

1 **Do habitat measurements in the vicinity of Atlantic salmon (*Salmo salar*) parr matter?**

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19 **Abstract**

20 Atlantic salmon (*Salmo salar*) parr habitat characterization is usually performed by *in situ*
21 measures of key environmental variables, taken at the exact fish location, or conversely, in large
22 sampling sections, often ignoring variability in the immediate vicinity around individuals. These
23 data may have a critical importance in development and validation of habitat preference models.
24 The influences of seven increasing distances of measurements, the number of considered
25 measures, and two depth of velocity measurement were tested in the calculations of HSI (Habitat
26 Suitability Index) from a multiple-experts fuzzy model. The radius of 50 cm around the fish, an
27 average measure of 6 measurements in the neighbouring environment and a velocity measured at
28 60% of the depth gave the highest HSI values. These results show some potential for the use of
29 an intermediate study scale, between micro- and mesohabitat, and questions how the fish habitat
30 conditions are currently measured.

31 **Keywords:** Fuzzy logic, Multiple-expert fuzzy modelling, Habitat model, Habitat measurement
32 methods, Intermediate scale measurements, Habitat Suitability Index.

33

35 In salmon habitat assessment, environmental measures characterizing this habitat are most often
36 unique focal measurements taken at the exact location of the fish (e.g. Morantz et al., 1987; Guay
37 et al., 2000), or a composite data from several measurements realized in large sampling section
38 (e.g. Hedger et al., 2005). Atlantic salmon parr are territorial: they present agonistic behavior to
39 defend a territory in order to maintain the best possible place to feed, shelter and grow (Gerking,
40 1953; Heland & Dumas, 1994; Höjesjö, Kaspersson, & Armstrong, 2015). The size of this
41 territory varies according to factors such as fish size, age, habitat heterogeneity and food
42 availability (Grant, Steingrimsson, Keeley, & Cunjak, 1998; Keeley & Grant, 1995; Lindeman,
43 Grant, & Desjardins, 2015), centred on a “home rock” around which the parr moves (Guay et al.,
44 2000). This behavior could interfere with the results of habitat models and characterization
45 analyses, since the data (i.e. the measurements of environmental variables) collected at the focal
46 location or in the whole section may not reflect the actual habitat being used by fish. In addition,
47 depending on the methodology and the fishing gears used to sample fish, *in situ* measurements of
48 habitat variables associated with fish presence can be done at different scales (Heggenes, 1990;
49 Wildman & Neumann, 2003). For example, sampling techniques such as direct viewing by
50 snorkeling (Flebbe & Dolloff, 1995) or detection by electronic tags can provide information on
51 the exact location of the fish and allow the researcher to associate habitat variable measurements
52 at the precise location of the fish. Alternatively, seining or electrofishing (Foldvik, Einum, &
53 Finstad, 2016; Mäki-Petäys, Erkinaro, Niemelä, Huusko, & Muotka, 2004) do not allow this
54 precision and the subsequent measurements of environmental conditions are related to a larger
55 area of capture. Finally, the protocols used to describe salmon habitat are different: the water
56 velocity can be measured at or near the bottom (e.g. Heggenes et al., 1995) or at a variable

57 depths according to the total depth of the water column (e.g. Morantz et al., 1987). In addition,
58 the habitat can be characterized by a single measurement per fish (e.g. Morantz et al., 1987) or
59 up to 15 measurements in the entire reach (e.g. Hedger et al., 2005).

60 Yet, these uncertainties may have a critical importance in habitat modeling. Indeed, before being
61 made available for managers, models have to be validated, i.e. they have to be tested with data
62 not used for model development and calibration. During the validation process, they have to
63 reach the required performance standards (Rykiel, 1996). In the specific case of habitat models,
64 one validation process consists in confronting the modeled predictions of habitat quantity and
65 suitability indices against field observations of presence or absence of fish (Fukuda & Hiramatsu,
66 2008; Mocq, St-Hilaire, & Cunjak, 2013; Mouton et al., 2008). Uncertainties in environmental
67 data and approximation in habitat characterization may induce errors, distort the validation, and
68 even affect the robustness of the model when the measures are used as input data.

69 We questioned the influence of these uncertainties in measured data in habitat models, and we
70 investigated the impact of the scale at which measurements are taken by hypothesizing that this
71 scale can affect the model results or the validation efficiency. We tested the hypothesis that focal
72 measurements (i.e. at the exact location of the fish) may not be the best representation of habitat
73 variables that determine (in part) fish presence, by using as a tool a previously developed fuzzy
74 model for Atlantic salmon (*Salmo salar*) parr habitat (Mocq et al., 2013). Fuzzy logic is regularly
75 used to model efficiently fish habitat (Muñoz-Mas et al., 2016) and this approach allows
76 determining a scale that provides a better habitat description according to the expert system. In
77 addition, the influences of the velocity in such a fuzzy system were assessed when it has been
78 measured at the bottom or at 60% of total depth, which is used as an estimate of average velocity
79 over the entire water column for shallow rivers. Finally, focusing on a study scale between
80 micro- and mesohabitat, we varied the number of measurements in a close neighborhood around

81 the individuals used to evaluate salmon habitat quality to highlight their impact in the model and
82 determine which numbers provide the best results, with the objective of improving the usual
83 method of sampling.

84 2 MATERIAL AND METHOD

85 2.1 Fuzzy model

86 To build the fuzzy logic Atlantic salmon parr rearing habitat model (Mocq et al., 2013), three of
87 the most important variables defining salmon distribution and abundance, i.e. depth, velocity and
88 mean substrate diameter (Armstrong, Kemp, Kennedy, Ladle, & Milner, 2003; Bardonnet &
89 Baglinière, 2000; Heggenes, 1990) were chosen as input variables, and Habitat Suitability Index
90 (HSI) as the output variable. Each variable domain was split into three categories defined by
91 combinations of linear membership functions, which constitute the fuzzy sets. Then HSI
92 consequences of each possible combination of every category of the three variables were
93 determined with “If...Then...” fuzzy rules, i.e. 27 possible rules. A comprehensive presentation
94 of the fuzzy method and model building is provided by Mocq et al. (2013).

95 Since the experts’ geographic range of knowledge showed influences on the fuzzy model results
96 (Mocq, St-Hilaire, & Cunjak, 2015), only those from eastern Canada were selected. Twenty
97 experts defined individual fuzzy sets and fuzzy rules, which were integrated and treated with the
98 fuzzy package FuzzyToolkitUoN in R (R Development Core Team, 2016). The Mamdani
99 inference method was used (Mamdani, 1977; Shepard, 2005) to process data from fuzzy input
100 sets to the fuzzy output set. The defuzzification (i.e. the transformation from the final fuzzy sets
101 to a crisp number) was done by the commonly used method of centre of gravity (Jorde,

102 Schneider, Peter, & Zoellner, 2001), providing an HSI value between 0 (representing an
103 unsuitable habitat) and 1 (representing the most suitable habitat) for each expert.

104 2.2 Sampling campaign

105 2.2.1 Study sites

106 Environmental measurements were taken in three Canadian Atlantic salmon rivers (Fig.1): the
107 Sainte-Marguerite River (Québec), the Little Southwest Miramichi River and its tributary,
108 Catamaran Brook (New-Brunswick).

109 Catamaran Brook flows for 20.5 km for a drainage basin of 50 km² (Cunjak, Caissie, & El-Jabi,
110 1990) with a mean annual discharge of 0.6 m³ s⁻¹ (Benyahya, Daigle, Caissie, Beveridge, & St-
111 Hilaire, 2009). The Little Southwest Miramichi (Cunjak et al., 1990; Johnston, 1997) drains a
112 1340 km² basin with a mean annual discharge of 32.2 m³ s⁻¹ (Benyahya et al., 2009). These
113 streams are characterized as relatively pristine (Cunjak et al., 1993). Finally, the Sainte-
114 Marguerite River is a 100 km-long river (Guay et al., 2000), with a catchment of 2100 km² and a
115 mean annual discharge of 30.93 m³ s⁻¹ (Benyahya et al., 2009). Wild Atlantic salmon populations
116 are present in all three rivers.

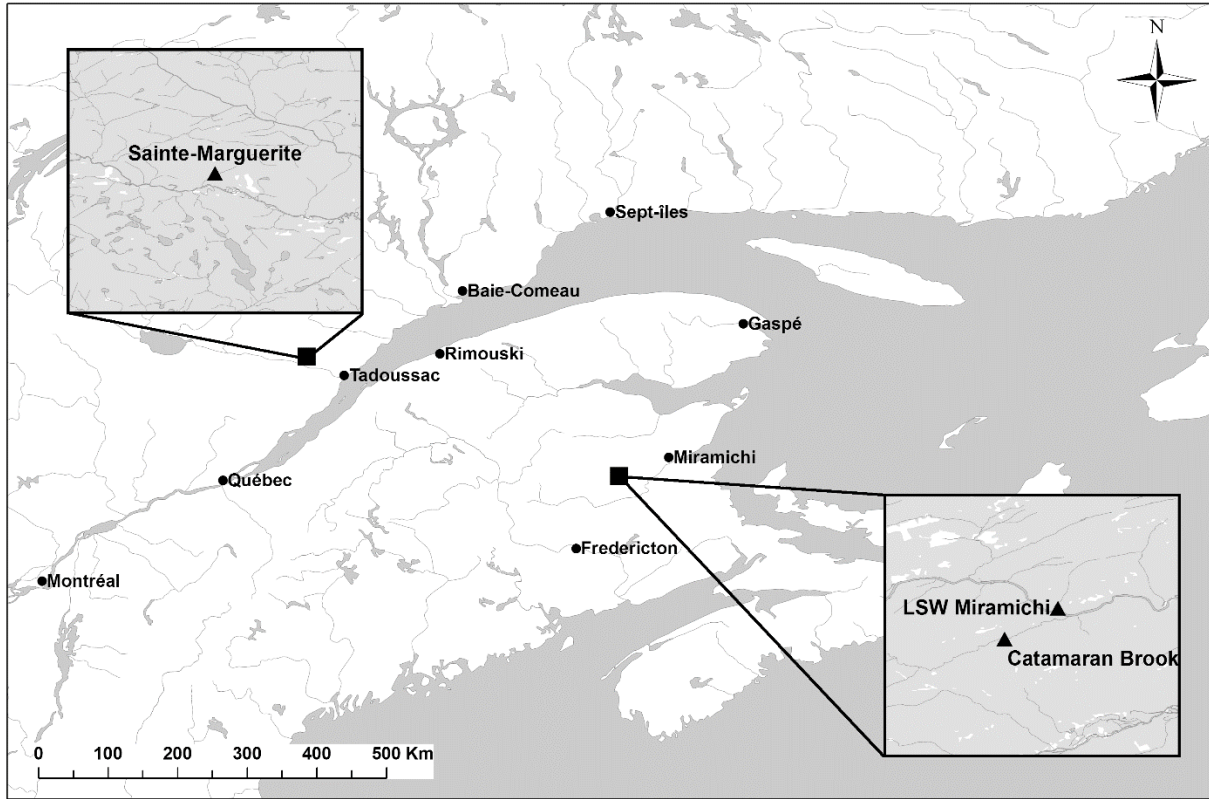


Figure 1: Localization map of the Sainte-Marguerite River (Québec), Little Southwest Miramichi River and its tributary, Catamaran Brook (New-Brunswick).

117

118 2.2.1 Sampling method

119 The sampling campaign took place in July 2012, at four sites in Catamaran Brook and two sites
 120 in Little Southwest Miramichi. Two sites were sampled in the Sainte-Marguerite River, in June
 121 and September of 2012. The considered reaches were sections of length 5 times larger than the
 122 width when the width was lower than 8 m. If the width was larger than 8m, a 6x25 m subsection
 123 along the banks was sampled.

124 The protocol was divided into three steps. First, salmon parr (1 and 2+ year-old) location was
125 assessed by a snorkeling diver moving upstream in zigzag patterns, reaching the fish location
126 from behind to avoid flight. Upon recognition, the diver waited motionless for a minimum of one
127 minute to ensure that the fish position was not influenced by the observer. A painted stone was
128 dropped onto the bottom as a position marker.

129 Environmental measurements were done for the whole reach, including in the vicinity of the
130 spotted fish. The whole reach were assessed by measurements along transects, every 2 m, with
131 nine measurements per transect, each measure constituting a coordinated node of a grid, dividing
132 the section into cells. Depth was measured with a ruler, and velocity with an electronic
133 flowmeter Flo-Mate model 2000 (Marsh-McBirney, inc.) during at least 2 min, at the bottom (i.e.
134 the position of the parr on its home-rock) and at 60% of the depth in the water column (i.e.
135 classical depth of velocity measurements). Substrate composition was assessed by evaluating the
136 proportion of the different classes of grain size according to the modified Wentworth scale
137 (Schoeneberger, Wysocki, Benham, Soil Survey Staff, & Natural Resources Conservation
138 Service, 2012) and a mean substrate size was calculated by weighting the diameter of each class
139 by the evaluated proportion observed at the site.

140 Microhabitat measures were made at the precise position and in the vicinity of the located fish.
141 Velocity and depth were measured, first at the exact location of the fish, then at distance of 10,
142 25 and 50 cm from it, representing respectively a circular area of 0.04, 0.2 and 0.79 m², for a
143 total of 5 points around a circle for each radius (3 upstream, 2 downstream). The substrate was
144 assessed by evaluating the proportion of the different classes of grain size, at the exact position
145 of the fish first, then in a square of 10, 25 and 50 cm on each side always centered on the fish.

146 2.3 Data process and statistical analysis

147 First, for each measurement point, a mean Habitat Suitability Index (HSI) value was calculated
148 through the fuzzy inference system of each expert, providing a spatial distribution of HSI. One
149 HSI value was calculated for velocity measured at 60% of the total depth (V_{60}) and for velocity
150 measured at the bottom (V_{bot}). These two sets of mean HSI values were compared with a
151 Wilcoxon matched-pairs signed-ranks test. The variation between the two HSI values at a same
152 point were calculated for each expert and then averaged to visualize the consequence on the
153 differences of velocity measures on a map. The same process was repeated with environmental
154 measures at the exact location of fish. Since the sampled sites presented low parr density, it was
155 considered that the choice of location by the parr was made regarding habitat quality only (i.e.
156 likely no density-dependence effect). The values of HSI were compared with a Wilcoxon
157 matched-pairs test with correction for multiplicity. Our hypothesis was that a fish will choose a
158 location not only because of the conditions at a focal location, but because of the conditions
159 experienced in a short-range neighborhood around the focal location. Consequently, considering
160 the neighboring environmental conditions of the fish position should improve the model results
161 by better describing habitat at a fish location, highlighted by an increase of the mean HSI and/or
162 a decrease of the variability, until a limit where the HSI mean should decrease and/or the
163 variability should increase again.

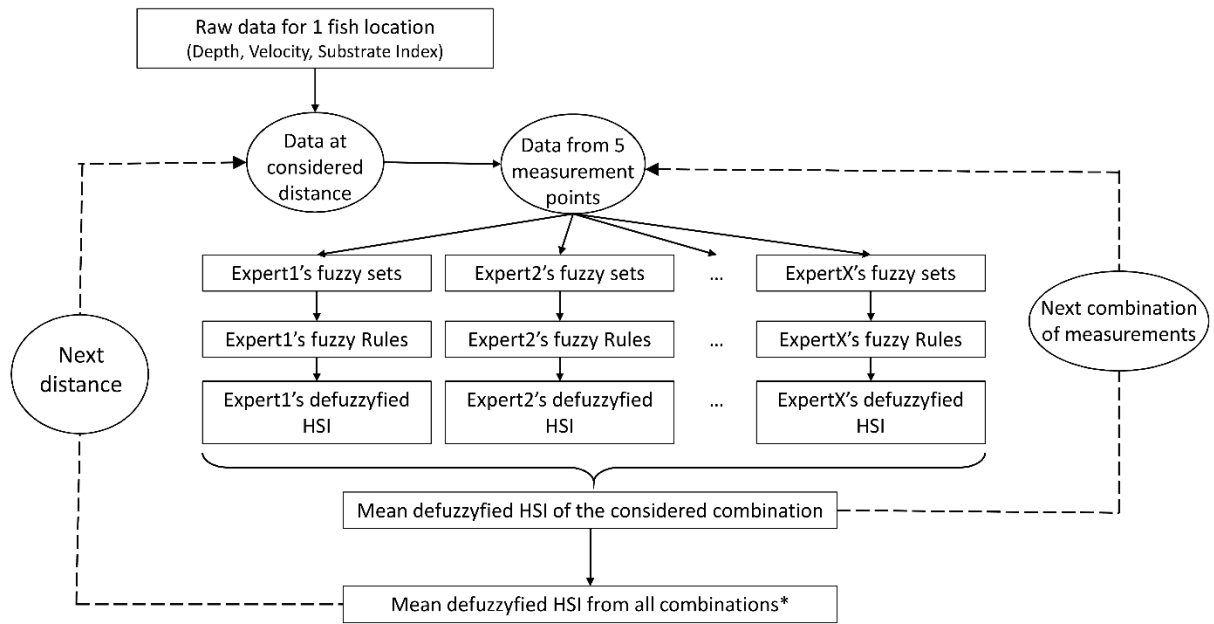


Figure 2: conceptual scheme of our study of impacts of increasing measurements distances on multiple-expert fuzzy models. (* except d310, half of combinations were performed)

164

165 The habitat quality was also assessed within an increasing area around the fish. The unique
 166 measure realized at the exact location of the fish provided the focal data, using a mean substrate
 167 size according to the proportion of different size classes, the velocity measured at 60% of the
 168 depth and the depth measured using a ruler with a 1cm precision. The integration of measures
 169 realized at three distances (i.e. 10, 25 and 50 cm) around the fish provided the data for the areas
 170 of 0.04, 0.2 and 0.79 m². Finally, 4, 11 and 30 m², represented by circles around the fish with
 171 respective approximate radii of 113, 190 and 310 cm, were assessed by integrating the measured

172 data from the grid when they were located in the appropriate area. Then, for each distance and
173 for each fish, mean HSI was calculated from the fuzzy sets and rules developed by each expert,
174 for all of the combination of 5 measurements encompassed in the considered area (Fig. 2). The
175 mean depth, velocity and substrate size of all considered measurements were used as input data
176 in the fuzzy system. For the distance of 310 cm, half of the possible combinations of a set of 5
177 out of 16 measurements, randomly drawn, were used because of computational limitations, but
178 representing a final 50 344 combinations, and 1,066,880 HSI values. The accuracy of each model
179 was estimated with the kappa statistics (see for example McHugh, 2012), with a threshold of HSI
180 at 0.5 beyond which the habitat is considered suitable. Since presenting the HSI values did not
181 allow to visualize the variation of the HSI value for a same fish, the variation of value from the
182 HSI value at the focal point, and the mean HSI value for the considered distance were calculated
183 for each fish. The HSI values according to the distance, and then the variation with the focal HSI
184 values were compared with a Friedman two-way analysis of variance, the non-parametric test
185 corresponding to a repeat-measure ANOVA, with a *post hoc* analysis i.e. a Wilcoxon matched-
186 pairs test with correction for multiplicity.

187 Finally, to evaluate the optimal number of environmental measurements, all measurements
188 included in a radius of 25 cm were considered for each parr (i.e. 11 measures for most of them).
189 All possible combinations of measures, including 1 to 11 measures, were averaged and a HSI
190 value was calculated for each expert fuzzy system and each fish. The accuracy of each model
191 was also estimated with the kappa statistics with a threshold of HSI at 0.5. The variation with
192 the HSI at the focal position of the fish were calculated for each expert, and then averaged for
193 each fish. The HSI variations were compared with a Friedman two-way analysis of variance with
194 a *post hoc* analysis (Wilcoxon matched-pairs test with correction for multiplicity). The optimal

195 number of measurements to characterize the salmon habitat was seen as the category providing
 196 the highest HSI associated with fish presence and/or the smallest variability.

197 **3 RESULTS**

198 On the 10 considered sites, a total of 1366 points were sampled, showing an overall HSI mean of
 199 0.42 +/- 0.12 (range from 0.17 to 0.7) for the velocity measured at 60% of the depth, and an
 200 overall HSI mean of 0.48 +/-0.15 (range from 0.17 to 0.71) when velocity is measured at the
 201 bottom (Tab. 1, Fig.3).

Table 1: Mean, minimum and maximum of Habitat Suitability Index (HSI) values from multiple-experts fuzzy system, with velocity measured at 60% of the depth (V_{60}) and at the bottom (V_{bottom}), and the calculated variation at a same point, considering the entire reach of the station.

River	Site	Year	HSI V_{60}			HSI V_{bottom}			Variation		
			Mean	Min	Max	Mean	Min	Max	Mean	Min	Max
Catamaran	Div2	2012	0.35	0.17	0.59	0.37	0.17	0.66	-0.03	-0.27	0.05
Catamaran	Lor	2012	0.35	0.17	0.64	0.36	0.17	0.63	-0.01	-0.12	0.15
Catamaran	Moc	2012	0.38	0.18	0.66	0.42	0.18	0.67	-0.04	-0.25	0.16
Catamaran	Tom	2012	0.33	0.17	0.59	0.34	0.17	0.57	-0.01	-0.19	0.21
St-Marguerite	Smp	2012-06	0.51	0.29	0.69	0.59	0.31	0.71	-0.08	-0.29	0.15
St-Marguerite	Smp	2012-09	0.51	0.29	0.67	0.58	0.27	0.71	-0.07	-0.30	0.23
St-Marguerite	Smt	2012-06	0.39	0.19	0.70	0.48	0.19	0.68	-0.09	-0.25	0.13
St-Marguerite	Smt	2012-09	0.38	0.21	0.66	0.43	0.21	0.68	-0.05	-0.28	0.22
Miramichi	Alx	2012	0.53	0.30	0.69	0.65	0.30	0.71	-0.12	-0.31	0.03
Miramichi	Spm	2012	0.49	0.28	0.67	0.58	0.31	0.70	-0.09	-0.30	0.21

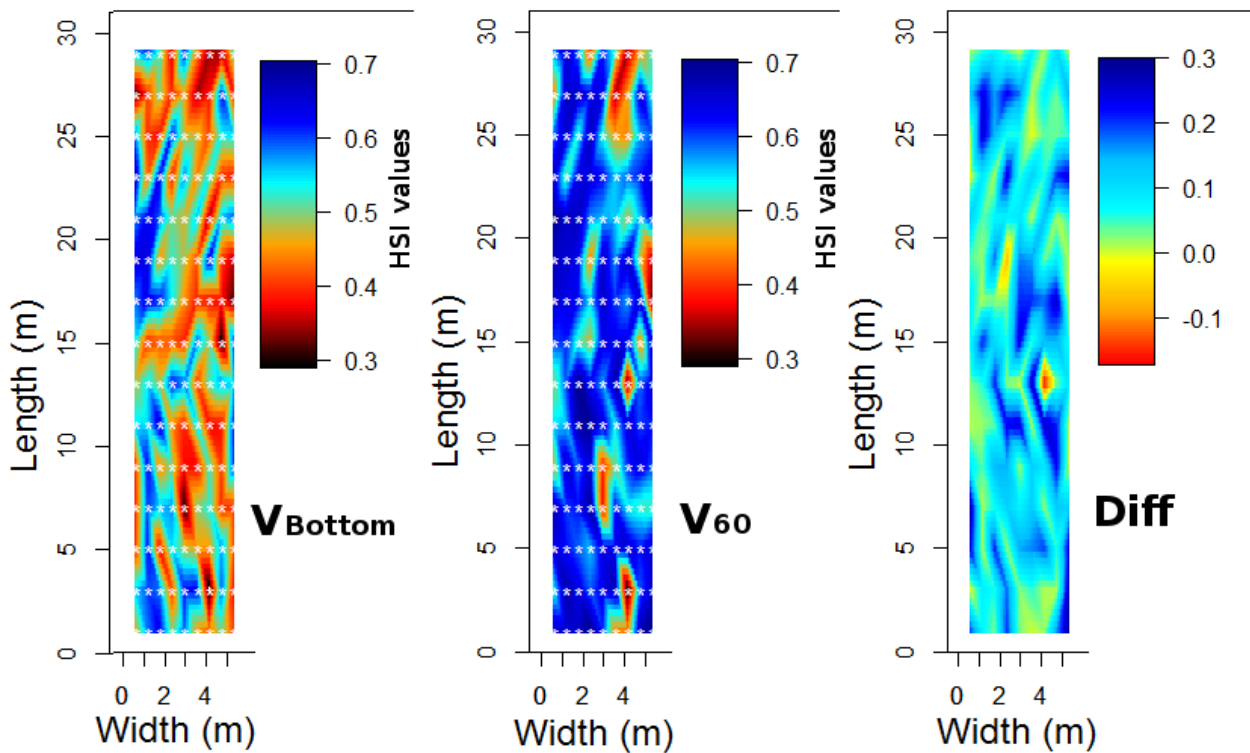


Figure 3: Map of HSI value distribution, after spline interpolation from measuring points (white stars), for velocity measured at the bottom (left panel) or at 60% of the depth in the water column (middle panel), and the difference between them (right panel), in the Miramichi River (Spm site).

203 The two sets of HSI were significantly different (Wilcoxon matched-pairs signed-ranks test with
 204 correction for multiplicity, p -value $< 2.2e-16$). The variations of HSI values were ranged from -
 205 0.30 to +0.23 (mean= -0.06 +/-0.09).

206 In the three rivers, 93 fish were observed. HSI were calculated at the focal location of the fish,
 207 using the velocity measured at the bottom, then at 60% of depth (Fig.4). The resulting HSI
 208 values were significantly higher (Wilcoxon matched-pairs signed-ranks test, p -value < 0.001),

209 when the velocity was measured at 60% of the depth (mean= 0.47 +/- 0.11) than at the bottom
210 (mean= 0.42 +/- 0.1).

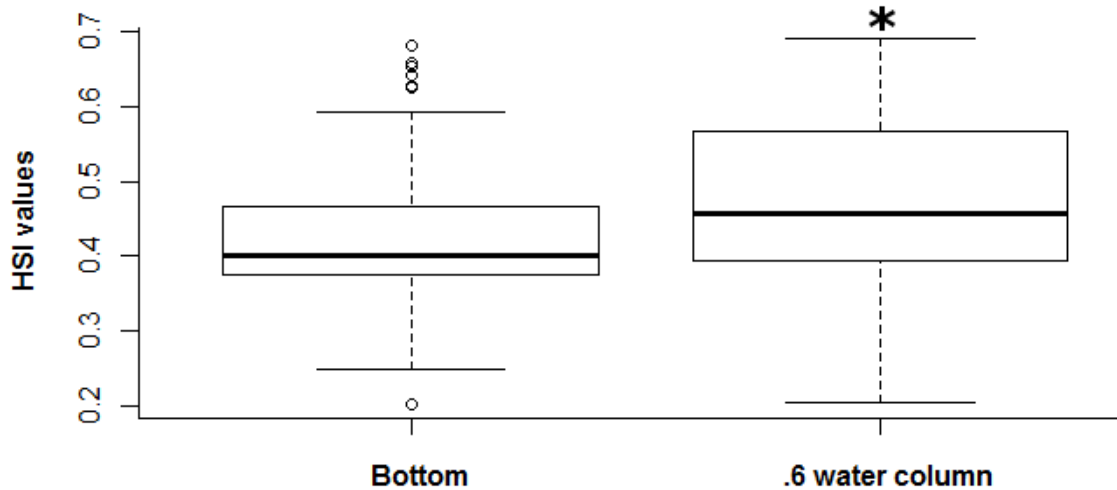


Figure 4: Distribution of HSI values with velocity measured at the bottom (left panel) or at 60% of the depth (right panel) at the exact fish location. Asterisks (*) indicate significant difference (Wilcoxon matched-pairs signed-ranks test, p-value <0.001).

211
212 Regarding the influence of increasing area of measures around the fish, 7 distances were tested: 0
213 (focal measure), 10, 25, 50, 113, 190 and 310 cm around the parr (Tab.2, Fig.5). The focal
214 measures provided the lowest mean HSI and the largest variability (mean= 0.47 +/- 0.11). The
215 HSI increased slowly until d50, and decreased afterward, while the variability decreased
216 progressively. Regarding the variation of HSI by fish, the highest positive difference with the
217 focal HSI value occurred at d50 (mean= $+3.92 \cdot 10^{-2}$ +/- $6.35 \cdot 10^{-2}$) and the highest kappa value is
218 reached at the same distance ($\kappa = 0.66$ at d50; Tab.2). Friedman's test was significant for the HSI
219 values and for the variations with the focal value (both *p-value* <0.01). The post hoc analysis
220 found only a significant difference between d0 and all other distance (all *p-value* <0.01) but no

221 other significant difference was found afterwards. The same analysis on the variation with the
 222 focal values highlighted significant differences between every distances (all *p-value* <0.05) but
 223 d113 and d190 (*p-value*= 0.28).

Table 2: mean values and standard deviation (SD) for HSI values, and for the variation from the focal HSI values for each parr, provided by a multiple-expert fuzzy system on every possible combination of 5 measures encompassed in seven increasing distances (0, 10, 25, 50, 113, 190 and 310 cm) around the considered parr.

Distance	Mean HSI	SD HSI	Mean variation with focal values	SD variation with focal values
d0	0.475	0.112	0	0
d10	0.499	0.106	0.024	0.058
d25	0.506	0.105	0.031	0.062
d50	0.514	0.099	0.039	0.064
d113	0.514	0.095	0.039	0.07
d190	0.514	0.096	0.039	0.08
d310	0.512	0.096	0.037	0.082

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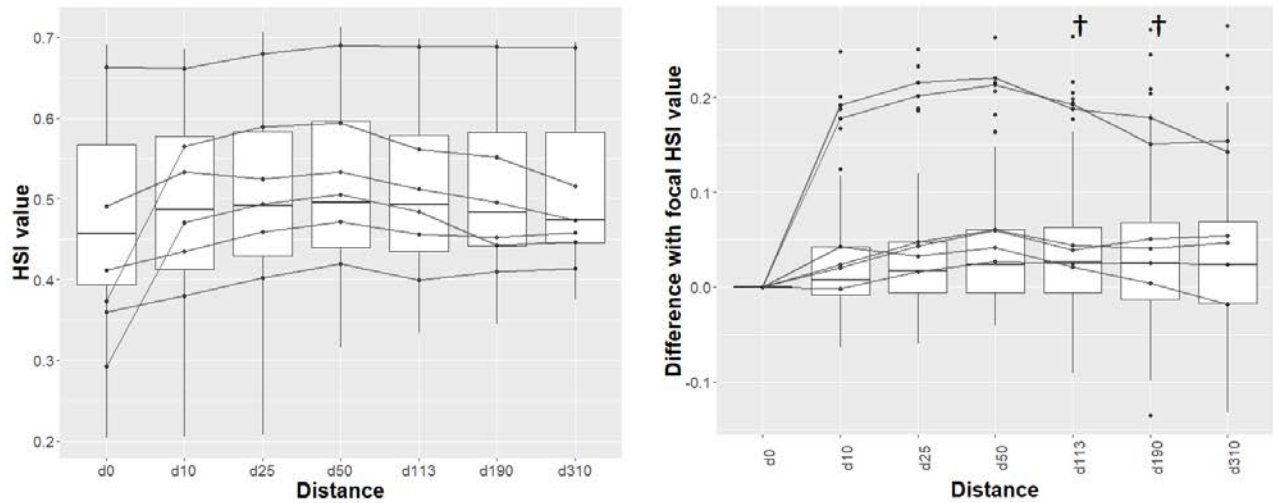


Figure 5: Distribution of HSI values calculated with a multiple-experts fuzzy model (left panel) and the variation of this HSI values with the focal HSI value for each parr (right panel), according to the distance inside which measures were considered as input data for the model; the HSI values (lines) of a selection of 6 (out of 93) different fish are provided. Asterisks (*) indicate a significant difference with the other categories. Daggers (†) indicate non-significant difference between the two categories, all other differences are significant (both post hoc Wilcoxon matched-pairs signed-ranks test).

226

227 For the number of measurements to include in the models, the HSI provided by the different
 228 categories of included measures were averaged for each fish and for each expert. All HSI
 229 calculated showed an increase compared with the value calculated from the focal position 0*.
 230 The maximum is reached with 11 measures but the Cohen's kappa was the highest with 6
 231 measures ($\kappa = 0.66$), providing 94% of the maximum improvement of HSI values (Tab. 3, Fig.
 232 6). The variations of HSI values between categories were significantly different from each other
 233 (Friedman rank sum test, p -value < 0.01). Only the categories of 9, 10 and 11 measures showed

234 significant differences with each other (Wilcoxon matched-pairs signed-ranks test, p -value <0.01
 235 for these three categories, p -value ≥ 0.42 for the other combinations).

Table 3: Means and standard deviations of differences in values of calculated HIS, between focal environmental measures of Salmon parr location (0*) and averaged HSI from every combination of measures, including from 1 to 11 measures.

Number of included measures	0*	1	2	3	4	5	6	7	8	9	10	11
Mean	0	0.023	0.03	0.032	0.037	0.034	0.033	0.033	0.033	0.031	0.032	0.034
SD	0	0.07	0.073	0.072	0.07	0.072	0.073	0.073	0.073	0.069	0.068	0.066

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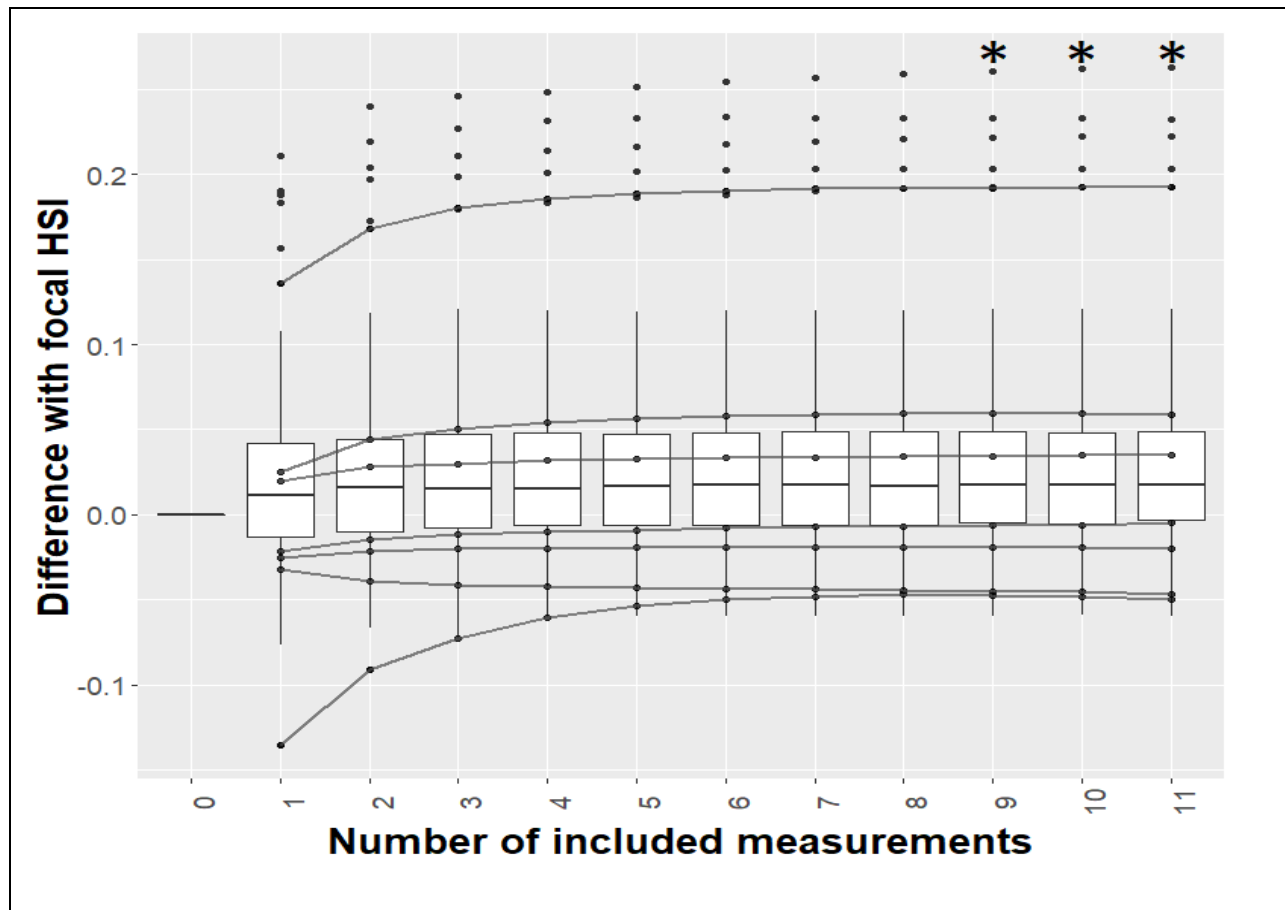


Figure 6: Differences in values of calculated HSI, between focal environmental measures of salmon parr location (0*) and averaged HSI from every combination of measures, including 1 to 11 measures; the differences in HSI values (lines) of a selection of 7 (out of 93) different fish are provided. Asterisks (*) indicate categories with no significant differences from each other (post hoc Wilcoxon matched-pairs signed-ranks test).

237

238 4 DISCUSSION

239 Highlighting and calculating the radial distance from the fish defining a circle in which
 240 measurements could be taken to describe accurately the physical habitat is one possible way to
 241 improve habitat characterization and subsequent habitat modeling. Seven different distances

242 were tested for each presence. The null distance, i.e. the focal measure taken at the exact location
243 of the fish, gave globally the lowest HSI values and the highest variability. This is potentially
244 explained by the fact that only one measurement of each habitat variable may introduce an
245 important bias: assuming the fact that the parr may chose a suboptimal focal location with good
246 vicinity, if the exact location did not present good conditions for the fish, the error is not
247 attenuated by other measurements. Consequently, despite the geographic proximity of their
248 measures, the multiple-measures 10 cm-distance and the unique focal measure showed
249 significant differences in their calculated HSI, the 10 cm-distance providing a slightly better HSI
250 than focal measurements. Then, increasing the distance improved the model performance: for a
251 same fish, the HSI calculated by the fuzzy system reached a peak at a distance of 50 cm, which
252 corresponded with an area of 0.79 m², consistent with previous assessment of parr territory sizes
253 (Keeley & Grant, 1995; Lindeman et al., 2015). Beyond this distance, the calculated HSI values
254 tend to remain constant, the loss of variability in HSI values being the sign of a homogenization
255 of the measures at a large scale instead of a model improvement, as proved by the large
256 variability of HSI values at large distance.

257 Our protocol gave the opportunity to explore the influence of the location of the velocity
258 measurements in an expert fuzzy system. As expected, the data highlighted the fact that the
259 velocities at the bottom were slower than at 60%, because of frictions with the substrate, and
260 sometimes even counter-currents were observed. Near the bottom, parr can save some energy
261 while having access to the drifting food, the habitat quality should be consequently higher there
262 than further up in the water column. The HSI values were slightly lower when the velocities
263 were measured at the bottom than at 60% of total depth. In addition, with the measurements
264 taken at the fish location, our result showed an improvement of the model with higher HSI when
265 the velocity is measured at 60% of the depth. Thus, having better calculated HSI at 60% of the

266 depth instead of at the bottom for a benthic fish is unexpected and raises some questions. It is
267 unclear if the experts unconsciously referred to the velocity in the middle of the water column
268 when they built their memberships functions and rules, or if they consider the 60%-depth
269 velocity is more representative of the velocity associated with the drifting food.

270 It is usual, when characterizing salmon habitat, to take one measure of velocity and depth, and
271 an assessment of the substrate at the exact location of the fish. Our results showed that the values
272 of HSI provided by one measure at the exact location of the fish were generally lower than
273 several measures realized in its direct environment. The average highest HSI value was obtained
274 when all the measurements of each habitat variable, at randomly selected locations (within a
275 radius of 25 cm), were aggregated, but 6 measurements provided the best model. More than 6
276 measurements improved marginally the model outputs and the extra time and efforts required is
277 not warranted in this case. However, a fixed number of measurements to describe the
278 environmental conditions cannot be applicable to every circumstance and should be adapted to
279 local habitat complexity. In addition, our results exhibit a snapshot of habitat suitability, for a
280 short time period and the integration of flow dynamics would be important to model adequately
281 the salmon habitat (Boavida, Harby, Clarke, & Heggenes, 2017). Nevertheless, our study shows
282 a clear difference in the models outputs, highlighting the needs to take into consideration
283 multiple measurements in a close range around the individuals.

284 Our results suggest that the best description of the parr rearing habitat during summer diurnal
285 period to be used in the fuzzy logic model described by Mocq et al. (2013) is reached by taking
286 into consideration the neighboring conditions, i.e. measuring variables at 50 cm from the fish and
287 adding them to calculate a mean value. In addition, this multiple-points sampling protocol could
288 improve the models. Moreover, the fuzzy model is based on expert knowledge and it is possible
289 that they defined the fuzzy sets and rules using characteristics of parr habitat that include all or a

290 part of its territory instead of only its home-rock, as our results indicate. Our hypothesis about
291 the model efficiency was that a high calculated HSI for a presence indicated that habitat
292 description used in the calculations was accurate and consequently, led to a more efficient model.
293 However, parr were frequently present in habitats with calculated HSI under 0.5, while the
294 reaches presented large bands of good quality. Field observations indicate that, in the three
295 sampled rivers, densities extended from 0.5 to 13.5 ind/100 m², were too low to force some parr
296 to use poor-quality habitat. Therefore, density dependence-related biases are not the likely cause
297 of presence of fish in relatively poor habitat. This observation can more likely be explained by
298 expert's unsure or ill-translated knowledge in the fuzzy sets and rules. These biases are found in
299 the fuzzy sets and rules definition: the unsure or ill-defined knowledge could be represented in
300 fuzzy logic by an important overlap between successive categories or by membership functions
301 limits irrelevant with ecological reality. Indeed, accurate codification of the expert knowledge is
302 an important obstacle. The difficulties could come from the expert (unwanted forgetting of
303 information, difficulty of expression or abstraction, fear of personal knowledge disclosure and
304 use; Chevrie and Guély, 1998, Drescher et al., 2013), or from the method (poor ergonomics of
305 the worksheet used to collect data, lack of precision in words or concepts description; Knol et al.,
306 2010). In addition, the growth of the parr evolves quickly in few months, especially for the
307 youngest parr, modifying their habitat preferences and complicating the definition of the related
308 fuzzy sets. Finally, another explanation could be related with the physical factors that were not
309 included in the model: indeed, the habitat selection by parr is under the influences of multiple
310 factors (Armstrong et al., 2003), generally of lesser importance than the three selected variables
311 but sufficiently important to modify substantially the habitat selection of some of the fish.

312 This study is linked to the problematic study of scale, i.e. the spatial and temporal dimension of
313 a process or an entity (Lewis et al., 1996). Considering the Atlantic salmon (*Salmo salar*)

314 physical habitat, some studies work at the microhabitat scale, measuring environmental physical
315 variables in the vicinity of the fish, on the order of the cm² (Armstrong et al., 2003; Heggenes et
316 al., 1995; Heggenes, Bagliniere, & Cunjak, 1999). By contrast, other studies are concerned with
317 the mesohabitat scale, from m² to some tens of m², equivalent to Channel Morphological Units
318 (CMU, e.g. riffle, glide, pool; Folt et al., 1998). For the habitat characterization, the mesohabitat
319 is too large to precisely describe salmon needs and preferences, while the microhabitat neglects
320 environmental elements which could influence the selection and the occupation of the habitat by
321 the fish (Shirvell, 1994). An intermediate scale, considering the exact location of the fish and its
322 close environment as potential used habitat, could be useful to describe more precisely the
323 habitat and improve the model predictions. Moreover, lots of scientific studies and protocols
324 about fish habitat are based, for their field data, on the measures of environmental variables at
325 the exact location of the fish: our results suggest a change in this protocol may be needed, in
326 order to improve the habitat description.

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338 References

339 Armstrong, J. D., Kemp, P. S., Kennedy, G. J. A., Ladle, M., & Milner, N. J. (2003). Habitat
340 requirements of Atlantic salmon and brown trout in rivers and streams. *Fisheries Research*,
341 62(2), 143–170. [http://doi.org/Pii S0165-7836\(02\)00160-1](http://doi.org/Pii%20S0165-7836(02)00160-1)

342 Bardonnet, A., & Baglinière, J.-L. (2000). Freshwater habitat of Atlantic salmon (*Salmo salar*).
343 *Canadian Journal of Fisheries and Aquatic Sciences*, 57(2), 497–506.

344 Benyahya, L., Daigle, A., Caissie, D., Beveridge, D., & St-Hilaire, A. (2009). *Caractérisation du*
345 *régime naturel du débit des bassins versants de l'Est du Canada* (Vol. Rapport R1). INRS-
346 ETE.

347 Boavida, I., Harby, A., Clarke, K. D., & Heggenes, J. (2017). Move or stay: habitat use and
348 movements by Atlantic salmon parr (*Salmo salar*) during induced rapid flow variations.
349 *Hydrobiologia*, 785(1), 261–275. <http://doi.org/10.1007/s10750-016-2931-3>

350 Chevrie, F., & Guély, F. (1998). *Cahier technique n°191: La logique floue*. Groupe Schneider.

351 Cunjak, R., Caissie, D., & El-Jabi, N. (1990). *The Catamaran Brook Habitat Research Project:*
352 *Description and General Design of Study*. *Canadian Technical Report of Fisheries and*
353 *Aquatic Sciences* (Vol. 1751). Moncton: Department of Fisheries and Oceans.

354 Cunjak, R., Caissie, D., el-Jabi, N., Hardie, P., Conlon, J. H., Pollock, T. L., ... Komadina-
355 Douthwright, S. (1993). *The Catamaran Brook (New Brunswick) Habitat Research Project:*
356 *Biological, Physical and Chemical Conditions (1990-1992)*. *Canadian Technical Report*
357 *of Fisheries and Aquatic Sciences*. Moncton: Department of Fisheries and Oceans.

358 Drescher, M., Perera, A. H., Johnson, C. J., Buse, L. J., Drew, C. A., & Burgman, M. A. (2013).
359 Toward rigorous use of expert knowledge in ecological research. *Ecosphere*, 4(7), 1–26.
360 <http://doi.org/10.1890/ES12-00415.1>

361 Flebbe, P. A., & Dolloff, C. A. (1995). Trout Use of Woody Debris and Habitat in Appalachian
362 Wilderness Streams of North Carolina. *North American Journal of Fisheries Management*,

363 15(3), 579–590. [http://doi.org/10.1577/1548-8675\(1995\)015<0579:TUOWDA>2.3.CO;2](http://doi.org/10.1577/1548-8675(1995)015<0579:TUOWDA>2.3.CO;2)

364 Foldvik, A., Einum, S., & Finstad, A. G. (2016). Spatial diffusion modelling of juvenile Atlantic
365 salmon (*Salmo salar*) shows ontogenetic increase in movement rates. *Canadian Journal of*
366 *Fisheries and Aquatic Sciences*, 1–6. <http://doi.org/10.1139/cjfas-2015-0315>

367 Folt, C. L., Nislow, K. H., & Power, M. E. (1998). Implications of temporal and spatial scale for
368 Atlantic salmon (*Salmo salar*) research. *Canadian Journal of Fisheries and Aquatic*
369 *Sciences*, 55, 9–21.

370 Fukuda, S., & Hiramatsu, K. (2008). Prediction ability and sensitivity of artificial intelligence-
371 based habitat preference models for predicting spatial distribution of Japanese medaka
372 (*Oryzias latipes*). *Ecological Modelling*, 215(4), 301–313. Retrieved from
373 [http://www.sciencedirect.com/science/article/B6VBS-4SK2V3R-](http://www.sciencedirect.com/science/article/B6VBS-4SK2V3R-1/2/49220e5317c43e908ca2066127001686)
374 [1/2/49220e5317c43e908ca2066127001686](http://www.sciencedirect.com/science/article/B6VBS-4SK2V3R-1/2/49220e5317c43e908ca2066127001686)

375 Gerking, S. D. (1953). Evidence for the Concepts of Home Range and Territory in Stream
376 Fishes. *Ecology*, 34(2), 347–365. Retrieved from <http://www.jstor.org/stable/1930901>

377 Grant, J. W. A., Steingrimsson, S. O., Keeley, E. R., & Cunjak, R. A. (1998). Implications of
378 territory size for the measurement and prediction of salmonid abundance in streams.
379 *Canadian Journal of Fisheries and Aquatic Sciences*, 55, 181–190.

380 Guay, J. C., Boisclair, D., Rioux, M., Leclerc, M., Lapointe, M., & Legendre, P. (2000).
381 *Development and validation of numerical habitat models for juveniles of Atlantic salmon*
382 *(Salmo salar)* (Vol. 57). Ottawa, ON, CANADA: National Research Council of Canada.

383 Hedger, R. D., Dodson, J. J., Bergeron, N. E., & Caron, F. (2005). Habitat selection by juvenile
384 Atlantic salmon: the interaction between physical habitat and abundance. *Journal of Fish*
385 *Biology*, 67(4), 1054–1071. <http://doi.org/10.1111/j.0022-1112.2005.00808.x>

386 Heggenes, J. (1990). Habitat utilization and preferences in juvenile atlantic salmon (*salmo salar*)
387 in streams. *Regulated Rivers: Research & Management*, 5(4), 341–354.
388 <http://doi.org/10.1002/rrr.3450050406>

389 Heggenes, J., Bagliniere, J. L., & Cunjak, R. (1995). Note de synthèse sur la sélection de niche
390 spatiale et la compétition chez le jeune saumon Atlantique (*Salmo salar*) et la truite
391 commune (*Salmo trutta*) en milieu lotique. *Bulletin Français de La Pêche et de La*

392 *Pisciculture*, (337-338–339), 231–239. Retrieved from
393 <http://dx.doi.org/10.1051/kmae:1995026>

394 Heggenes, J., Bagliniere, J. L., & Cunjak, R. A. (1999). Spatial niche variability for young
395 Atlantic salmon (*Salmo salar*) and brown trout (*S-trutta*) in heterogeneous streams. *Ecology*
396 *of Freshwater Fish*, 8(1), 1–21.

397 Heland, M., & Dumas, J. (1994). Ecologie et comportements des juvéniles. In J. Guéguen & P.
398 Prouzet (Eds.), *Le Saumon Atlantique* (pp. 29–46). Ifremer.

399 Höjesjö, J., Kaspersson, R., & Armstrong, J. D. (2015). Size-related habitat use in juvenile
400 Atlantic salmon: the importance of intercohort competition. *Canadian Journal of Fisheries*
401 *and Aquatic Sciences*, 73(8), 1182–1189. <http://doi.org/10.1139/cjfas-2015-0446>

402 Johnston, T. A. (1997). Downstream movements of young-of-the-year fishes in Catamaran
403 Brook and the Little Southwest Miramichi River, New Brunswick. *Journal of Fish Biology*,
404 51(5), 1047–1062. <http://doi.org/10.1111/j.1095-8649.1997.tb01543.x>

405 Jorde, K., Schneider, M., Peter, A., & Zoellner, F. (2001). Fuzzy based models for the evaluation
406 of fish habitat quality and instream flow assessment. *Proceedings of the 3rd International*
407 *Symposium on Environmental Hydraulics*. Tempe, AZ.

408 Keeley, E. R., & Grant, J. W. A. (1995). Allometric and environmental correlates of territory size
409 in juvenile Atlantic salmon (*Salmo salar*). *Canadian Journal of Fisheries and Aquatic*
410 *Sciences*, 52(1), 186–196. <http://doi.org/10.1139/f95-019>

411 Knol, A., Slottje, P., van der Sluijs, J., & Lebrecht, E. (2010). The use of expert elicitation in
412 environmental health impact assessment: a seven step procedure. *Environmental Health*,
413 9(1), 19. Retrieved from <http://www.ehjournal.net/content/9/1/19>

414 Lewis, C. A., Lester, N. P., Bradshaw, A. D., Fitzgibbon, J. E., Fuller, K., Hakanson, L., &
415 Richards, C. (1996). Considerations of scale in habitat conservation and restoration.
416 *Canadian Journal of Fisheries and Aquatic Sciences*, 53(S1), 440–445.
417 <http://doi.org/10.1139/f96-021>

418 Lindeman, A. A., Grant, J. W. A., & Desjardins, C. M. (2015). Density-dependent territory size
419 and individual growth rate in juvenile Atlantic salmon (*Salmo salar*). *Ecology of Freshwater*
420 *Fish*, 24(1), 15–22. <http://doi.org/10.1111/eff.12120>

- 421 Mäki-Petäys, A., Erkinaro, J., Niemelä, E., Huusko, A., & Muotka, T. (2004). Spatial
422 distribution of juvenile Atlantic salmon (*Salmo salar*) in a subarctic river: size-specific
423 changes in a strongly seasonal environment. *Canadian Journal of Fisheries and Aquatic*
424 *Sciences*, 61(12), 2329–2338. <http://doi.org/10.1139/f04-218>
- 425 Mamdani, E. H. (1977). Application of fuzzy logic to approximate reasoning using linguistic
426 synthesis. *Proceedings of the Sixth International Symposium on Multiple-Valued Logic*.
427 Logan, Utah, United States: IEEE Computer Society Press.
- 428 McHugh, M. L. (2012). Interrater reliability: the kappa statistic. *Biochemia Medica*, 22(3), 276–
429 282. Retrieved from <http://www.ncbi.nlm.nih.gov/pmc/articles/PMC3900052/>
- 430 Mocq, J., St-Hilaire, A., & Cunjak, R. a. (2015). Influences of Experts' Personal Experiences in
431 Fuzzy Logic Modeling of Atlantic Salmon Habitat. *North American Journal of Fisheries*
432 *Management*, 35(2), 271–280. Retrieved from
433 <http://www.tandfonline.com/doi/full/10.1080/02755947.2014.996684>
- 434 Mocq, J., St-Hilaire, A., & Cunjak, R. A. (2013). Assessment of Atlantic salmon (*Salmo salar*)
435 habitat quality and its uncertainty using a multiple-expert fuzzy model applied to the
436 Romaine River (Canada). *Ecological Modelling*, 265, 14–25.
- 437 Morantz, D. L., Sweeney, R. K., Shirvell, C. S., & Longard, D. A. (1987). Selection of
438 Microhabitat in Summer by Juvenile Atlantic Salmon (*Salmo salar*). *Canadian Journal of*
439 *Fisheries and Aquatic Sciences*, 44(1), 120–129. <http://doi.org/10.1139/f87-015>
- 440 Mouton, A. M., Schneider, M., Peter, A., Holzer, G., Müller, R., Goethals, P. L. M., & De Pauw,
441 N. (2008). Optimisation of a fuzzy physical habitat model for spawning European grayling
442 (*Thymallus thymallus* L.) in the Aare river (Thun, Switzerland). *Ecological Modelling*,
443 215(1–3), 122–132. <http://doi.org/10.1016/j.ecolmodel.2008.02.028>
- 444 Muñoz-Mas, R., Papadaki, C., Martínez-Capel, F., Zogaris, S., Ntoanidis, L., & Dimitriou, E.
445 (2016). Generalized additive and fuzzy models in environmental flow assessment: A
446 comparison employing the West Balkan trout (*Salmo farioides*; Karaman, 1938). *Ecological*
447 *Engineering*, 91, 365–377. <http://doi.org/http://dx.doi.org/10.1016/j.ecoleng.2016.03.009>
- 448 R Development Core Team, R. (2016). R: A Language and Environment for Statistical
449 Computing. *R Foundation for Statistical Computing*. Vienna: R Foundation for Statistical
450 Computing. <http://doi.org/10.1007/978-3-540-74686-7>

- 451 Rykiel, E. J. (1996). Testing ecological models: the meaning of validation. *Ecological*
452 *Modelling*, 90(3), 229–244. [http://doi.org/10.1016/0304-3800\(95\)00152-2](http://doi.org/10.1016/0304-3800(95)00152-2)
- 453 Schoeneberger, P. J., Wysocki, D. A., Benham, E. C., Soil Survey Staff, & Natural Resources
454 Conservation Service, N. S. S. C. (2012). *Field Book for Describing and Sampling Soils,*
455 *version 3.0*. Lincoln, NE: Natural Resources Conservation Service, National Soil Survey
456 Center.
- 457 Shepard, R. B. (2005). *Quantifying Environmental Impact Assessment Using Fuzzy Logic*. (B. N.
458 Anderson, R. W. Howarth, & L. R. Walker, Eds.). New-York: Springer.
- 459 Shirvell, C. S. (1994). Effect of changes in streamflow on the microhabitat use and movements
460 of sympatric juvenile coho salmon (*Oncorhynchus kisutch*) and chinook salmon (*O.*
461 *tshawytscha*) in a natural stream. *Can. J. Fish. Aquat. Sci.*, 51(7), 1644–1652.
- 462 Wildman, T. L., & Neumann, R. M. (2003). Comparison of snorkeling and electrofishing for
463 estimating abundance and size structure of brook trout and brown trout in two southern New
464 England streams. *Fisheries Research*, 60(1), 131–139. [http://doi.org/10.1016/s0165-](http://doi.org/10.1016/s0165-7836(02)00060-7)
465 [7836\(02\)00060-7](http://doi.org/10.1016/s0165-7836(02)00060-7)
- 466