduced as the sinking rate increased (Figs. 4b,c & 5b,c). This is because sinking removed phytoplankton biomass from the euphotic zone and allowed more regeneration at depth, i.e. below the euphotic zone.

Flushing by the freshwater runoff: a 2-D approach

Results of the 2-D simulation using the observed baroclinic velocity profiles at Stn 23 (Fig. 2c) are shown in Fig. 7. The simulated phytoplankton-nutrient system rapidly achieved a steady state (in less than 5 d) that generated slightly increasing phytoplankton biomass (from 14 to 46 mg chl a m⁻²) and slightly decreasing surface nutrient concentration in the downstream portion (150 to 200 km) of the simulated area. The rapid steady state achieved in the 2-D simulations results from the residence time of phytoplankton biomass in the simulated area, i.e. the phytoplankton-nutrient system developed up to time $t = (200 \text{ km})/U_s$ (approximately 4 d for the corresponding 1-D simulation). Depth-integrated primary production showed a similar but more pronounced seaward increase with maximum values ca $2.3~g~C~m^{-2}~d^{-1}$ at the downstream boundary of the simulated area.

The simulation made using half the observed baroclinic velocity ($U=\frac{1}{2}U_{\rm mes}$) led to a slower attainment of steady state (in approximately 10 d) with more marked seaward increases in phytoplankton biomass and decreases in surface nutrient concentrations in the downstream half of the section (Fig. 8). The depth-integrated primary production shows a zone of maximum values (ca 2.8 g C m $^{-2}$ d $^{-1}$) 110 km downstream which corresponds to the beginning of the nutrient-impoverished area. In contrast, the maximum values for depth-integrated biomass, of ca 100 mg chl a m $^{-2}$, were still at the end of the simulated estuary.

This spatial uncoupling between primary production and maximum phytoplankton biomass was partly due

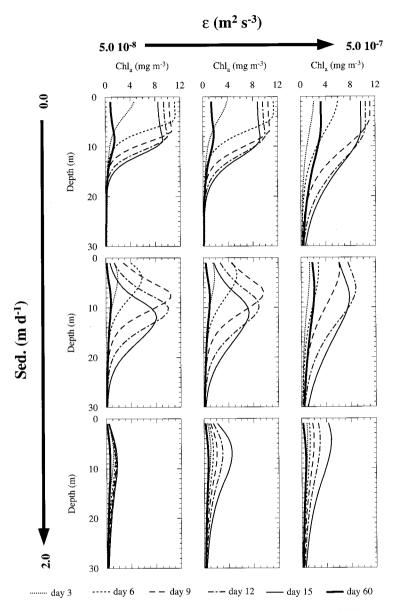


Fig. 6. Temporal evolution of the simulated vertical profiles of phytoplankton biomass for the first 15 d and for various regimes of vertical turbulent mixing (ϵ ranging from 5.0×10^{-8} to 5.0×10^{-7} m 3 s $^{-1}$) and sinking rates (Sed. varying from 0 to 2 m d $^{-1}$); bold solid line in each panel: quasi steady-state vertical profile after 60 d

to the temporal lag previously described in the 1-D simulation (e.g. Fig. 4a), which appears as a spatial uncoupling due to the advection of the whole phytoplankton-nutrient system. Note that the time needed to achieve the larger primary production rates in the 1-D simulation was ca 5 d, which corresponds to a distance of 130 km, close to the highest primary production zone in Fig. 8. Nevertheless, the depth-integrated daily assimilation ratios (as an index of these biomass-production uncouplings) were nearly twice as large in the 2-D simulation [ca 56 g C (g chl a)⁻¹ d⁻¹ in the high-

est primary production zone with $U = \frac{1}{2}U_{\text{mes}}$] than in the 1-D standard simulation [ca 32 g C (g chl a)⁻¹ d⁻¹ at the time of highest primary production]. This implies that this spatial uncoupling also resulted partly from rapid advection of the phytoplankton biomass out of the high primary production zone.

Comparisons with the COUPPB-90 observations

Fig. 9 shows the depth-integrated (0 to 15 m) phytoplankton biomass along the COUPPB-90 area from Legs G2 to G4. Note first the consistently low phytoplankton biomass upstream and a general tendency to a downstream increase in agreement with the 2-D simulations. For Leg G2, the along-shore increase is small and, except for 3 near-shore stations, bloom-like biomasses are restricted to the last third of the area. For Leg G3, the data suggest a biomass maximum near the middle of the section, while phytoplankton biomass remains high downstream. This maximum probably reflects the influence of the strong transverse current described by Savenkoff et al. (1997). This transverse current reduces the downstream residual advection and may allow the bloom to build up farther upstream. Leg G4 shows a similar tendency as Leg G3, but with a higher spatial variability. The more complex circulation patterns observed during Leg G4 are likely to explain this higher variability about the trend expected from the model simulations. The COUPPB-90 area starts 70 km downstream of the head of the Laurentian Channel and is 70 km long. Observed values are well bounded by the simulated depth-integrated biomass between 70 and 140 km for seaward advections of ca 0.3 to 0.6 m s⁻¹ (Fig. 9).

The along-channel variability of the vertical profiles of chl *a* (Fig. 10) confirms the above findings but we noted the rapid occurrence of DCMs as the bloom develops downstream.

Figs. 5 to 7 in Savenkoff et al. (1997) show that the surface inflow from the Gulf was negligible during COUPPB-90, so that the observed patterns in the vertical profiles of chl a would have been produced locally. Such DCMs can originate from 2 main processes: (1) increase at depth of the chl a/C ratio due to adaptation to low light and (2) development of a deep biomass maxima following the nutrient impoverishment of the surface layer. The first process is unlikely to occur in shallow and turbulent euphotic layers as is the case here. The second process is the long-term normal out-

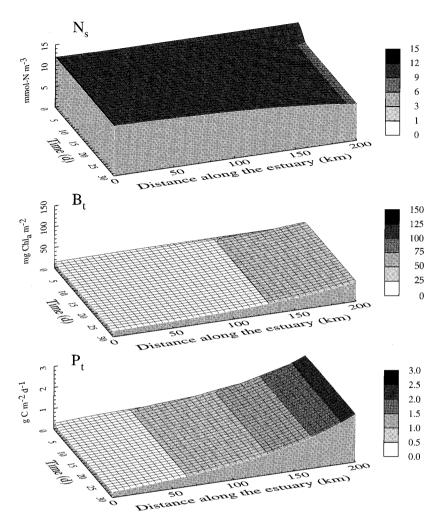


Fig. 7. Temporal evolution along the simulated estuary of surface nutrients $(N_s = NO_3 + NH_4)$ averaged over the euphotic zone), depth-integrated phytoplankton biomass (B_t) and primary production (P_t) obtained with the baroclinic velocity profile observed at Stn 23 (maximum surface current of ca $0.6~\mathrm{m~s^{-1}}$) during Leg G2 of the COUPPB-90 experiment

come of the bloom but it is more rapid for slowly sinking cells (Fig. 6). Hence, to perform comparisons with the simulated vertical profiles, we have to define a mean residence time for the phytoplankton biomass in the COUPPB-90 area. Estimating a mean seaward advection is difficult due to the rapid changes in the COUPPB-90 observed circulation patterns. However, Savenkoff et al. (1997) estimated the velocity of the pulse-like feature which led to this circulation variability to be ca 0.15 m s^{-1} . Taking this value as a background seaward velocity leads to a residence time of 5.5 d in the COUPPB-90 area and 11 d if we include the distance from the head of the Laurentian Channel. It can be seen from Fig. 6 that only a slowly sinking phytoplankton population may be able to produce DCMs in such short times.

In more detail, observed vertical profiles of chl a during Leg G2 (Fig. 10) may correspond to the simulated ones for the first 3 to 7 d of a non- or slowly sinking population ($w_s = \text{ca } 0 \text{ to } 1 \text{ m d}^{-1}$), coherent with a mean seaward advection of ca 0.25 m s^{-1} and higher (Fig. 11). For Leg G3, observed profiles at the end of the COUPPB-90 area are comparable to simulated ones for a slowly sinking population ($w_s = 1.5 \text{ m d}^{-1}$) of ca 4 to 8 d, leading to a corresponding seaward advection of ca 0.20 m s⁻¹ (Figs. 10 & 11). For Leg G4, the same reasoning gives 10 d and higher for residence time and a corresponding seaward advection of ca 0.15 m s^{-1} (Figs. 10 & 11).

The approach we used cannot capture all the possible short-term and smallscale processes that may affect the phytoplankton distribution and lead to DCMs, such as a local divergence of the turbulent flux or near-inertial wave shears (e.g. Franks 1994), subduction of the chl a-rich layer due to cross-frontal secondary circulation or, as discussed by Savenkoff et al. (1997), overflow of light, nutrient-rich and low-chl a waters on chl a-rich, low-nutrient surface water, so that the simulated DCMs show some discrepancies with the observed ones. Nevertheless, our approach should be able to reproduce the long-term trends of phytoplankton dynamics (e.g. Zakardjian & Prieur 1994). Moreover, the general tendency of the DCM to deepen from Legs G2 to G4 agrees with the scenario of a slowly sinking population, and the persistence of low chloro-

phyll values in the upstream part of the COUPPB-90 area (Figs. 9 & 10) agrees with the flushing scenario. Hence, the residence time of the surface waters and the simulated bloom development time match sufficiently well to explain the along-channel patterns in phytoplankton biomass observed during the COUPPB-90 experiment.

DISCUSSION

The turbulent mixing hypothesis

Results presented in this study proceed from a series of numerical tests of physical processes which may affect the timing and spatial location of the phyto-

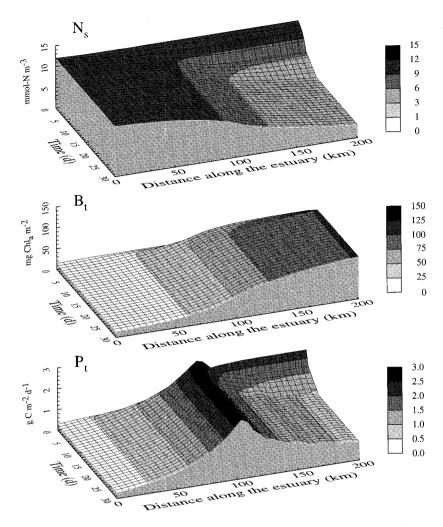


Fig. 8. Temporal evolution along the simulated estuary of surface nutrients ($N_{\rm s}={
m NO_3+NH_4}$ averaged over the euphotic zone), depth-integrated phytoplankton biomass (B_t) and primary production (P_t) obtained with a 2-fold weaker baroclinic velocity profile than that observed (maximum surface current of ca 0.3 m s⁻¹)

plankton bloom in the LSLE. The first process is related to vertical turbulent mixing, which affects light and nutrients conditions as well as the vertical distribution of the phytoplankton cells in the light and nutrient gradients. The effect of turbulent mixing on nutrient supply is negligible for the onset of the bloom as nutrients are generally abundant after winter mixing. For a stratified euphotic zone, as in the LSLE, the effect of vertical turbulent mixing on the light conditions can be estimated by comparison of the time scale $t_{\rm om} = 1/(2K_z\chi^2)$, which compares the time to mix through 1 optical depth with the doubling time in carbon biomass t_{c_i} as a function of light limitation (e.g. Prieur & Legendre 1988). When $t_{\rm om}$ is less than $t_{\rm cr}$ vertical turbulent mixing affects the light conditions experienced by phytoplankton in a way that decreases t_{c} , i.e. the cells receive a larger amount of light due to the non-linear change of irradiance with depth. The effect then favors phytoplankton growth and, as we are searching for factors that limit the bloom, does not need to be considered further.

Lucas et al. (1998) have shown that, for shallow estuaries (0 to 20 m), the main effect of vertical mixing on phytoplankton blooms is through turbulent leakage of phytoplankton cells out of the euphotic layer. In this study we tested the effect of vertical turbulent mixing by varying ε (the TKE dissipation rate) across the mean stratification observed in the LSLE in late spring and prior to the bloom. Our simulations have shown that high vertical turbulent mixing (ϵ up to 5.0×10^{-7} m² s⁻³) delays the bloom by only 5 to 10 d and, moreover, increased new production and phytoplankton biomass computed by the model. It appears then that the increase in stratification from Legs G2 to G4 (see Savenkoff et al. 1997), and increased stability of the water column, would not be the sole factor responsible for the timing of the bloom. None of these stratification regimes limits phytoplankton growth as the model shows that there is insufficient turbulent leakage of phytoplankton cells out of the euphotic zone even with the least stratified regime (Leg G2) and with high TKE dissipation rates. Note that in the LSLE the spring to early summer changes in stratification are mainly dependent on salinity changes, hence driven by the spring freshet, so that the stratification regime

tested here may have prevailed for several weeks before the COUPPB-90 experiment.

The sinking hypothesis

Our simulations indicate that phytoplankton cell sinking is a factor that may limit phytoplankton growth in the LSLE. Species that form the bulk of the phytoplankton bloom in July are mainly diatoms of *Thalassiosira*, *Nitzschia*, *Skeletonema* and *Chaetoceros* (Levasseur et al. 1984, Sime-Ngando et al. 1995, Savenkoff et al. 1997). Smayda (1970) reported mean values of sinking rate for actively growing individual cells of less than 2 m d⁻¹ for these genera, while sinking rates greater than 2 m d⁻¹ would be more relevant for senescent cells. In our simulations, the sinking rate of 2 m d⁻¹

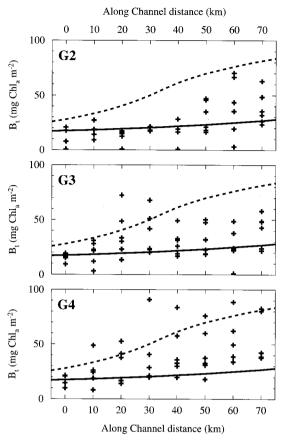


Fig. 9. Depth-integrated (0 to 15 m) phytoplankton biomass in the along-channel direction for the 3 legs, G2 to G4, of the COUPPB-90 experiment; the section starts (km 0) at Stns 43 to 48 in Fig. 1; solid and dashed lines: steady-state depth-integrated phytoplankton biomass extracted from the simulations using the observed and 2-fold weaker baroclinic velocity profiles, respectively, considering that the COUPPB-90 area begins 70 km downstream of the simulated area

is a critical threshold since phytoplankton biomass decreased rapidly for higher sinking rates. Indeed, the euphotic zone is typically 10 to 15 m deep in the LSLE, due to the high turbidity of the freshwater from the St. Lawrence, Saguenay and Manicouagan-Bersimis-Aux Outardes rivers. With a sinking rate of 2 m d⁻¹, the time for 1 phytoplankton cell to sink through the euphotic layer is close to the minimum time scale of the bloom (5 to 6 d). Obviously, sinking rates of 2 m d⁻¹ and more may have a more dramatic effect on the phytoplankton dynamics than vertical turbulent mixing.

This reasoning considers that only individual cells sink. Diatoms can also form chains and aggregates that can sink at speeds of several tens of meters per day (e.g. Smayda 1970, Alldredge & Jackson 1995). However, it is unlikely that the bloom could develop in the thin euphotic layer of the LSLE with such high sinking rates. This agrees with the finding that flocculation and aggregation occur largely after the bloom maximum.

Finally, inclusion of phytoplankton cell sinking in the 1-D model can lead to stationary low phytoplankton biomass under high-nutrient, high-light conditions (e.g. Fig. 5d). While such a steady state may account for the absence of blooms, as observed during Leg G2, it would not explain why the phytoplankton bloomed 10 d later, in a period during which the limiting effect of sinking would become increasingly important, as stratification increases from Legs G2 to G4. Limitation of phytoplankton dynamics by sinking is then unable to explain the 2 mo delay of the LSLE bloom.

The flushing hypothesis

As other hypotheses cannot account for the delay of the spring phytoplankton bloom in the LSLE, we must consider the effect of the seaward component of the buoyancy-induced current linked to the freshwater runoff of the St. Lawrence and Saguenay rivers. Our 2-D simulations show that a residual, seaward surface advection of 0.6 m s⁻¹ prevents biomass accumulation along a distance equivalent to the length of the LSLE. Imposing a similar but 50 % weaker baroclinic velocity profile resulted in a situation where only the second half of the simulated area showed the characteristics of a phytoplankton bloom, a situation which compares well with observations made during the COUPPB-90 experiment.

The 2-D simulations described here were done with the standard parameter set defined in the 1-D simulations, which assumed a low TKE dissipation rate and no phytoplankton cell sinking. This represents the shortest time scale for the simulated phytoplankton bloom (e.g. Figs. 4 & 5). Increasing the vertical turbulent regime or adding phytoplankton cell sinking lengthens the bloom time scale and, consequently, increases the limiting effect of seaward surface advection on phytoplankton biomass accumulation. Comparisons with the COUPPB-90 data set suggest that the sinking rates of the diatom population were low during the 1990 bloom. The effect of such low sinking rate on a 1-D phytoplankton bloom is mainly a slight delay (2 to 3 d by comparison with the standard simulation, see Figs. 4b & 5b) but its effect can be much stronger when advection is considered. With a background seaward velocity of 15 cm s⁻¹, this 2 to 3 d delay is equivalent to 25 to 40 km, or approximately half the length of the COUPPB-90 area. In other words, this delay is of the same order as the increased residence time in the COUPPB-90 area in response to a decrease in the seaward residual buoyancy-induced velocity from 20 to 15 cm s^{-1} (e.g. Fig. 11)

Due to this close match between the bloom development time and residence times in the LSLE, the seaward residual buoyancy-induced circulation is able to

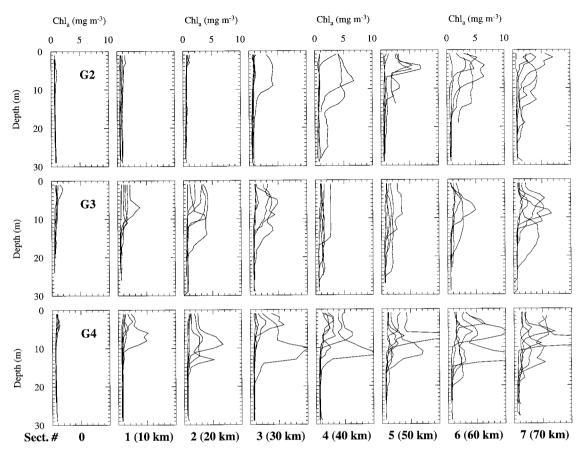


Fig. 10. Variability of the vertical profiles of chl a in the along-channel direction during the 3 legs, G2 to G4, of the COUPPB-90 experiment; sections are numbered 0 to 7 and are spaced 10 km apart

prevent accumulation of phytoplankton biomass, but small changes in its velocity may trigger the onset of the bloom. The onset of the phytoplankton bloom in the LSLE generally coincides with the decrease in the freshwater runoff from the St. Lawrence and Saguenay rivers that occurs in late June/early July (Koutitonsky & Bugden 1991), so that the subsequent decrease in seaward residual advection may determine the timing of the phytoplankton bloom in the LSLE. The bloom generally appears first in the downstream part of the LSLE and then propagates upstream (e.g. Therriault & Levasseur 1985), a train of events which agrees with the flushed euphotic zone hypothesis. CZCS (Coastal Zone Color Scanner) imagery of the gulf and the estuary has shown that discharge of the St. Lawrence and Saguenay rivers and their fluctuations are involved in the spatio-temporal variability of phytoplankton abundance in the LSLE and Northwest Gulf (Fuentes-Yaco et al. 1997). Trites & Drinkwater (1991) reported toxic algae blooms in the Western Gulf that would originate from the LSLE, suggesting that the buoyancy-induced seaward advection may limit phytoplankton accumulation in the estuary. Moreover, the role of transverse baroclinic currents at the onset of the phytoplankton bloom in the LSLE, as shown in Vézina et al. (1995) and Savenkoff et al. (1997), strengthens this flushing hypothesis.

The role of other environmental factors

The limiting effect of seaward advection depends on the ratio between the residence time of phytoplankton cells in the LSLE on the bloom time scale, i.e. the time scale of phytoplankton biomass development, and nutrient depletion which, in turn, depends on environmental conditions, such as light, temperature and vertical turbulent mixing, and on ecological conditions, such as grazing, regeneration and sinking. Any biological or physical process that lengthens the bloom development time will strengthen the flushing effect.

Here, temperature limitation of phytoplankton growth rate (Eppley 1972) is an important factor to consider when dealing with phytoplankton dynamics in subarctic areas such as the LSLE (e.g. Levasseur et al. 1984). The LSLE is known as a cold environment (e.g. Gratton

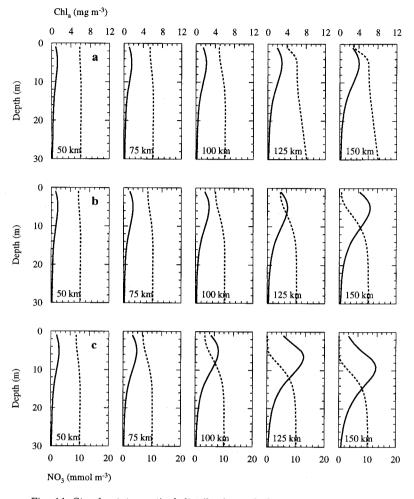


Fig. 11. Steady-state vertical distributions of phytoplankton biomass (as chl a, solid lines) and nitrate (NO₃, dashed lines) at selected distances corresponding to the COUPPB-90 area for a slowly sinking population ($w_s = 1.5 \text{ m d}^{-1}$) and baroclinic velocity profiles of (a) 40% (surface velocity of ca 0.20 m s⁻¹), (b) 30% (ca 0.25 m s⁻¹), and (c) 25% (ca 0.15 m s⁻¹) of the measured velocities

et al. 1988) but winter or early spring diatom blooms have been reported in a low temperature environment (<2°C) in North Atlantic coastal waters by Hitchcock & Smayda (1977) and Townsend et al. (1994). Here, we tested the effect temperature limitation of phytoplankton growth rate has on the bloom development time with a simple 1-D simulation using a 2-fold higher minimum doubling time $t_{\rm dmin}$, or equivalently a 2-fold lower maximum growth rate, $\mu_{\rm max}$. It is not a true use of the Eppley (1972) relationship, rather a simple illustration of the possible effect of assuming growth rates reduced by half due to very low temperatures.

Fig. 12a,b compares the result of this simulation to the standard simulation. A lower $\mu_{\rm max}$, as the result of a temperature limitation of phytoplankton growth rate, leads to a 5 to 10 d delay of the bloom. Nevertheless, the increase in surface temperature observed between

Legs G2 and G4 of the COUPPB-90 experiment was no more than 3°C, and monthly mean temperatures computed by Petrie et al. (1996) for the LSLE show that the warming of the upper 15 m of the water column takes place from April to June, non-limiting surface temperature (>5°C) being achieved during May. Hence, temperature limitation of phytoplankton growth rate is unable to explain the 2 mo delay of the bloom. Nevertheless, temperature limitation of phytoplankton growth rate may combine with the flushing during the earlier growing season (March-April) to prevent early blooms in the LSLE.

Likewise, we tested the effect of grazing on the bloom development time considering a 2-fold higher maximum grazing term, G_{max} . It can be seen from Fig. 12c that doubling the grazing pressure limits the development of the phytoplankton biomass but does not delay or prevent the bloom. Obviously, increasing grazing pressure further will remove enough phytoplankton biomass to lead to no-bloom conditions. To have an effect on the timing of the bloom, one may consider peculiar interactions between grazers and phytoplankton, i.e. an initially grazing-limited phytoplankton biomass which blooms due to the removal of its grazers. Given the estuarine circulation in the LSLE, this would imply differential flushing of phytoplankton and their grazers. However, ciliates and adult females and young stages of the dominant herbivorous copepods are surface-dwelling organisms, hence subject to the same flushing as phytoplankton.

Comparisons with other estuarine systems

Estuaries in which a late bloom is related to the spring-freshet-induced flushing of the surface layer estuaries are generally fjords which are deep and narrow (Sinclair et al. 1981). In a shallow well-mixed estuary, wind and tidal mixing have greater effects on phytoplankton dynamics, through either light conditions (e.g. Monbet 1992), leakage of phytoplankton cells out of the euphotic zone (Lucas et al. 1998) or seeding of the water column by resuspension of resting spores (e.g. Haas et al. 1981), than the flushing of the surface layer. However, some examples of flushed phytoplankton bloom are reported in shallow estuaries, but for short time scales and generally for nanoplankton populations of the oligo- to mesohaline zones, as in the Neuse River Estuary in North Carolina

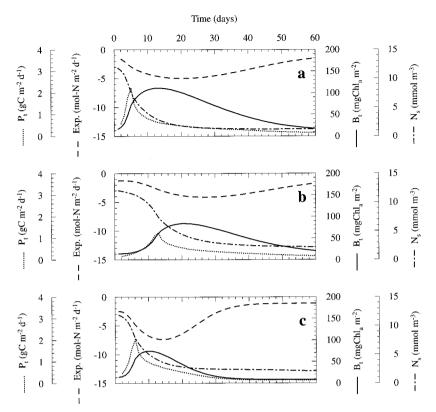


Fig. 12. Temporal evolution of depth-integrated primary production (P_{t} , 0–15 m), chl a biomass (0–15 m) and exported production (Exp., 0–150 m) and surface nutrients nutrients ($N_{s} = NO_{3}+NH_{4}$ averaged over the euphotic zone) for (a) the 60 d standard simulation compared with (b) a temperature limitation of phytoplankton growth rate scenario and (c) a higher grazing pressure scenario (see text for the 2 scenarios)

(Boyer et al. 1993), the James River Estuary in Virginia (Moon 1990) or in the Lower Hudson Estuary (Malone et al. 1980). In addition, the differential timing of the spring bloom between the lower, medium and upper Chesapeake Bay reported by Glibert et al. (1995) may be interpreted as a limited flushing effect. Indeed, it is likely that short-term fluctuations of other environmental conditions, i.e. turbulent mixing, turbidity, etc. (e.g. Malone et al. 1980, Legendre et al. 1982, Harding 1994, Glibert et al. 1995), would mask the effect of flushing on phytoplankton dynamics in shallow stratified estuaries.

Sinclair et al. (1981) argued that a necessary condition for a marked flushing effect is the narrowness of the estuary, which prevents the establishment of transverse motions. Our simulations suggest that flushing can occur in wide estuaries as well. The characteristic buoyancy-induced coastal jet considered in our simulations would not act on the whole LSLE width, as it is restricted to a width of 1 or 2 internal Rossby radii (ca 10 km in the LSLE), which is only one-quarter to one-third of the Estuary. Moreover, strong variations of the LSLE general circulation patterns occur during the

bloom period (late spring/early summer), including cyclonic or anticyclonic eddies which have been attributed to atmospheric events (e.g. Tee & Lim 1987, Mertz et al. 1992) and freshwater wave-like or pulse-induced instabilities (Mertz et al. 1988, Koutitonsky et al. 1990), the latter now considered to be dominant (Ingram & El-Sabh 1990).

Nevertheless, such mesoscale features would not prevent the flushing of the surface layer as long as their residence time in the area is less than the bloom development time. Savenkoff et al. (1997) estimated the seaward speed of the freshwater pulse that induced the onset of the 1990 bloom to be ca 0.15 m s⁻¹. Little is known about these mesoscale features in the pre-bloom period, although it is reasonable to assume they have short residence times in the LSLE, since they would be advected with speeds equivalent to the mean seaward buoyancy-induced residual current ($\geq 0.3 \text{ m s}^{-1}$). Hence, by influencing both the flushing time and environmental conditions (temperature and turbulent mixing), the seasonal evolution of the buoyancy circulation induced by freshwater discharge from the St. Lawrence and Saguenay rivers is likely to be the main factor regulating phytoplankton blooms in the LSLE.

General implications

The seasonal evolution of light conditions in the upper layer has been recognized as the main factor regulating the primary production seasonal cycle for the whole estuary and gulf system (Levasseur et al. 1984, Therriault & Levasseur 1985). The reasoning of Levasseur et al. (1984) was based on the Sverdrup-Riley critical depth model but it also appears from these studies that by April the light availability is not limiting phytoplankton growth in the LSLE. This indicates that the Sverdrup-Riley critical depth model sometime fails in estuarine systems, as shown by Lucas et al. (1998) in shallow estuaries. The Sverdrup-Riley model only gives necessary conditions for a phytoplankton bloom but, in its original mathematical form, cannot explain the delayed phytoplankton bloom in the LSLE: it only considers a simple loss term which stands for mean respiration and grazing, whereas sinking and physical loss terms are not included (e.g. Smetacek & Passow 1990, Lucas et al. 1998). Physical loss terms would consider turbulent mixing and horizontal advection losses in the LSLE.

The 2-D simulations also showed that seaward advection may spatially decouple zones of high phytoplankton biomass and new production at a mesoscale. This simple advective forcing can lead to spatial variability in the trophic characteristics of the ecosystem, which has important geochemical implications, for example, for the depth-integrated assimilation number or for the f-ratio. The use of such geochemical variables requires equilibrium of vertical fluxes through a laterally isolated (i.e. no lateral transport) water mass. Vézina (1994) reported strong mesoscale variability in nitrogen uptake and the f-ratio during the 1989 bloom in the LSLE and showed that this assumption would not be valid during the phytoplankton bloom due to a lack of equilibrium between biomass increase and biomass export out of the euphotic zone. Earlier, Sinclair (1978) reported a lack of coherence in measured phytoplankton physiological state indices and questioned their utility.

As suggested by our simulations, such a lack of equilibrium may be partly explained by the advective effects on the phytoplankton-nitrogen system of new production that results in export of phytoplankton out of the localized high primary production zone. Vertical motions associated with geostrophic dynamics can induce similar spatial uncoupling at a small scale (Zakardjian & Prieur 1998). Hence, the complex circulation patterns observed during the late spring bloom period can lead to strong small- and mesoscale heterogeneities of ecosystem characteristics that are difficult to interpret without taking into account the advective forcing of the ecosystem.

CONCLUDING REMARKS

Physical-biological modeling allows the evaluation of the influence of hydrodynamic conditions on biological time scales as well as their matching with physical time scales. This study highlights the major role of flushing of the euphotic zone by freshwater runoff in the control of the timing of the phytoplankton bloom in the LSLE. This is due to a close match between the time scales of bloom development and flushing. Other environmental conditions, such as the vertical turbulent mixing regime (through stratification or TKE dissipation rate), water turbidity and sinking of phytoplankton, affect the bloom development time and act in concert with advection to determine the timing, duration and intensity of blooms. An implication of this result is that interannual variation of the spring freshwater runoff may drive interannual variation in the dynamics of phytoplankton blooms in the LSLE. The

simple 2-D simulations described here cannot, however, take into account the complex mesoscale features that occur in late spring in the LSLE. Given the possible biogeochemical implications of the strong influences of physical processes on primary production in the LSLE, particularly the spatial decoupling of phytoplankton biomass and primary production, a complete, coupled biological-physical 3-D study is warranted.

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