

1 **Seasonal variation of total mercury and condition indices of Arctic charr (*Salvelinus***  
2 ***alpinus*) in Northern Québec, Canada**

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18 **Abstract**

19           The winter ecology of anadromous Arctic charr has remained poorly detailed in the literature  
20 beyond descriptions of seasonal fasting and resulting declines in condition. However, prolonged periods  
21 of reduced feeding can have significant consequences for other variables, such as tissue contaminant  
22 levels. To more thoroughly detail seasonal changes, biological information (fork length, total weight, age,  
23 sex, somatic condition), stable isotopes ( $\delta^{13}\text{C}$ , % carbon,  $\delta^{15}\text{N}$ , % nitrogen), dorsal muscle % lipid,  
24 caloric densities, and total mercury (THg) concentrations were assessed in anadromous Arctic charr  
25 collected from Deception Bay, Canada, during the summer and over-wintering periods. Significant  
26 reductions in somatic condition, total weight, and % nitrogen, consistent with prolonged periods of  
27 fasting, were found for post-winter captured Arctic charr, but % lipid and caloric densities were  
28 significantly higher in these fish. THg also varied seasonally and was significantly higher in summer  
29 collected tissue. When tested individually via linear regression, significant relationships were seasonally  
30 dependent, but limited in number. All previously mentioned parameters were then incorporated into  
31 multi-variable models which better explained variations in the data. While there was no clear best model  
32 for explaining the % lipid values, caloric densities, and THg, season, condition, and stable isotope values  
33 (% carbon and % nitrogen) were the best indicators of % lipid content and caloric densities. THg  
34 concentrations were best explained by total weight, somatic condition, and  $\delta^{13}\text{C}$ . Seasonal variation in  
35 fish condition measures and THg may be indicative of condition selective mortality that yields apparent  
36 improvement through the disproportionate removal of poorer conditioned fish from the population during  
37 the over-wintering period. This hypothesis was further supported by mortality estimates and the results of  
38 the multi-predictor variable models. Collectively, this research highlights the importance of seasonal  
39 dynamics for anadromous Arctic charr populations and suggests that future studies further consider  
40 seasonality when evaluating this species.

41 **Keywords**

42 Arctic charr; total mercury; Nunavik; seasonal variation; condition; lipid content; caloric density

## 43 **1. Introduction**

44 Arctic charr (*Salvelinus alpinus*) are the most northerly distributed freshwater fish species on  
45 Earth with populations ranging from southern temperate locations in eastern North America and the  
46 European Alps to the northern extent of land masses in Eurasia and North America (Johnson 1980). These  
47 fish are considered habitat generalists, occupying lakes, streams, rivers, and marine environments  
48 depending on the time of year and life history form (Power et al. 2008). Arctic charr exhibit diverse life  
49 history strategies (Jonsson et al. 1988) that include anadromous, lacustrine, and partial migratory types,  
50 where anadromous and non-anadromous fish co-exist (Jonsson and Jonsson 1993), with the decision to  
51 migrate dictated by environmental conditions, sex, genetics, and ontogenetic niche shifts (Jonsson and  
52 Jonsson 1993; Jørgensen and Johnsen 2014; Klemetsen et al. 2003a). Both sexually mature and immature  
53 fish perform migrations and first time migrants can be anywhere from 3 to 8 years of age (Johnson 1989;  
54 Nordeng 1983).

55 Seaward migrations of anadromous Arctic charr begin after ice break up in the spring (Johnson  
56 1983; Klemetsen et al. 2003a), with fish generally remaining in coastal areas for 1-2 months before  
57 returning to freshwater environments (Klemetsen et al. 2003a; Mathisen and Berg 1968). The relative  
58 abundance of anadromous individuals is variable within and among populations (Johnson 1980; Svenning  
59 et al. 1992) and dependent on conditions, such as freshwater growth opportunities and differences in  
60 productivity between freshwater and marine environments (Gross et al. 1988) and/or the physical  
61 characteristics of the migratory route (Kristoffersen 1994; Moore et al. 2016). However, at the northern  
62 extremes of their distribution lacustrine residency is favoured regardless of access to the marine  
63 environment (Power et al. 2008).

64 For Arctic charr, the period of marine residency is characterized by rapid growth, with  
65 anadromous fish growing faster than resident freshwater fish (Berg and Berg 1989; Johnson 1980; Moore  
66 and Moore 1974). Fish may double their body weight and experience a five-fold increase in lipid content  
67 while in the marine environment (Finstad and Heggberget 1993; Jørgensen et al. 1997; Mathisen and Berg

68 1968), with the carcass (head, skeleton, and skin) and muscle tissue accounting for 50% and 35-40%,  
69 respectively, of the total body lipid content when fish re-enter freshwater from the sea (Jobling et al.  
70 1998; Jørgensen et al. 1997; Jørgensen and Johnsen 2014). During the marine period fish feed  
71 opportunistically on zooplankton, amphipods, pelagic and benthic fishes (Dempson et al. 2002; Grønvik  
72 and Klemetsen 1987; Power et al. 2008).

73 Historically, studies of anadromous Arctic charr have focused on the period of marine residency  
74 and its consequences for growth and maturation (Berg and Berg 1989; Jørgensen et al. 1997; Murdoch et  
75 al. 2015), but have traditionally lacked winter data (Mulder et al. 2018b). The available over-wintering  
76 data that does exist have focused on resident lacustrine Arctic charr condition, diet, and habitat during the  
77 ice-covered months (Amundsen and Knudsen 2009; Eloranta et al. 2013; Klemetsen et al. 2003b). Thus  
78 the ecology of over-wintering anadromous Arctic charr has remained poorly described in the literature  
79 beyond the noted cessation of feeding and the resulting reduction of condition (e. g. (Boivin and Power  
80 1990; Jørgensen et al. 1997; Rikardsen et al. 2003). More recently, telemetry studies of over-wintering  
81 Arctic charr in southern Labrador have indicated reduced activity patterns (Mulder et al. 2018b) and  
82 temperature-dependent diurnal movement patterns (Mulder et al. 2019) occurring within a narrow range  
83 of temperatures (0.5–2 °C) that suggest a use of strategies to lower metabolic costs and minimize over-  
84 winter energy expenditure (Mulder et al. 2018a).

85 During winter, whole body lipid body reserves have been reported to decline by up to 30% for  
86 non-reproductive anadromous Arctic charr and between 35 – 46% for post-spawning individuals (Dutil  
87 1986; Jørgensen et al. 1997). Lipids are depleted from all tissues, but mobilization of muscle and carcass  
88 deposits has been determined to be the most significant during this period (Jobling et al. 1998; Jørgensen  
89 et al. 1997). Additionally, emaciation is greatest in females with individuals on average losing  
90 approximately 80% of their lipid stores during spawning and overwintering (Jobling et al. 1998;  
91 Jørgensen et al. 1997). The documented declines in energy reserves, the minimization of movement, use  
92 of colder water temperatures to reduce metabolism (Mulder et al. 2018b), lack of suitable prey (Boivin  
93 and Power 1990), and the implications of light restrictions on prey capture efficiency (Mazur and

94 Beauchamp 2003; Turesson and Brönmark 2007) all suggest significantly reduced feeding during the  
95 winter months. While in lacustrine populations of Arctic charr, winter feeding has been documented to  
96 some extent in the literature (e.g., (Eloranta et al. 2013; Klemetsen et al. 2003b; Power et al. 2009)), it has  
97 only been inferred for anadromous individuals based on activity and habitat use (Boivin 1987; Mulder et  
98 al. 2018b).

99         In addition to losses in lipids and overall reductions in condition, prolonged periods of reduced  
100 feeding can have other significant consequences for over-wintering fish. Critical among those effects are  
101 the possible associated changes in tissue contaminant levels. Jørgensen et al. (2006) demonstrated that  
102 winter fasting and subsequent emaciation of anadromous Arctic charr resulted in the redistribution of  
103 lipophilic PCBs to sensitive organs (e.g., liver and brain). This was accompanied by significant increases  
104 in hepatic biomarker activity and disruptions to endocrine mechanisms, immune function, and  
105 smoltification processes. The results suggests that seasonal lipid dynamics may result in increased  
106 sensitivity of Arctic animals to certain contaminants undergoing seasonal fasting (Jørgensen et al. 2002;  
107 Jørgensen et al. 2006). Methylmercury (MeHg) is one contaminant of specific concern due to its  
108 neurologically toxic health effects (Mergler et al. 2007) and its ability to bioaccumulate in aquatic food  
109 webs to reach high concentrations in large-bodied predatory fish (Gantner et al. 2010; Kidd et al. 1995;  
110 van der Velden et al. 2013a). In the muscle tissue of these fish, MeHg concentrations are typically greater  
111 than 90% of total mercury, THg, (Eagles-Smith et al. 2016; Hall et al. 1997; Lockhart et al. 2005), and  
112 generally increase with size and age (Lescord et al. 2018), making THg concentrations a suitable proxy  
113 for MeHg levels and of particular concern to the Inuit people who consume large quantities of fish  
114 (Lemire et al. 2015).

115         Measured concentrations of THg are often related to biological variables such as: fish size, age,  
116 trophic position ( $\delta^{15}\text{N}$ ), feeding strategies and habitat use ( $\delta^{13}\text{C}$ ), growth, and somatic condition (Dittman  
117 and Driscoll 2009; Power et al. 2002; Wiener et al. 2003), but also depend on lipid and protein contents  
118 (Eisler 1987; Kahilainen et al. 2016; Swanson and Kidd 2010). THg in fish tissues is derived almost  
119 exclusively from prey consumption (Hall et al. 1997) and after digestion THg is translocated to the liver

120 via blood and subsequently stored in muscle tissues (Oliveira Ribeiro et al. 1999; Wang and Wang 2015)  
121 where it is almost exclusively bound within proteins (Amlund et al. 2007; Bury et al. 2003; Mason et al.  
122 1995). Growth during the summer can lower THg concentrations during the growing season as  
123 individuals accumulate tissue faster than THg (Karimi et al. 2007; Lepak et al. 2012; Olk et al. 2016),  
124 while lipid losses prompted by a cessation of feeding during the winter months can result in a  
125 phenomenon termed starvation – concentration (Cizdziel et al. 2003; Cizdziel et al. 2002) that increases  
126 mercury in remaining protein-containing tissues (Kahilainen et al. 2016). The result is higher THg during  
127 the ice-covered period (Keva et al. 2017; Olk et al. 2016). While seasonal variations in THg have been  
128 reported in lacustrine resident Arctic charr and other fish species (Kahilainen et al. 2016; Keva et al.  
129 2017; Olk et al. 2016), such variations are relatively understudied and have yet to be documented for  
130 anadromous Arctic charr. Additionally, quantifying seasonal variations in THg concentrations in  
131 anadromous Arctic charr may help determine differences in seasonal risks for capture and consumption  
132 associated with the winter fishery for this species, which provides an important component of the year-  
133 round diet of Inuit people, including those of the Nunavik region of northern Québec (e.g., (Boivin and  
134 Power 1990)). Similar to other areas of Arctic Canada, anadromous Arctic charr in Deception Bay,  
135 Nunavik, Québec, are of significant cultural and economic importance to the local Indigenous  
136 communities (e.g., Salluit and Kangiqsujuaq). However, literature on Arctic charr populations in this  
137 region remains scarce (e.g., (Boivin and Power 1990; Murdoch et al. 2013; Murdoch and Power 2013)),  
138 with only one study having directly examined issues related to seasonality (Boivin and Power 1990).

139         Here samples collected from Deception Bay, Nunavik, where Arctic charr migrate and feed in  
140 summer, and lakes Duquet and François-Malherbe, where Arctic charr spawn and over-winter, are used  
141 to investigate patterns of seasonal change in fish condition measures and THg concentrations.  
142 Specifically, it was hypothesized that (i) over-wintering anorexia and fasting (e.g., as represented by  
143 reduced total weight and somatic condition (Fulton’s K) (Jobling 1981; Jørgensen et al. 1997) would  
144 result in significant declines in percent lipid (% lipid) and caloric densities of anadromous Arctic charr  
145 collected during post-winter sampling as stored energy reserves are mobilized to meet metabolic demands

146 during this period (Jobling et al. 1998; Jørgensen et al. 1997). Additionally, seasonal relationships with  
147 co-measured biological information (fork length, total weight, age, sex, and somatic condition) and stable  
148 isotopes used as proxies for feeding behaviour, habitat use, and trophic position ( $\delta^{13}\text{C}$ , % carbon (C),  
149  $\delta^{15}\text{N}$ , % nitrogen (N)) were evaluated.

150 Samples were also used to test a contamination hypothesis: (ii) that THg values would be higher  
151 during the post-winter sampling period than in the late summer sampling period as a result of fasting-  
152 induced emaciation and the subsequent increase of mercury in the remaining tissues (Kahilainen et al.  
153 2016; Keva et al. 2017; Olk et al. 2016). Again, seasonal correlations with biological information and  
154 stable isotope ratios were assessed, with specific relationships between measured THg concentrations %  
155 lipid and caloric density values being evaluated. Finally, data were used to model dorsal muscle % lipid  
156 content, caloric density, and THg concentrations as a function of multi-variable statistical models  
157 inclusive of combinations of the above tested biological information, stable isotope values, and their  
158 interactions. Specifically, we hypothesized that (iii) season and/or fish condition would best describe %  
159 lipid values and caloric content (Dutil 1986; Thompson et al. 1991; Todd et al. 2008) and variables  
160 associated with bioaccumulation (fork length, total weight and/or age and trophic position represented by  
161  $\delta^{15}\text{N}$ ) (Cizdziel et al. 2002; Power et al. 2002; van der Velden et al. 2013a) and season (Kahilainen et al.  
162 2016; Keva et al. 2017; Köck et al. 1996), would best describe THg concentrations as these variables are  
163 known to significantly influence measured THg concentrations in fish tissue (Kahilainen et al. 2016;  
164 Wiener et al. 2003).

## 165 **2. Methods**

### 166 *2.1 Study Area*

167 The Deception River and its tributaries (Fig. 1) are located east of Salluit, Québec ( $62^{\circ}10'46\text{ N}$ ,  
168  $75^{\circ}40'13\text{ W}$ ) with a watershed spanning an area of  $3870\text{ km}^2$  between latitudes  $61^{\circ}31'26''\text{ N}$  and  
169  $62^{\circ}11'01''\text{ N}$ . The river flows into Deception Bay, which is associated with the Hudson Strait marine  
170 ecosystem (Goldsmit et al. 2014), a deep and wide channel that connects Hudson Bay and the Foxe Basin



171 with the Labrador Sea and the Davis Strait. Average daily temperatures range from -25.6°C in February,  
172 to 10.5°C in August (Environment Canada 2018a; Environment Canada 2018b), with a growing season of  
173 less than 120 days per year. In addition to traditional hunting and fishing, the area is impacted by two  
174 nickel and copper mining projects: Glencore - the Raglan Mine Project and Canadian Royalties Inc. – the  
175 Nunavik Nickel Project, and is proximate to the now closed Asbestos Hill Mine (Purtiniq). A 95  
176 kilometre road connects the main Raglan mining site with an additional camp and deep-water port in  
177 Deception Bay. Mine personnel are present year-round and the road closely follows the Deception River  
178 and its tributaries for most of its length. Some contaminant input is believed to result from the mining  
179 facilities proximate to Deception Bay in addition to atmospheric deposition.

180 Arctic charr spawn and overwinter in headwater lakes Duquet (62°03'18 N, 74°31'51 W) and  
181 François-Malherbe (62°00'06 N, 74°15'25 W) from October to June. Lake Duquet is less than half the  
182 size of Lake François-Malherbe and the lakes are located 2.5 km and 15 km, respectively, upstream of  
183 the river mouth. There is usually a commercial fishing permit active for Deception Bay and both lakes,  
184 and a Raglan Mine sport fishing permit is active for Deception Bay. Lake Watts, a third lake in the  
185 Deception River system upstream of Lake François-Malherbe, is assumed to have received some direct  
186 input of mining waste prior to the 1984 closure of the Asbestos Hill Mine located 10 km from the lake.  
187 Traditional knowledge suggests that Arctic charr are predominately non-migratory in Lake Watts and  
188 access to this lake was not possible during the course of this study.

## 189 *2.2 Sample Collection*

190 Summer collection of anadromous Arctic charr was completed using multi-mesh experimental  
191 gill nets (25 – 150 mm mesh panels with a length of 120 m and a hanging depth of 2 m) set coincident  
192 with the returning upstream migration period. Arctic charr were caught at eight locations equally  
193 distributed throughout Deception Bay and from the mouth of the Deception River in August of 2016. A  
194 post-winter sample was obtained from lakes François-Malherbe (May 8<sup>th</sup> – 11<sup>th</sup>, 2017) and Duquet (May  
195 12<sup>th</sup> – 13<sup>th</sup>, 2017), approximately a month prior to ice break-up (as reported for the area by the Canadian  
196 Ice Service (2018) as June 26<sup>th</sup>, 2017). Post-winter dorsal muscle samples were collected from fish

197 captured throughout the lakes with jigging lines by Nunavik Research Centre (NRC) staff in collaboration  
198 with Inuit elders during the Salluit community Elder's Spring Fishing Event hosted by Qaqqalik  
199 Landholding Corporation and supported by Raglan Mine. All fish captured were sacrificed after capture  
200 with a sharp blow to the head and all sample collection was performed in accordance with standards  
201 dictated by the Ministère des Forêts, de la Faune, et des Parcs – Direction de la gestion de la faune du  
202 Nord-du-Québec (permis de gestion de la faune #2016-02-199-152-10-G-P N/D : 9053\_36).

203 Fish were sexed, weighed ( $\pm 1$  g) and measured for fork length ( $\pm 1$  mm). Measurements were  
204 used to calculate Fulton's condition factor ( $K = 10^5 * W/L^3$ ), which was only determined after performing  
205 standardized weight-length regressions and ensuring that the slope of this regression did not significantly  
206 deviate from a value of three (Ricker 1975). A sample of dorsal muscle tissue (mass  $\approx 10$  g wet weight)  
207 was removed from above the lateral line and posterior to the dorsal fin on the left side of each Arctic charr  
208 (van der Velden et al. 2013a) and frozen at  $-20^\circ\text{C}$  for subsequent laboratory analyses. A random sample  
209 of 50 fish from the summer sampling was selected for THg analysis. A sub-sample of 40 fish was used  
210 for lipid analysis and a sub-sample of 30 fish was used for bomb calorimetry. From the winter sampling a  
211 total of 55 fish dorsal muscle samples were provided by the Salluit community Elder's Spring Fishing  
212 Event. To guard against inadvertent use of resident fish from the post-winter sampling, anadromy was  
213 confirmed with  $\delta^{34}\text{S}$  stable isotope analysis following methods described in Doucett et al. (1999). After  
214 removal of resident fish, the remaining anadromous fish were used for THg analyses. Similar to the  
215 summer sampling, a sub-sample was retained for lipid and bomb calorimetry analyses. Dorsal muscle  
216 samples were generally used for all (lipid, bomb calorimetry and THg) analyses except where tissue was  
217 limited.

218 To minimize accidental metal contamination, tissues for THg analysis were placed in Eppendorf  
219 polypropylene tubes. These tubes had been acid washed in 15%  $\text{HNO}_3$  for at least 24 hours before being  
220 rinsed 5 times with distilled water, twice with ultrapure water, and then dried under a laminar-flow fume  
221 hood before use. The upper gastrointestinal tracts (e.g., esophagus and stomach) of all post-winter  
222 collected fish were examined for evidence of short-term winter feeding. Aging of all sampled fish was

223 completed by NRC staff, with fish ages determined by submersing the otolith in water and examining the  
224 distal surface with reflected light under a dissecting microscope (Chilton and Beamish 1982). Ages were  
225 used to estimate von Bertalanffy growth equations (Ricker 1975) for each seasonal sample from which  
226 the von Bertalanffy growth coefficient (K) was used to estimate mortality following (Jensen 1996).

### 227 *2.3 Stable Isotope Analysis*

228 For stable isotope analyses, dorsal muscle tissue was dried at 50°C for 48 h and pulverized into a  
229 homogenate powder with a mortar and pestle. After being weighed to 0.275 – 0.300 mg (UMX2, Mettler-  
230 Toledo GmbH, Greifensee, Switzerland), each tissue sample was simultaneously analyzed for carbon and  
231 nitrogen stable isotope ratios ( $\delta^{13}\text{C}$ ,  $\delta^{15}\text{N}$ ) at the Environmental Isotope Laboratory, University of  
232 Waterloo, Ontario, Canada with a Delta Plus Continuous Flow Stable Isotope Ratio Mass Spectrometer  
233 (Thermo Finnigan, Bremen, Germany) coupled to a Carlo Erba Elemental Analyzer (CHNS-O EA1108,  
234 Carlo Erba, Milan, Italy) following methods described in van der Velden et al. (2013a). Elemental  
235 compositions were expressed in percentage terms based on pre-analysis weights. All stable isotope  
236 measurements were expressed using standard delta notation ( $\delta$ ) as parts per thousand differences (‰)  
237 with respect to the international reference standards of Vienna Peedee Belemnite carbonate rock for  $\delta^{13}\text{C}$   
238 (Craig 1957) and nitrogen gas in the atmosphere for  $\delta^{15}\text{N}$  (Mariotti 1983):

$$239 \quad \delta R = \left[ \frac{R_{\text{sample}} - R_{\text{standard}}}{R_{\text{standard}}} \right] \times 1000$$

240 where  $\delta R$  is the measured carbon ( $^{13}\text{C}/^{12}\text{C}$ ) or nitrogen ( $^{15}\text{N}/^{14}\text{N}$ ) isotope ratio expressed with respect to  
241 the appropriate international standard. Machine analytical precision was determined to be  $\pm 0.2\text{‰}$  for  $\delta^{13}\text{C}$   
242 and  $\pm 0.3\text{‰}$  for  $\delta^{15}\text{N}$  and was established via repeat analysis of internal laboratory working standards  
243 (IAEA-N<sub>1</sub> + N<sub>2</sub>, IAEA-CH<sub>3</sub> + CH<sub>6</sub>) cross calibrated to International Atomic Energy Agency (IAEA)  
244 standards: CH<sub>6</sub> for  $\delta^{13}\text{C}$  and N<sub>1</sub> and N<sub>2</sub> for  $\delta^{15}\text{N}$ . Internal standards were placed at the beginning, middle  
245 and end of every run of samples and repeatability was assessed by repeat analysis of 1 in 10 samples. As  
246 C:N ratios were consistently below the 4.0 threshold above which lipid extraction is required (Jardine et

247 al. 2011; Logan et al. 2008; Sanderson et al. 2009),  $\delta^{13}\text{C}$  values were not lipid extracted or mathematically  
248 normalized for lipid content.

#### 249 *2.4 Lipid Analysis and Bomb Calorimetry*

250 Lipid analysis and bomb calorimetry were performed at the University of Waterloo, Waterloo,  
251 Ontario, Canada. A modified version of the procedure outlined in Folch (1957) was used for lipid  
252 extraction as this method provides accurate estimates of lipid content when lipids comprise greater than  
253 2% of tissue (Iverson et al. 2001). After freeze drying (Freezone Plus 2.5 Liter Cascade Benchtop Freeze  
254 Dry Systems, Labconco, Kansas City, USA), Arctic charr dorsal muscle tissue was ground with a mortar  
255 and pestle and weighed to approximately 30.0 mg  $\pm$  0.1 mg (XS205DU Analytical Balance, Mettler  
256 Toledo, Mississauga, Canada). 2 mL of a 2:1 chloroform-methanol solution and 1.6 mL of a 0.9% KCl  
257 solution were then added to the ground tissue. The resulting solution was homogenized with a vortex  
258 (Fisherbrand Analog Vortex Mixer, Fisher Scientific, Hampton, USA) and centrifuged  
259 (Fisherbrand™ Centrif Model 225A Benchtop Centrifuge, Fisher Scientific, Hampton, USA) at 2000  
260 RPM for 5 minutes until the KCl, tissue, and chloroform-methanol layers were completely separated. The  
261 lipid containing solution was extracted via Pasteur pipette through the KCl and residual biomass layers.  
262 Three iterations of the procedure were performed until a final lipid solution of 8 mL was obtained. The  
263 lipid-containing solution was then evaporated to dryness. Once dry, an additional 2 mL of 2:1  
264 chloroform-methanol solution were added to the dried material and two, 100  $\mu\text{L}$  aliquots were then  
265 transferred to pre-weighed tin cups. The solution was evaporated at room temperature overnight until only  
266 dry lipids remained. Remains were weighed on a micro-balance (UMX2, Mettler-Toledo GmbH,  
267 Greifensee, Switzerland) to determine the percentage of lipid in the dorsal muscle tissue expressed as:

$$268 \quad \% \text{ Lipid} = \left( \left( \frac{\text{Mass}_{\text{Dry}} (\text{g})}{\text{Mass}_{\text{Ground}} (\text{g})} \right) \right) * (1 - P_{\text{Water}}) * 100\% * 20;$$

269 where  $\text{Mass}_{\text{Dry}}$  is the weight of the dried lipid following the extraction procedure,  $\text{Mass}_{\text{Ground}}$  is the initial  
270 ground mass of the tissue preceding extraction,  $P_{\text{Water}}$  is the proportion of water in the analyzed dorsal  
271 muscle tissue (wet tissue mass (g) – dry tissue mass (g)), and 20 represents a correction as a result of

272 using only a subset of the final extraction solution volume for establishing the final dried mass (Folch  
273 1957).

274 For bomb calorimetry, Arctic charr dorsal muscle tissue was dried at 50°C for 48 hours and  
275 ground to obtain a homogenized sample using a mortar and pestle. Pellets were formed (Parr Pellet Press,  
276 Parr Instrument Company, Moline, USA) with weights not exceeding 50.0 mg  $\pm$  0.1 mg (XS205DU  
277 Analytical Balance, Mettler Toledo, Mississauga, Canada) before ignition in a Parr Semi-micro  
278 Calorimeter 6725 (Parr Instrument Company, Moline, USA) to measure caloric density (cal·g<sup>-1</sup> dry mass).  
279 The wet mass caloric density was determined by multiplying the dry mass caloric density by the  
280 proportion of final dry mass to original wet mass (Glover et al. 2010). Results were reported as wet  
281 weight (ww) caloric density means  $\pm$  standard deviation. Benzoic acid pellets with a caloric density of  
282 6318.4 cal·g<sup>-1</sup> dry weight (dw) were used to standardize the calorimeter and assess recovery every 10<sup>th</sup>  
283 sample. Percent recovery  $\pm$  standard deviation of benzoic acid pellets ( $n = 11$ ) was determined to be  
284 100.35%  $\pm$  1.07%.

### 285 *2.5 THg Analysis*

286 Mercury analysis was performed at the Institut National de la Recherche Scientifique (INRS) in  
287 Québec City, Québec, Canada and at the University of Waterloo in Waterloo, Ontario, Canada. After  
288 freeze-drying (FTS Systems TMM, Kinetics Thermal Systems, Longueil, QC, Canada; Freezone Plus 2.5  
289 Liter Cascade Benchtop Freeze Dry Systems, Labconco, Kansas City, USA), tissue was weighed to  
290 approximately 50.0 – 100.0 mg  $\pm$  0.1 mg (Series 321 LT 220A Balance, Precisa Gravimetrics AG,  
291 Dietikon, Switzerland; Mettler Toledo, Mississauga, Canada). Analysis was then performed with a direct  
292 mercury analyzer (DMA-80, Milestone Inc., Shelton, USA) which enables assessment of THg through  
293 thermal decomposition followed by atomic absorption spectroscopy as described in U.S. EPA method  
294 7473 (United States Environmental Protection Agency 1998). Results were converted from dry weight to  
295 wet weight using percent moisture calculations determined from weights ( $\pm$  0.1 mg) taken before and  
296 after lyophilisation (Eikenberry et al. 2015).

297 Past studies have indicated that age (van der Velden et al. 2013b) and/or length (Rigét et al. 2000;  
298 Swanson et al. 2011) are often strongly correlated with THg concentrations such that THg concentrations  
299 require age or length adjustment to permit comparisons among individuals. Accordingly, data were  
300 examined for evidence of significant length and/or age correlations using linear regression ( $\log_{10}$ THg vs.  
301  $\log_{10}$ fork length, fork length, or  $\log_{10}$ age, and age) (Tran et al. 2015). The effect of lipids on analytical  
302 results was examined by comparing differences in relationships obtained using raw and lipid corrected  
303 THg concentrations, with lipid corrected THg concentrations computed following methods described in  
304 Kahilainen et al. (2016).

305 Method detection limits and percent recoveries are reported as mean percentage of certified  
306 values  $\pm$  standard deviation. Tissues were evaluated in triplicate with certified reference materials from  
307 the National Research Council of Canada (NRCC) (TORT-3, DOLT-4, and DOLT-5). Blanks were used  
308 every fifth sample in the same analytical cycle to establish accuracy and recovery rates. The method  
309 detection limit, determined as  $3\times$  the standard deviation of blanks, was 2.91 ng Hg (approximately 0.003  
310 ( $\text{mg}\cdot\text{kg}^{-1}$ )) and mean relative standard deviation of the triplicates was 5.53% ( $n = 108$ ). Percent recoveries  
311 were determined to be  $99.32\% \pm 7.42\%$  ( $n = 52$ ),  $93.16\% \pm 9.04\%$  ( $n = 47$ ), and  $88.95\% \pm 3.88\%$  ( $n = 16$ )  
312 for TORT-3, DOLT-4, and DOLT-5 respectively.

### 313 *2.6 Statistical Analysis*

314 All statistical analyses were performed using JMP® statistical software (v. 13.0.0, SAS Institute,  
315 CA) and Type I error was set to  $\alpha = 0.05$ . Data consistency with normality and homoscedasticity  
316 assumptions were verified using residual diagnostic histograms, visual assessment of Q-Q plots, and the  
317 Shapiro-Wilk W test (Shapiro and Wilk 1965). Data that did not meet parametric assumptions were  $\log_{10}$   
318 transformed (Zar 2007). Linear regressions were used to determine the relationship significance between  
319 specific variables and outliers that may have unduly influenced regression results were assessed using  
320 Cook's Distance statistic (Cook 1977) and subsequently removed. Un-paired, two-sample t-tests adjusted  
321 for homogeneity of variance assumptions were used to determine significant differences among seasons  
322 (Zar 2007) and the Wilcoxon approach was used when data did not conform to the required parametric

323 assumptions (Zar 2007). The Fisher exact test was used to test for significant differences among  
324 proportions (Zar 2007) and incidental correlations were assessed using Pearson's correlation coefficient.

325         General linear models (GLM) inclusive of season, fork length, total weight, age, somatic  
326 condition,  $\delta^{13}\text{C}$ , % C,  $\delta^{15}\text{N}$  and % N were used to determine the best model to describe dorsal muscle %  
327 lipid values, caloric densities, and THg concentrations. Pearson correlation analysis (Appendix A Tables  
328 A.1 – A.3) and/or linear regression were used to establish the significance of variables expected to  
329 correlate with dorsal muscle % lipid values, caloric densities, and THg concentrations, with only  
330 significant variables being retained for use in GLM models. Significant two-way interactions among  
331 variables were similarly assessed, with significant interactions retained for use in GLM models. To reduce  
332 statistical issues associated with multicollinearity (Zar 2007), possible explanatory variables that were  
333 highly correlated ( $r^2 > 0.70$ ) or known to have biological redundancy (e.g., fork length and total weight)  
334 were not included in the same model. To predict seasonal dorsal muscle % lipid values, caloric densities,  
335 or THg concentrations models that included all possible combinations of the feasible set of significant  
336 explanatory variables and two-way interactions, as determined above, were considered.

337         Model selection was performed using the Akaike Information Criteria adjusted for small sample  
338 sizes (*AICc*). The model with the lowest *AICc* score was considered the most accurate, except in  
339 circumstances where *AICc* values differed by less than two (Burnham and Anderson 2003). Detailed  
340 description of the computation of the analytical metrics associated with the *AICc* methodology used here  
341 can be found in Burnham and Anderson (2003).

### 342 **3. Results**

343         Of the 105 anadromous Arctic charr samples obtained, 100 were used in the comparison of  
344 summer and post-winter levels of  $\log_{10}$  THg,  $\log_{10}$  % lipids, and  $\log_{10}$  caloric densities. A single sample  
345 from the summer sampling was removed from use because of sample desiccation prior to shipment.  
346 Testing with  $\delta^{34}\text{S}$  stable isotope analyses indicated 4 Arctic charr from the post-winter collection were

347 non-anadromous and these fish were excluded from further consideration. This resulted in the loss of 4  
348 Arctic charr samples from the lipid testing subset and 3 samples from the calorimetry testing subset.

### 349 *3.1 Seasonal Variation in Biological Variables and Stable Isotope Values*

350 Means  $\pm$  standard deviations and ranges for biological variables and stable isotope values for all  
351 summer sampled and post-winter captured Arctic charr can be found in Table 1. Statistical testing of the  
352 slope of the standardized weight-length regression (Deception Bay sampling – 3.12, post-winter sampling  
353 – 3.03) confirmed isometric growth (Deception Bay t-test,  $p = 0.457$ , post-winter sampling t-test,  $p =$   
354  $0.255$ ), thereby allowing the use of Fulton's K. Estimated seasonal von Bertalanffy growth models yielded  
355 von Bertalanffy growth coefficients from which differing mortality rates were estimated (summer  
356 mortality: 22.3%; winter mortality: 24.5%).  $\log_{10}$  somatic condition significantly declined in post-winter  
357 Arctic charr ( $Z_{(1,98)} = 4.637$ ,  $p < 0.001$ ) indicating possible over-wintering anorexia, as a result of a  
358 significant decline in  $\log_{10}$  total weight during the over-wintering period ( $Z_{(1,98)} = 2.444$ ,  $p = 0.015$ ).  
359  $\log_{10}$  fork length ( $t_{(1,98)} = 1.43$ ,  $p = 0.156$ ),  $\log_{10}$  age ( $t_{(1,95)} = 1.309$ ,  $p = 0.194$ ), % carbon ( $Z_{(1,98)} = -0.831$ ,  
360  $p = 0.406$ ), and  $\delta^{15}\text{N}$  values ( $t_{(1,98)} = -1.258$ ,  $p = 0.211$ ) of captured fish did not vary seasonally.  $\delta^{13}\text{C}$  was  
361 significantly more depleted in post-winter sampled Arctic charr ( $Z_{(1,98)} = -5.754$ ,  $p < 0.001$ ) as was % N  
362 ( $Z_{(1,98)} = 5.623$ ,  $p < 0.001$ ). The sex ratio of summer and post-winter sampled fish did not vary between  
363 seasons (Fisher's exact  $\Pi_2 = 0.440$ ,  $p < 0.657$ ) and sex was not significantly correlated with % lipids,  
364 caloric density, or THg in any of the seasonal samples (all correlation  $p > 0.466$ ). Furthermore, no  
365 evidence of short-term winter feeding was determined via examination of the upper gastrointestinal tract.

### 366 *3.2 Lipid Content and Caloric Density*

367 Means  $\pm$  standard deviations and ranges for % lipid values and caloric densities are given in  
368 Table 2. Significant seasonal variation existed for  $\log_{10}$  lipid content ( $t_{(1,74)} = -6.49$ ,  $p < 0.001$ ) and  $\log_{10}$   
369 caloric density ( $t_{(1,55)} = -11.70$ ,  $p < 0.001$ ) (Fig. 2).  $\log_{10}$  lipid content explained 28% of the variation in  
370  $\log_{10}$  caloric density of Arctic charr during the post-winter fishery ( $r^2 = 0.28$ ,  $p = 0.005$ ), but values were  
371 not significantly correlated in Arctic charr captured returning from the marine environment ( $r^2 = 0.09$ ,  $p =$



0.118). Significant relationships between  $\log_{10}$  % lipid values and  $\log_{10}$  caloric densities and studied correlates are plotted in Fig. 3 and Fig. 4. A limited number of weak significant relationships were found and there was no consistent pattern of associations when comparing summer and post-winter samples.  $\log_{10}$  % lipids were not significantly correlated with any of the biological parameters (e.g., length, weight). However,  $\log_{10}$  % lipid values of late summer migrants significantly declined with increased offshore feeding ( $r^2 = 0.12, p = 0.030$ ) and were generally related to % elemental composition, increasing with % C in post-winter sampled fish ( $r^2 = 0.11, p = 0.048$ ) and decreasing with % N in both summer and post-winter sampled fish (summer ( $r^2 = 0.14, p = 0.017$ ); post-winter ( $r^2 = 0.12, p = 0.035$ )) (Fig. 3).  $\log_{10}$  caloric density increased with fork length in summer sampled Arctic charr ( $r^2 = 0.16, p = 0.031$ ) (Fig. 4) and in post-winter sampled fish was positively related to % C ( $r^2 = 0.30, p = 0.003$ ) and negatively related to % N ( $r^2 = 0.43, p < 0.001$ ) (Fig. 4).

### 3.3 THg

Mean THg concentrations  $\pm$  standard deviations are reported in Table 2 and all measured concentrations were determined to be below the regulatory  $0.5 \text{ mg}\cdot\text{kg}^{-1}$  of mercury detailed in Health Canada's commercial guideline of maximum levels for chemical contaminants in foods (Government of Canada 2019; Health Canada 2018). Contrary to what was hypothesized,  $\log_{10}$  THg concentrations of summer captured Arctic charr were significantly higher than those obtained from post-winter fish ( $t_{(1,98)} = 2.59, p = 0.011$ ) (Fig. 2). Significant relationships between  $\log_{10}$  THg concentrations and tested variables in both summer and post-winter sampled Arctic charr are plotted in Fig. 5. Regressions of fork length, weight and age against  $\log_{10}$  THg concentrations in summer or post-winter sampled fish were either non-significant or yielded regressions that poorly explained variation in  $\log_{10}$  THg concentrations (summer –  $\log_{10}$  THg vs. age ( $r^2 = 0.20, p = 0.001$ ); post-winter –  $\log_{10}$  THg vs. fork length ( $r^2 = 0.09, p = 0.034$ ); post-winter –  $\log_{10}$  THg vs. weight ( $r^2 = 0.14, p = 0.006$ )). Additionally, significant heteroscedasticity indicative of their statistical inadequacy was displayed, as such there was no compelling statistical evidence for length or age standardization of the  $\log_{10}$  THg concentrations data. Examination of differences in the relationships between tested variables and raw and altered  $\log_{10}$  THg values also

398 provided no compelling evidence for the correction of  $\log_{10}$  THg concentrations for lipids following  
399 (Kahilainen et al. 2016), as relationships between the parameters did not significantly change after  
400 transformation. Therefore, raw  $\log_{10}$  transformed results were used in all subsequent statistical analyses.  
401 Weak and seasonally differentiated patterns of association between  $\log_{10}$ THg and stable isotope data  
402 ( $\delta^{13}\text{C}$ , % C,  $\delta^{15}\text{N}$ ) were found (Fig. 5).  $\log_{10}$  THg increased with  $\delta^{13}\text{C}$  ( $r^2 = 0.23$ ,  $p < 0.0001$ ) in summer  
403 and  $\delta^{15}\text{N}$  in post-winter sampled fish ( $r^2 = 0.09$ ,  $p = 0.031$ ), and decreased with % C only in summer  
404 sampled Arctic charr ( $r^2 = 0.19$ ,  $p = 0.002$ ). Significant relationships between  $\log_{10}$ THg concentrations  
405 and  $\log_{10}$  % lipid values or  $\log_{10}$  caloric densities showed significant association (negative) only with  $\log_{10}$   
406 % lipids (Fig. 5) in late summer captured Arctic charr ( $r^2 = 0.20$ ,  $p = 0.018$ ).

### 407 *3.4 General Linear Models*

408 Models best explaining variation in  $\log_{10}$  % lipid values,  $\log_{10}$  caloric densities and  $\log_{10}$ THg  
409 concentrations are reported in Table 3. *AICc* model selection yielded multiple models with a  $\Delta_i$  within 2  
410 of the best model that were considered as essentially as good as the best model. For  $\log_{10}$  % lipid values  
411 these included two models including subsets of the variables season, % C, and % N, although Akaike  
412 weight ( $w_i$ ) and evidence ratio ( $ER_i$ ) measures were better for the model including all three variables. For  
413 caloric density there were three equivalent models that included combinations of the variables: season,  
414 age, condition, fork length, total weight, % C and % N. Akaike weight ( $w_i$ ) and evidence ratio ( $ER_i$ )  
415 measures suggested better support for the models including either age or fork length as variables, rather  
416 than weight. Similarly for  $\log_{10}$  THg, three equivalent models inclusive of subsets of the variables season,  
417 total weight, condition  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  were estimated, with the model including season having the lowest  
418 likelihood of being the best approximating model for describing the  $\log_{10}$  THg concentration data.

## 419 **4. Discussion**

420 Somatic condition and weight declined in post-winter captured fish as hypothesized. Significant  
421 seasonal differences in %N (decline) and  $\delta^{13}\text{C}$  (increase) were also observed, whereas no significant  
422 seasonal variation was evident for the other biological parameters. Significant seasonal variation existed

423 for lipid content and caloric density, but contrary to what was hypothesized, values determined from  
424 Arctic charr collected during the post-winter sampling were significantly higher than those determined  
425 from fish returning from the marine environment in late summer. Percent lipid values and caloric  
426 densities were correlated in the tissue of post-winter sampled Arctic charr, but were not significantly  
427 related in late summer migrants. Additionally, significant relationships between % lipid values and caloric  
428 densities existed between some, but not all of the studied variables of interest. Seasonal variation existed  
429 for THg concentrations and was contrary to what was expected, declining in post-winter sampled Arctic  
430 charr. THg concentrations were significantly related to some, but not all, of the tested variables and were  
431 unrelated to % lipid values and caloric densities, except in late summer sampled Arctic charr where  
432 concentrations significantly declined with increasing lipid content. Models best supporting the data from  
433 the previously stated analyses better explained variation in data when inclusive of multiple parameters.  
434 While no single model was exclusively best, models generally included variables considered of traditional  
435 importance (i.e., fish size, condition and feeding patterns) as well as season.

#### 436 *4.1 Seasonal Variation in Biological Variables and Stable Isotope Values*

437 Condition, driven by decreases in mean weight rather changes in mean length, declined between  
438 seasonal sampling periods, with mean K values of post-winter collected Arctic charr being significantly  
439 less than those of returning marine migrants. Similar significant over-winter reductions in somatic  
440 condition have been reported for Norwegian resident lacustrine Arctic charr despite winter feeding  
441 (Amundsen and Knudsen 2009; Klemetsen et al. 2003b). However, end of winter condition for non-  
442 reproductive individuals captured near Kangiqsualujjuaq, Nunavik, Québec, was reported to consistently  
443 exceed 1.0 (Boivin and Power 1990). Thus, variation in the extent of post-winter declines in condition is  
444 apparent among populations and likely among years.

445 Declines in condition appear linked to reduced seasonal feeding reflected in the seasonal  
446 differences of % N in the muscle tissue, with declines in % N having been reported for resident lacustrine  
447 Arctic charr after prolonged periods of fasting (Power et al. 2009). Reductions in % N have similarly been  
448 demonstrated in other species (Elliott 1976; Guerin-Ancey 1976; Van Weerd et al. 1995) and are

449 associated with continuous losses of nitrogenous compounds via waste products during fasting (Elliott  
450 1976; Hobson et al. 1993; Steele and Daniel 1978). These losses are substantive when compared to  
451 continuously fed fish (Van Weerd et al. 1995). During prolonged periods of fasting, enrichment of  $\delta^{15}\text{N}$   
452 values is anticipated via similar catabolization processes associated with reductions in % N values  
453 (Gannes et al. 1997; Hobson et al. 1993; McCutchan Jr et al. 2003). While here tests of differences in  
454  $\delta^{15}\text{N}$  between periods showed no significant changes, the differences in mean size between periods may  
455 have masked the extent of the overall increase in  $\delta^{15}\text{N}$  occurring during the fasting period. For example,  
456 the significant relationship between fork length and  $\delta^{15}\text{N}$  often reported in the literature (e.g., (Gantner et  
457 al. 2010; van der Velden et al. 2013a) and seen here ( $r^2 = 0.501$ ,  $p < 0.031$ ) yields a  $\delta^{15}\text{N}$  prediction range  
458 (12.51-13.04) for summer fish of a size equivalent to the observed mean size of post-winter fish that does  
459 not include the mean reported  $\delta^{15}\text{N}$  value (13.33) found for post-winter fish. Furthermore, the overall  
460 absolute increase in  $\delta^{15}\text{N}$  between periods (0.27) observed here fell within the 95% confidence limits  
461 (0.26-0.74) of the mean effect size for starvation induced changes in  $\delta^{15}\text{N}$  estimated from a meta-analysis  
462 of fasting studies (Hertz et al. 2015). Overall, the evidence suggests a biologically significant increase in  
463  $\delta^{15}\text{N}$  occurred as a result of fasting in over-wintering fish. Similarly, use of lipid reserves as an over-  
464 winter energy source (e.g., (Dutil 1986; Jørgensen et al. 1997)), typically in the range of -24.91 to -28.51  
465 for salmonids (Aursand et al. 2000), suggests observable declines in  $\delta^{13}\text{C}$  would occur over the winter  
466 period as tissues equilibrated over time with catabolized lipid reserves (e.g., (Herzka and Holt 2000)). The  
467 observed absence of short term feeding by post-winter captured Arctic charr, established here through an  
468 analysis of the upper gastrointestinal tract, in conjunction with noted declines in weight, somatic  
469 condition, %N, and  $\delta^{13}\text{C}$  and the directional shifts in  $\delta^{15}\text{N}$  are all indicative of prolonged periods of  
470 fasting exceeding several months consistent with the over-wintering period (Hesslein et al. 1993; Power  
471 et al. 2009).

#### 472 *4.2 Lipid Content and Caloric Density*

473 Somatic condition, weight loss and % N values indicative of prolonged periods of fasting are  
474 consistent with previous literature detailing the over-winter period for Arctic charr (Amundsen and

475 Knudsen 2009; Jørgensen et al. 1997; Power et al. 2009). Over-winter changes in lipid content and caloric  
476 densities ran contrary to what was hypothesized and were significantly higher in the dorsal muscle tissue  
477 of post-wintering Arctic charr. Elevated post-winter % lipid values and caloric densities might be  
478 indicative of winter feeding, as previously inferred through observations of foraging behaviour (Boivin  
479 1987) and examination of winter movement activities (Mulder et al. 2018b). However, evidence for the  
480 absence of short term feeding provided by empty stomachs and upper gastrointestinal tracts, reductions in  
481 somatic condition, weight, and % N, is more consistent with prolonged periods of non-feeding (Hesslein  
482 et al. 1993; Power et al. 2009). While previous literature (Jobling et al. 1998; Jørgensen et al. 1997) has  
483 reported female anadromous Arctic charr exhibiting significantly greater depletion of lipid stores  
484 compared to their male counterparts during the over-wintering months, sex was not correlated with %  
485 lipid or caloric densities in this study. The lack of a significant relationship between sex and fish  
486 condition indices measures has been observed with other species (Kitts et al. 2004; Usydus et al. 2012),  
487 with differences among studies possibly related to the proportion of spawning ready adults (Henderson  
488 and Tocher 1987; Jobling et al. 1998; Medford and Mackay 1978).

489         Considered together, the evidence of over-winter fasting, increased mortality, reduced weight and  
490 condition, and paradoxical increases in % lipids and caloric densities suggest condition selective mortality  
491 may be operating over the winter period. Condition selective mortality acts through a range of phenotypic  
492 and genotypic variables to remove poorly conditioned individuals from a population with consequences  
493 for evolution and population demographics (Chen and Maklakov 2012; Gagliano et al. 2007; Ronget et al.  
494 2017). Sources of condition dependent mortality often include starvation, thermal stress, predation, failure  
495 to transition between ontogenetic life stages, and the interactions between the multiple factors (Gagliano  
496 et al. 2007; Miranda and Hubbard 1994; Ronget et al. 2017). For a species such as Arctic charr, over-  
497 wintering in low productivity, oligotrophic environments, and given the evidence noted above, starvation  
498 would appear to be the mechanism of most interest.

499         Starvation or reduced feeding is a consistent driver of over-wintering and selective mortality and  
500 has been implicated as the cause of mortality for different life stages of several fish species, including

501 Arctic charr (Biro et al. 2004; Byström et al. 2006; Post and Evans 1989). Fish size often has considerable  
502 influence, as smaller fish with higher metabolic demands and lower lipid and protein reserves relative to  
503 larger conspecifics deplete critical energy reserves at an escalated rate, resulting in more rapid starvation  
504 and higher subsequent mortality (Henderson et al. 1988; Smith and Griffith 1994; Thompson et al. 1991).  
505 For example, studies of condition selective mortality among age-0 walleye Pollock (*Theragra*  
506 *chalcogramma*) showed that while lipid stores and body condition for the test group as a whole were  
507 rapidly reduced by starvation, survivors had significantly higher lipid content than mortalities, with values  
508 often exceeding those of pre-starvation fish (Sogard and Olla 2000). Additionally, Searcy and Sponaugle  
509 (2001) examined mortality as a function of early life history traits (size-at-age and growth rates) at critical  
510 periods in the Bluehead wrasse (*Thalassoma bifasciatum*) and the Slippery dick (*Halichoeres bivittatus*)  
511 and similarly noted that the better conditioned fish survived. The effect of condition selective mortality,  
512 therefore, would be to shift the mean of the trait distribution in ways that would yield no apparent effect  
513 of over-wintering on lipid and body reserves, or as here, apparent improvement as a result of the  
514 systematic removal of poorer conditioned individuals. Further, condition selective mortality should  
515 increase apparent mortality within the population as was also noted here.

516         The life history of Arctic charr further argues for the influence of condition selective mortality.  
517 Arctic charr are a long lived fish species (Johnson 1983; Johnson 1989; Power et al. 2008) and the  
518 prevalence of condition dependent mortality has been theoretically linked with increased life spans (Chen  
519 and Maklakov 2012). Arctic charr are also fall spawning and body lipids in fish generally decrease  
520 coincident with maturation as individuals mobilize lipids into maturing gonads (Henderson and Tocher  
521 1987). In Arctic charr, the overall change in body lipids maybe as high as 30 – 80% after spawning and  
522 over-wintering (Dutil 1986; Jobling 1981; Jørgensen et al. 1997) with post-spawners being much more  
523 depleted when compared to non-reproductive individuals (Dutil 1986). Spawning has also previously  
524 been linked to increased over-winter mortality in Largemouth bass, *Micropterus salmoides* (Post et al.  
525 1998). If condition-dependent mortality was influencing the Deception River population, it would be  
526 expected to increase the mortality of spawned fish. Given the inverse relationship between body lipid

527 content and the gonadosomatic index (Henderson and Tocher 1987), fish preparing to spawn would have  
528 lower muscle tissue reserves in comparison to non-spawning fish. Higher mortality among spawning fish  
529 with lower muscle lipid reserves would, therefore, effectively act to remove a greater proportion of the  
530 returning migrants at the lower end of % lipid spectrum. The net effect would be an increase in measured  
531 mean % lipids among over-winter survivors, yielding the pattern of seasonal differences in % lipid and  
532 caloric density values observed here. While we did not measure the proportion of spawning ready adults,  
533 the seasonal effect can be expected to be particularly strong in years when there is a high proportion of  
534 spawning ready fish among the marine re-entry migrants.

#### 535 *4.3 THg*

536 Quantification of seasonal variation in THg concentrations determined all measured values were  
537 below the regulatory  $0.5 \text{ mg} \cdot \text{kg}^{-1}$  of mercury recommended in Health Canada's commercial guideline of  
538 maximum levels for chemical contaminants in foods (Government of Canada 2019; Health Canada 2018).  
539 Thus the seasonal risks for capture and consumption associated with the winter fishery for anadromous  
540 Arctic charr in the Deception River system appear minimal. However, further evaluation of seasonal THg  
541 consumption risks for the system's resident Arctic charr is recommended given previously documented  
542 higher winter THg concentrations in other lacustrine resident Arctic charr populations (Kahilainen et al.  
543 2016) and consistently higher THg concentrations in lake-dwelling life history forms for this species  
544 (Swanson et al. 2011; Van der Velden et al. 2015; van der Velden et al. 2013b).

545 Hypothesized over-wintering anorexia, prompted by a cessation of feeding by anadromous Arctic  
546 charr during the winter months, leading to seasonal energy reserve losses and increases in THg  
547 concentrations (Cizdziel et al. 2002; Kahilainen et al. 2016; Keva et al. 2017) was not observed in the  
548 Deception fish. Rather, THg concentrations were significantly higher in tissue collected from Arctic charr  
549 sampled during the summer in the marine environment. Similar to seasonal % lipid and caloric density  
550 data, seasonal differences in THg concentrations may have resulted from the consequences of condition  
551 selective mortality, as suggested by estimated seasonal increases in mortality. The removal of poorly  
552 conditioned, lower lipid content individuals via condition selective mortality would increase measured

553 mean lipid content, which itself is often negatively correlated with THg concentrations (Post et al. 2007;  
554 Wiener et al. 2003). Thus, better conditioned post winter survivors with higher lipids would shift the  
555 distribution of sampled THg concentrations, yielding a decrease in mean THg.

556 Differences in prey THg content in the seasonally occupied habitats and elimination of tissue  
557 THg concentrations, can be reasonably excluded as potential drivers of the observed results. The absence  
558 of short term prey intake, reduced somatic condition, weight, and % N point to a significant period of  
559 fasting. Further, as elimination is a metabolically-dependent process correlated positively with  
560 temperature (Trudel and Rasmussen 1997) it is unlikely to have played a major role in determining  
561 seasonal differences because of the narrow range of cold (0.5-2°C) temperatures occupied by over-  
562 wintering Arctic charr to reduce metabolism (Mulder et al. 2018a). The continued loss of proteins during  
563 fasting via nitrogenous waste products (Elliott 1976; Hobson et al. 1993; Steele and Daniel 1978), e.g.,  
564 reduced % N in post-winter fish, may have contributed to the declines in measured THg as THg in fish  
565 tissues is almost exclusively bound within proteins (Amlund et al. 2007; Bury et al. 2003; Mason et al.  
566 1995). The increase in caloric densities, which includes consideration of protein content, in post-winter  
567 Arctic charr, however, suggests that protein loss during the over-wintering period did not substantially  
568 influence seasonal variation in THg concentrations in Deception River anadromous Arctic charr. The  
569 elimination and depuration of THg is remarkably slow (Amlund et al. 2007; Oliveira Ribeiro et al. 1999),  
570 implying that physiological processes affecting protein content were not the main driver of the  
571 significantly reduced THg concentrations observed in post-winter samples.

#### 572 *4.4 General Linear Models*

573 While there were no clear best models for explaining the % lipid values, caloric densities, and  
574 THg concentrations data, all dependent variable models were better supported when containing multiple  
575 variables. The % lipids models all depended on season and % N, with the best model also including  
576 consideration of % C. Key for % lipids then are variables associated with controlling food intake (season)  
577 and/or the metabolic consequences of fasting (% N), with season and metabolism known to hold  
578 implications for body energy reserves in fish (Dutil 1986; Jørgensen et al. 1997; Thompson et al. 1991)



579 and changes in % N being linked to continuous, non-replaced N losses via excretion during periods of  
580 fasting (Post et al. 2007; Power et al. 2009). In contrast to % lipids, variation in caloric densities were  
581 linked to a suite of biological and feeding variables, predominantly season, and variables reflective of  
582 feeding status. Results here parallel previous research detailing seasonal feeding and its implications for  
583 somatic condition and body reserve depletion in Arctic charr (Amundsen and Knudsen 2009; Rikardsen et  
584 al. 2003; Sæther et al. 1996). Additionally, the inclusion of variables that have associations with both size  
585 and reduced feeding were consistent with the hypothesis that condition selective mortality may be  
586 exerting an influence on this population of Arctic charr, as size, starvation, and reduced feeding are  
587 commonly associated with condition selective mortality (Biro et al. 2004; Byström et al. 2006; Sogard  
588 and Olla 2000). Finally, models for describing variations in THg concentrations accorded with previous  
589 research that has detailed relationships between metal concentrations and weight and condition (Dittman  
590 and Driscoll 2009; Eastwood and Couture 2002; Swanson and Kidd 2010), season (Keva et al. 2017) and  
591 indicators of diet and feeding habits (Goutte et al. 2015; Power et al. 2002; van der Velden et al. 2013a).  
592 Collectively, the multivariate evidence for each of the studied variables suggests that single variable  
593 correlations alone are likely insufficient for understanding variations in the environmental conditions  
594 associated THg concentrations in anadromous Arctic charr.

#### 595 *4.5 Conclusions*

596 In addition to addressing knowledge gaps associated with available regional and population  
597 specific data, this research has increased the limited scientific data available on patterns of metal  
598 contamination in the over-wintering anadromous Arctic charr. To our knowledge, the observed results are  
599 the first detailing higher lipid content and caloric density in Arctic charr captured during the winter  
600 months, a period that has previously been associated with significant reductions in body reserves.  
601 Condition selective mortality is argued to be the most plausible explanation for the observed results,  
602 particularly as concurrently measured variables (e.g., % N, condition, and weight loss) suggested  
603 prolonged periods of fasting coincident with the known seasonal feeding behaviour of Arctic charr.  
604 Relationships between THg and descriptive biological variables were consistent with the seasonally

605 dependent hypothesis, although season itself was not the single best predictor of THg concentrations.  
606 Rather, multi-predictor models better described variation in the data for THg, % lipids and caloric  
607 densities, with variables such as season, somatic condition, age, body size (length or weight) and feeding  
608 tactics ( $\delta^{13}\text{C}$ ,  $\delta^{15}\text{N}$ ) being implicated in varying combinations for the best description of the data sets.  
609 While there is large literature examining the associations between THg and biological or feeding  
610 variables, much less is known about the effect of season. Quantification of seasonal variation in THg  
611 concentrations has determined that winter consumption risks for anadromous Arctic charr in this region  
612 are minimal. However, further work is necessary to elucidate global patterns of seasonal THg  
613 accumulation, especially as seasonal trends may differ between life history forms for this species. Overall,  
614 our corroboration of earlier studies that highlight the importance of season suggests further work is  
615 required to understand seasonality in Arctic charr, especially in the face of climate change and the likely  
616 effects it will have on Arctic aquatic environments.

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627 **References**

628  
629 Amlund H, Lundebye A-K, Berntssen MH (2007) Accumulation and elimination of methylmercury in  
630 Atlantic cod (*Gadus morhua* L.) following dietary exposure. *Aquat Toxicol* 83:323-330  
631 Amundsen P-A, Knudsen R (2009) Winter ecology of Arctic charr (*Salvelinus alpinus*) and Brown trout  
632 (*Salmo trutta*) in a subarctic lake, Norway. *Aquat Ecol* 43:765-775 doi:10.1007/s10452-009-  
633 9261-8  
634 Aursand M, Mabon F, Martin G (2000) Characterization of farmed and wild salmon (*Salmo salar*) by a  
635 combined use of compositional and isotopic analyses. *Journal of the American Oil Chemists'*  
636 *Society* 77:659-666  
637 Berg OK, Berg M (1989) Sea growth and time of migration of anadromous Arctic char (*Salvelinus*  
638 *alpinus*) from the Vardnes River, in Northern Norway. *Can J Fish Aquat Sci* 46:955-960 doi:DOI  
639 10.1139/f89-123  
640 Biro PA, Morton AE, Post JR, Parkinson EA (2004) Over-winter lipid depletion and mortality of age-0  
641 Rainbow trout (*Oncorhynchus mykiss*). *Can J Fish Aquat Sci* 61:1513-1519 doi:10.1139/F04-083  
642 Boivin TG (1987) The winter fishery of Kangiqsualujuaq, Quebec, and winter physiology of Arctic char.  
643 University of Waterloo  
644 Boivin TG, Power G (1990) Winter condition and proximate composition of anadromous Arctic Charr  
645 (*Salvelinus alpinus*) in Eastern Ungava Bay, Québec. *Can J Zool* 68:2284-2289 doi:DOI  
646 10.1139/z90-319  
647 Burnham KP, Anderson DR (2003) Model selection and multimodel inference: a practical information-  
648 theoretic approach. 2<sup>nd</sup> edn. Springer-Verlag, New York, NY, USA  
649 Bury NR, Walker PA, Glover CN (2003) Nutritive metal uptake in teleost fish. *J Exp Biol* 206:11-23  
650 doi:10.1242/jeb.00068  
651 Byström P, Andersson J, Kiessling A, Eriksson LO (2006) Size and temperature dependent foraging  
652 capacities and metabolism: consequences for winter starvation mortality in fish. *Oikos* 115:43-52  
653 doi:10.1111/j.2006.0030-1299.15014.x  
654 Canadian Ice Service (2018) Ice archive – daily ice charts – black and white (1999 – 2018) for the  
655 Hudson Bay region. Government of Canada.  
656 [http://climate.weather.gc.ca/climate\\_data/daily\\_data\\_e.html?StationID=52378&Year=2016&Month=8#](http://climate.weather.gc.ca/climate_data/daily_data_e.html?StationID=52378&Year=2016&Month=8#). Accessed December 8 2017  
657  
658 Chen H-y, Maklakov AA (2012) Longer life span evolves under high rates of condition-dependent  
659 mortality. *Curr Biol* 22:2140-2143 doi:10.1016/j.cub.2012.09.021  
660 Chilton DE, Beamish RJ (1982) Age determination methods for fishes studied by the groundfish program  
661 at the Pacific Biological Station. *Can J Fish Aquat Sci* 60:1-102  
662 Cizdziel J, Hinners T, Cross C, Pollard J (2003) Distribution of mercury in the tissues of five species of  
663 freshwater fish from Lake Mead, USA. *J Environ Monitor* 5:802-807 doi:10.1039/b307641p  
664 Cizdziel J, Hinners T, Pollard J, Heithmar E, Cross C (2002) Mercury concentrations in fish from Lake  
665 Mead, USA, related to fish size, condition, trophic level, location, and consumption risk. *Arch*  
666 *Environ Con Tox* 43:309-317 doi:10.1007/s00244-002-1191-6  
667 Cook RD (1977) Detection of influential observation in linear regression. *Technometrics* 19:15-18  
668 doi:Doi 10.2307/1268249  
669 Craig H (1957) Isotopic standards for carbon and oxygen and correction factors for mass spectrometric  
670 analysis of carbon dioxide. *Geochim Cosmochim Ac* 12:133-149  
671 Dempson JB, Shears M, Bloom M (2002) Spatial and temporal variability in the diet of anadromous  
672 Arctic charr, *Salvelinus alpinus*, in northern Labrador. In: *Ecology, behaviour and conservation*  
673 *of the charrs, genus Salvelinus*. Springer, Dordrecht, Netherlands pp 49-62  
674 Dittman JA, Driscoll CT (2009) Factors influencing changes in mercury concentrations in lake water and  
675 yellow perch (*Perca flavescens*) in Adirondack lakes. *Biogeochemistry* 93:179-196  
676 doi:10.1007/s10533-009-9289-9

677 Doucett RR, Hooper W, Power G (1999) Identification of anadromous and nonanadromous adult Brook  
678 trout and their progeny in the Tabusintac River, New Brunswick, by means of multiple stable  
679 isotope analysis. *T Am Fish Soc* 128:278-288 doi:Doi 10.1577/1548-  
680 8659(1999)128<0278:Ioana>2.0.Co;2

681 Dutil J (1986) Energetic constraints and spawning interval in the anadromous Arctic charr (*Salvelinus*  
682 *alpinus*). *Copeia* 1986:945-955 doi:Doi 10.2307/1445291

683 Eagles-Smith CA et al. (2016) Mercury in western North America: A synthesis of environmental  
684 contamination, fluxes, bioaccumulation, and risk to fish and wildlife. *Sci Total Environ*  
685 568:1213-1226

686 Eastwood S, Couture P (2002) Seasonal variations in condition and liver metal concentrations of yellow  
687 perch (*Perca flavescens*) from a metal-contaminated environment. *Aquat Toxicol* 58:43-56  
688 doi:10.1016/s0166-445x(01)00218-1

689 Eikenberry BCS, Riva-Murray K, Knightes CD, Journey CA, Chasar LC, Brigham ME, Bradley PM  
690 (2015) Optimizing fish sampling for fish-mercury bioaccumulation factors. *Chemosphere*  
691 135:467-473 doi:10.1016/j.chemosphere.2014.12.068

692 Eisler R (1987) Mercury hazards to fish, wildlife, and invertebrates: a synoptic review. Laurel, MD, USA

693 Elliott J (1976) Energy losses in the waste products of Brown trout (*Salmo trutta* L.). *J Anim Ecol*:561-  
694 580

695 Eloranta AP, Mariash HL, Rautio M, Power M (2013) Lipid-rich zooplankton subsidise the winter diet of  
696 benthivorous Arctic charr (*Salvelinus alpinus*) in a subarctic lake. *Freshwater Biol* 58:2541-2554  
697 doi:10.1111/fwb.12231

698 Environment Canada (2018a) Daily data report for August 2016, Salluit, Québec. Government of Canada.  
699 [http://climate.weather.gc.ca/climate\\_data/daily\\_data\\_e.html?StationID=52378&Year=2016&Month=8#](http://climate.weather.gc.ca/climate_data/daily_data_e.html?StationID=52378&Year=2016&Month=8#). Accessed February 8 2019

700 Environment Canada (2018b) Daily data report for February 2016, Salluit, Québec. Government of  
701 Canada.  
702 [http://climate.weather.gc.ca/climate\\_data/daily\\_data\\_e.html?StationID=52378&timeframe=2&StartYear=1840&EndYear=2017&Day=29&Year=2016&Month=2#](http://climate.weather.gc.ca/climate_data/daily_data_e.html?StationID=52378&timeframe=2&StartYear=1840&EndYear=2017&Day=29&Year=2016&Month=2#). Accessed February 8 2019

703 Finstad B, Heggberget T (1993) Migration, growth and survival of wild and hatchery-reared anadromous  
704 Arctic charr (*Salvelinus alpinus*) in Finnmark, Northern Norway. *J Fish Biol* 43:303-312 doi:DOI  
705 10.1111/j.1095-8649.1993.tb00430.x

706 Folch J (1957) A simple method for the isolation and purification of total lipids from animal tissues. *J*  
707 *Biol Chem* 226:497-509

708 Gagliano M, McCormick MI, Meekan MG (2007) Survival against the odds: ontogenetic changes in  
709 selective pressure mediate growth-mortality trade-offs in a marine fish. *Proc R Soc Lond B Biol*  
710 *Sci* 274:1575-1582 doi:10.1098/rspb.2007.0242

711 Gannes LZ, O'Brien DM, Del Rio CM (1997) Stable isotopes in animal ecology: assumptions, caveats,  
712 and a call for more laboratory experiments. *Ecology* 78:1271-1276

713 Gantner K et al. (2010) Mercury concentrations in landlocked Arctic char (*Salvelinus alpinus*) from the  
714 Canadian Arctic. Part II: influence of lake biotic and abiotic characteristics on geographic trends  
715 in 27 populations. *Environ Toxicol Chem* 29:633-643 doi:10.1002/etc.96

716 Glover DC, DeVries DR, Wright RA, Davis DA (2010) Sample preparation techniques for determination  
717 of fish energy density via bomb calorimetry: An evaluation using largemouth bass. *T Am Fish*  
718 *Soc* 139:671-675 doi:10.1577/T09-110.1

719 Goldsmit J, Howland KL, Archambault P (2014) Establishing a baseline for early detection of non-  
720 indigenous species in ports of the Canadian Arctic. *Aquat Invasions* 9:327-342  
721 doi:10.3391/ai.2014.9.3.08

722 Goutte A, Cherel Y, Churlaud C, Ponthus J-P, Massé G, Bustamante P (2015) Trace elements in Antarctic  
723 fish species and the influence of foraging habitats and dietary habits on mercury levels. *Sci Total*  
724 *Environ* 538:743-749 doi:10.1016/j.scitotenv.2015.08.103

727 Government of Canada (2019) Mercury in fish - questions and answers. Government of Canada.  
728 Accessed November 27th 2019

729 Grønvik S, Klemetsen A (1987) Marine food and diet overlap of co-occurring Arctic charr *Salvelinus*  
730 *alpinus* (L.), Brown trout *Salmo trutta* L. and Atlantic salmon *S. salar* L. off Senja, N. Norway.  
731 Polar Biol 7:173-177

732 Gross MR, Coleman RM, McDowall RM (1988) Aquatic productivity and the evolution of diadromous  
733 fish migration. Science 239:1291-1293

734 Guerin-Ancey O (1976) Etude expérimentale de l'excrétion azotée du bar (*Dicentrarchus labrax*) en cours  
735 de croissance IV. Effets de la manipulation et du MS222 Sandoz sur l'excrétion d'ammoniac et  
736 d'urée. Aquac Res 9:367-372

737 Hall B, Bodaly R, Fudge R, Rudd J, Rosenberg D (1997) Food as the dominant pathway of  
738 methylmercury uptake by fish. Water Air Soil Poll 100:13-24

739 Health Canada (2018) Health Canada's maximum levels for chemical contaminants in foods. Government  
740 of Canada Accessed February 1st 2018

741 Henderson P, Holmes R, Bamber RN (1988) Size-selective overwintering mortality in the Sand smelt,  
742 *Atherina boyeri risso*, and its role in population regulation. J Fish Biol 33:221-233 doi:DOI  
743 10.1111/j.1095-8649.1988.tb05465.x

744 Henderson RJ, Tocher DR (1987) The lipid composition and biochemistry of freshwater fish. Prog Lipid  
745 Res 26:281-347 doi:10.1016/0163-7827(87)90002-6

746 Hertz E, Trudel M, Cox MK, Mazumder A (2015) Effects of fasting and nutritional restriction on the  
747 isotopic ratios of nitrogen and carbon: a meta-analysis. Ecology and evolution 5:4829-4839

748 Herzka SZ, Holt GJ (2000) Changes in isotopic composition of red drum (*Sciaenops ocellatus*) larvae in  
749 response to dietary shifts: potential applications to settlement studies. Can J Fish Aquat Sci  
750 57:137-147

751 Hesslein RH, Hallard K, Ramlal P (1993) Replacement of sulfur, carbon, and nitrogen in tissue of  
752 growing Broad whitefish (*Coregonus nasus*) in response to a change in diet traced by  $\delta^{34}\text{S}$ ,  $\delta^{13}\text{C}$ ,  
753 and  $\delta^{15}\text{N}$ . Can J Fish Aquat Sci 50:2071-2076

754 Hobson KA, Alisauskas RT, Clark RG (1993) Stable-nitrogen isotope enrichment in avian tissues due to  
755 fasting and nutritional stress - Implications for isotopic analyses of diet. Condor 95:388-394  
756 doi:Doi 10.2307/1369361

757 Iverson SJ, Lang SL, Cooper MH (2001) Comparison of the Bligh and Dyer and Folch methods for total  
758 lipid determination in a broad range of marine tissue. Lipids 36:1283-1287 doi:10.1007/s11745-  
759 001-0843-0

760 Jardine TD, Hunt RJ, Pusey BJ, Bunn SE (2011) A non-lethal sampling method for stable carbon and  
761 nitrogen isotope studies of tropical fishes. Mar Freshw Res 62:83-90 doi:10.1071/Mf10211

762 Jensen A (1996) Beverton and Holt life history invariants result from optimal trade-off of reproduction  
763 and survival. Can J Fish Aquat Sci 53:820-822

764 Jobling M (1981) The influences of feeding on the metabolic rate of fishes - a short review. J Fish Biol  
765 18:385-400 doi:DOI 10.1111/j.1095-8649.1981.tb03780.x

766 Jobling M, Johansen S, Foshaug H, Burkow I, Jørgensen E (1998) Lipid dynamics in anadromous Arctic  
767 charr, *Salvelinus alpinus* (L.): seasonal variations in lipid storage depots and lipid class  
768 composition. Fish Physiol Biochem 18:225-240

769 Johnson L (1980) The arctic charr, *Salvelinus alpinus*. In: Balon EK (ed) Charrs, salmonid fishes of the  
770 genus *Salvelinus*. Dr W. Junk Publishers, The Hague, Netherlands, pp 15-98

771 Johnson L (1983) Homeostatic characteristics of single species fish stocks in Arctic lakes. Can J Fish  
772 Aquat Sci 40:987-1024 doi:DOI 10.1139/f83-125

773 Johnson L (1989) The anadromous arctic charr, *Salvelinus alpinus*, of Nauyuk Lake, NWT, Canada.  
774 Physiol Ecol Jpn 1:201-227

775 Jonsson B, Jonsson N (1993) Partial migration: niche shift versus sexual maturation in fishes. Rev Fish  
776 Biol Fisher 3:348-365

777 Jonsson B et al. (1988) Life history variation of polymorphic Arctic charr (*Salvelinus alpinus*) in  
778 Thingvallavatn, Iceland. Can J Fish Aquat Sci 45:1537-1547 doi:DOI 10.1139/f88-182

779 Jørgensen E, Johansen S, Jobling M (1997) Seasonal patterns of growth, lipid deposition and lipid  
780 depletion in anadromous Arctic charr. J Fish Biol 51:312-326

781 Jørgensen EH, Foshaug H, Andersson P, Burkow IC, Jobling M (2002) Polychlorinated biphenyl  
782 toxicokinetics and P4501A responses in anadromous Arctic charr during winter emaciation.  
783 Environ Toxicol Chem 21:1745-1752

784 Jørgensen EH, Johnsen HK (2014) Rhythmic life of the Arctic charr: Adaptations to life at the edge. Mar  
785 Genom 14:71-81 doi:<https://doi.org/10.1016/j.margen.2013.10.005>

786 Jørgensen EH, Vijayan MM, Killie J-EA, Aluru N, Aas-Hansen Ø, Maule A (2006) Toxicokinetics and  
787 effects of PCBs in Arctic fish: a review of studies on Arctic charr. J Toxicol Environ Health A  
788 69:37-52

789 Kahilainen K et al. (2016) Seasonal dietary shift to zooplankton influences stable isotope ratios and total  
790 mercury concentrations in Arctic charr (*Salvelinus alpinus* (L.)). Hydrobiologia 783:47-63  
791 doi:10.1007/s10750-016-2685-y

792 Karimi R, Chen CY, Pickhardt PC, Fisher NS, Folt CL (2007) Stoichiometric controls of mercury dilution  
793 by growth. PNAS 104:7477-7482

794 Keva O, Hayden B, Harrod C, Kahilainen K (2017) Total mercury concentrations in liver and muscle of  
795 European whitefish (*Coregonus lavaretus* (L.)) in a subarctic lake - Assessing the factors driving  
796 year-round variation. Environ Pollut 231:1518-1528 doi:10.1016/j.envpol.2017.09.012

797 Kidd K, Hesslein R, Fudge R, Hallard K (1995) The influence of trophic level as measured by  $\delta^{15}\text{N}$  on  
798 mercury concentrations in freshwater organisms. In: Mercury as a global pollutant. Springer,  
799 Dordrecht, Netherlands, pp 1011-1015

800 Kitts DD, Huynh MD, Hu C, Trites AW (2004) Season variation in nutrient composition of Alaskan  
801 walleye pollock. Can J Zool 82:1408-1415

802 Klemetsen A, Amundsen PA, Dempson J, Jonsson B, Jonsson N, O'connell M, Mortensen E (2003a)  
803 Atlantic salmon *Salmo salar* L., Brown trout *Salmo trutta* L. and Arctic charr *Salvelinus alpinus*  
804 (L.): a review of aspects of their life histories. Ecol Freshw Fish 12:1-59 doi:10.1034/j.1600-  
805 0633.2003.00010.x

806 Klemetsen A, Knudsen R, Staldivik F, Amundsen P-A (2003b) Habitat, diet and food assimilation of  
807 Arctic charr under the winter ice in two subarctic lakes. J Fish Biol 62:1082-1098  
808 doi:10.1046/j.1095-8649.2003.00101.x

809 Köck G, Triendl M, Hofer R (1996) Seasonal patterns of metal accumulation in Arctic char (*Salvelinus*  
810 *alpinus*) from an oligotrophic Alpine lake related to temperature. Can J Fish Aquat Sci 53:780-  
811 786 doi:10.1139/cjfas-53-4-780

812 Kristoffersen K (1994) The influence of physical watercourse parameters on the degree of anadromy in  
813 different lake populations of Arctic charr (*Salvelinus alpinus* (L.)) in northern Norway. Ecol  
814 Freshw Fish 3:80-91

815 Lemire M, Kwan M, Laouan-Sidi AE, Muckle G, Pirkle C, Ayotte P, Dewailly E (2015) Local country  
816 food sources of methylmercury, selenium and omega-3 fatty acids in Nunavik, Northern Quebec.  
817 Sci Total Environ 509-510:248-259 doi:<https://doi.org/10.1016/j.scitotenv.2014.07.102>

818 Lepak JM et al. (2012) Manipulation of growth to reduce mercury concentrations in sport fish on a whole-  
819 system scale. Can J Fish Aquat Sci 69:122-135

820 Lescord GL, Johnston TA, Branfireun BA, Gunn JM (2018) Percentage of methylmercury in the muscle  
821 tissue of freshwater fish varies with body size and age and among species. Environ Toxicol Chem  
822 37:2682-2691

823 Lockhart W et al. (2005) A history of total mercury in edible muscle of fish from lakes in northern  
824 Canada. Sci Total Environ 351-352:427-463 doi:10.1016/j.scitotenv.2004.11.027

825 Logan JM, Jardine TD, Miller TJ, Bunn SE, Cunjak RA, Lutcavage ME (2008) Lipid corrections in  
826 carbon and nitrogen stable isotope analyses: comparison of chemical extraction and modelling  
827 methods. J Anim Ecol 77:838-846 doi:10.1111/j.1365-2656.2008.01394.x

828 Mariotti A (1983) Atmospheric nitrogen is a reliable standard for natural  $\delta^{15}\text{N}$  abundance measurements.  
829 Nature 303:685

830 Mason R, Reinfelder J, Morel FM (1995) Bioaccumulation of mercury and methylmercury. Water Air  
831 Soil Pollut 80:915-921

832 Mathisen O, Berg M (1968) Growth rates of the char *Salvelinus alpinus* (L.) in the Vardnes River, Troms,  
833 northern Norway. Rep Inst Freshw Res Drottningholm 48:177-186

834 Mazur MM, Beauchamp DA (2003) A comparison of visual prey detection among species of piscivorous  
835 salmonids: effects of light and low turbidities. Environ Biol Fish 67:397-405  
836 doi:10.1023/A:1025807711512

837 McCutchan Jr JH, Lewis Jr WM, Kendall C, McGrath CC (2003) Variation in trophic shift for stable  
838 isotope ratios of carbon, nitrogen, and sulfur. Oikos 102:378-390

839 Medford BA, Mackay WC (1978) Protein and lipid content of gonads, liver, and muscle of Northern Pike  
840 (*Esox lucius*) in relation to gonad growth. J Fish Res Board Can 35:213-219 doi:10.1139/f78-035

841 Mergler D, Anderson HA, Chan LHM, Mahaffey KR, Murray M, Sakamoto M, Stern AH (2007)  
842 Methylmercury exposure and health effects in humans: a worldwide concern. Ambio 36:3-11

843 Miranda LE, Hubbard WD (1994) Length-dependent winter survival and lipid composition of age-0  
844 Largemouth bass in Bay Springs reservoir, Mississippi. T Am Fish Soc 123:80-87  
845 doi:10.1577/1548-8659(1994)123<0080:LDWSAL>2.3.CO;2

846 Moore J-S, Harris LN, Kessel ST, Bernatchez L, Tallman RF, Fisk AT (2016) Preference for nearshore  
847 and estuarine habitats in anadromous Arctic char (*Salvelinus alpinus*) from the Canadian high  
848 Arctic (Victoria Island, Nunavut) revealed by acoustic telemetry. Can J Fish Aquat Sci 73:1434-  
849 1445

850 Moore JW, Moore IA (1974) Food and growth of Arctic char, *Salvelinus alpinus* (L.), in the Cumberland  
851 Sound area of Baffin Island. J Fish Biol 6:79-92 doi:10.1111/j.1095-8649.1974.tb04525.x

852 Mulder I, Dempson J, Fleming I, Power M (2019) Diel activity patterns in overwintering Labrador  
853 anadromous Arctic charr. Hydrobiologia 840:89-102

854 Mulder IM, Morris CJ, Dempson JB, Fleming IA, Power M (2018a) Overwinter thermal habitat use in  
855 lakes by anadromous Arctic charr. Can J Fish Aquat Sci 75:2343-2353

856 Mulder IM, Morris CJ, Dempson JB, Fleming IA, Power M (2018b) Winter movement activity patterns  
857 of anadromous Arctic charr in two Labrador lakes. Ecol Freshw Fish 27:785-797

858 Murdoch A, Dempson JB, Martin F, Power M (2015) Temperature–growth patterns of individually  
859 tagged anadromous Arctic charr *Salvelinus alpinus* in Ungava and Labrador, Canada. Ecol  
860 Freshw Fish 24:193-203 doi:10.1111/eff.12133

861 Murdoch A, Klein G, Doidge DW, Power M (2013) Assessing the food web impacts of an anadromous  
862 Arctic charr introduction to a sub-Arctic watershed using stable isotopes. Fish Manage Ecol  
863 20:302-314 doi:10.1111/fme.12012

864 Murdoch A, Power M (2013) The effect of lake morphometry on thermal habitat use and growth in Arctic  
865 charr populations: implications for understanding climate-change impacts. Ecol Freshw Fish  
866 22:453-466

867 Nordeng H (1983) Solution to the " char problem" based on Arctic char (*Salvelinus alpinus*) in Norway.  
868 Can J Fish Aquat Sci 40:1372-1387

869 Oliveira Ribeiro CA, Rouleau C, Pelletier E, Audet C, Tjälve H (1999) Distribution kinetics of dietary  
870 methylmercury in the arctic charr (*Salvelinus alpinus*). Environ Sci Technol 33:902-907

871 Olk TR, Karlsson T, Lydersen E, Økelsrud A (2016) Seasonal variations in the use of profundal habitat  
872 among freshwater fishes in Lake Norsjø, Southern Norway, and subsequent effects on fish  
873 mercury concentrations. Environments 3:29

874 Post D, Layman C, Arrington D, Takimoto G, Quattrochi J, Montaña C (2007) Getting to the fat of the  
875 matter: Models, methods and assumptions for dealing with lipids in stable isotope analyses.  
876 Oecologia 152:179-189 doi:10.1007/s00442-006-0630-x



877 Post DM, Kitchell JF, Hodgson JR (1998) Interactions among adult demography, spawning date, growth  
878 rate, predation, overwinter mortality, and the recruitment of Largemouth bass in a northern lake.  
879 Can J Fish Aquat Sci 55:2588-2600 doi:10.1139/f98-139

880 Post JR, Evans DO (1989) Size-dependent overwinter mortality of young-of-the-year Yellow perch  
881 (*Perca flavescens*): laboratory, in situ enclosure, and field experiments. Can J Fish Aquat Sci  
882 46:1958-1968 doi:10.1139/f89-246

883 Power M, Klein G, Guiguer K, Kwan M (2002) Mercury accumulation in the fish community of a sub-  
884 Arctic lake in relation to trophic position and carbon sources. J Appl Ecol 39:819-830

885 Power M, Power G, Reist JD, Bajno R (2009) Ecological and genetic differentiation among the Arctic  
886 charr of Lake Aigueau, Northern Québec. Ecol Freshw Fish 18:445-460 doi:10.1111/j.1600-  
887 0633.2009.00362.x

888 Power M, Reist JD, Dempson JB (2008) Fish in high-latitude Arctic lakes. In: Polar lakes and rivers. pp  
889 249-267

890 Ricker WE (1975) Computation and interpretation of biological statistics of fish populations. B Fish Res  
891 Board Can 191:1-382

892 Rigét F, Asmund G, Aastrup P (2000) Mercury in Arctic char (*Salvelinus alpinus*) populations from  
893 Greenland. Sci Total Environ 245:161-172 doi:[https://doi.org/10.1016/S0048-9697\(99\)00441-6](https://doi.org/10.1016/S0048-9697(99)00441-6)

894 Rikardsen A, Amundsen P-A, Bodin P (2003) Growth and diet of anadromous Arctic charr after their  
895 return to freshwater. Ecol Freshw Fish 12:74-80 doi:10.1034/j.1600-0633.2003.00001.x

896 Ronget V, Garratt M, Lemaître J-F, Gaillard J-M (2017) The "evo-demo" implications of condition-  
897 dependent mortality. Trends Ecol Evol 32:909-921

898 Sæther B-S, Johnsen HK, Jobling M (1996) Seasonal changes in food consumption and growth of Arctic  
899 charr exposed to either simulated natural or a 12:12 LD photoperiod at constant water  
900 temperature. J Fish Biol 48:1113-1122 doi:10.1111/j.1095-8649.1996.tb01808.x

901 Sanderson BL, Tran CD, Coe HJ, Pelekis V, Steel EA, Reichert WL (2009) Nonlethal sampling of fish  
902 caudal fins yields valuable stable isotope data for threatened and endangered fishes. T Am Fish  
903 Soc 138:1166-1177 doi:10.1577/T08-086.1

904 Searcy SP, Sponaugle S (2001) Selective mortality during the larval–juvenile transition in two coral reef  
905 fishes. Ecology 82:2452-2470

906 Shapiro SS, Wilk MB (1965) An analysis of variance test for normality (complete samples). Biometrika  
907 52:591-611

908 Smith R, Griffith J (1994) Survival of Rainbow trout during their first winter in the Henrys Fork of the  
909 Snake River, Idaho. T Am Fish Soc 123:747-756

910 Sogard S, Olla B (2000) Endurance of simulated winter conditions by age-0 Walleye pollock: effects of  
911 body size, water temperature and energy stores. J Fish Biol 56:1-21

912 Steele KW, Daniel RMJ (1978) Fractionation of nitrogen isotopes by animals: Further complication to the  
913 use of variations in the natural abundance of  $\delta^{15}\text{N}$  for tracer studies. J Agric Sci 90:7-9  
914 doi:10.1017/S002185960004853X

915 Svenning M, Smith-Nilsen A, Jobling M (1992) Sea water migration of Arctic charr (*Salvelinus alpinus*  
916 L.)-- correlation between freshwater growth and seaward migration, based on back-calculation  
917 from otoliths. Nordic journal of freshwater research Drottningholm 67:18-26

918 Swanson H, Gantner N, Kidd K, Muir DCG, Reist J (2011) Comparison of mercury concentrations in  
919 landlocked, resident, and sea-run fish (*Salvelinus* spp.) from Nunavut, Canada. Environ Toxicol  
920 Chem 30:1459-1467 doi:10.1002/etc.517

921 Swanson HK, Kidd KA (2010) Mercury concentrations in Arctic food fishes reflect the presence of  
922 anadromous Arctic charr (*Salvelinus alpinus*), species, and life history. Environ Sci Technol  
923 44:3286-3292 doi:10.1021/es100439t

924 Thompson JM, Bergersen EP, Carlson CA, Kaeding LR (1991) Role of size, condition, and lipid content  
925 in the overwinter survival of age-0 Colorado squawfish. T Am Fish Soc 120:346-353

- 926 Todd CD, Hughes SL, Marshall CT, Maclean JC, Lonergan ME, Biuw EM (2008) Detrimental effects of  
 927 recent ocean surface warming on growth condition of Atlantic salmon. *Glob Change Biol* 14:958-  
 928 970 doi:10.1111/j.1365-2486.2007.01522.x
- 929 Tran L, Reist JD, Power M (2015) Total mercury concentrations in anadromous Northern Dolly Varden  
 930 from the northwestern Canadian Arctic: A historical baseline study. *Sci Total Environ* 509-  
 931 510:154-164 doi:<https://doi.org/10.1016/j.scitotenv.2014.04.099>
- 932 Trudel M, Rasmussen JB (1997) Modeling the elimination of mercury by fish. *Environ Sci Technol*  
 933 31:1716-1722
- 934 Turesson H, Brönmark C (2007) Predator–prey encounter rates in freshwater piscivores: effects of prey  
 935 density and water transparency. *Oecologia* 153:281-290 doi:10.1007/s00442-007-0728-9
- 936 United States Environmental Protection Agency (1998) Method 7473: mercury in solids and solutions by  
 937 thermal decomposition, amalgamation, and atomic absorption spectrophotometry, Revision 0 edn.  
 938 United States Environmental Protection Agency, Washington, DC, USA
- 939 Usydus Z, Szlifder-Richert J, Adamczyk M (2012) Variations in proximate composition and fatty acid  
 940 profiles of Baltic sprat (*Sprattus sprattus balticus*). *Food Chem* 130:97-103  
 941 doi:<https://doi.org/10.1016/j.foodchem.2011.07.003>
- 942 Van der Velden S, Dempson J, Power M (2015) Comparing mercury concentrations across a thirty year  
 943 time span in anadromous and non-anadromous Arctic charr from Labrador, Canada. *Sci Total*  
 944 *Environ* 509:165-174
- 945 van der Velden S, Dempson JB, Evans MS, Muir DCG, Power M (2013a) Basal mercury concentrations  
 946 and biomagnification rates in freshwater and marine food webs: Effects on Arctic charr  
 947 (*Salvelinus alpinus*) from eastern Canada. *Sci Total Environ* 444:531-542  
 948 doi:10.1016/j.scitotenv.2012.11.099
- 949 van der Velden S, Evans MS, Dempson JB, Muir DCG, Power M (2013b) Comparative analysis of total  
 950 mercury concentrations in anadromous and non-anadromous Arctic charr (*Salvelinus alpinus*)  
 951 from eastern Canada. *Sci Total Environ* 447:438-449  
 952 doi:<https://doi.org/10.1016/j.scitotenv.2012.12.092>
- 953 Van Weerd JH, Verástegui AM, Tijssen PAT (1995) Nitrogen excretion and determination of nitrogen  
 954 and energy budgets in Rainbow trout (*Oncorhynchus mykiss* R.) under different feeding regimes.  
 955 *J Appl Ichthyol* 11:322-328 doi:10.1111/j.1439-0426.1995.tb00034.x
- 956 Wang X, Wang W-X (2015) Physiologically based pharmacokinetic model for inorganic and  
 957 methylmercury in a marine fish. *Environ Sci Technol* 49:10173-10181
- 958 Wiener JG, Krabbenhoft DP, Heinz GH, Scheuhammer AM (2003) Ecotoxicology of mercury. In:  
 959 Hoffman DJ, A. RB, Jr. BGA, Jr. CJ (eds) *Handbook of ecotoxicology*. 2<sup>nd</sup> edn. CRC Press, Boca  
 960 Raton, FL, USA, pp 409-463
- 961 Zar JH (2007) *Biostatistical analysis*. 5<sup>th</sup> edn. Prentice Hall, Inc., Upper Saddle River, NJ, USA

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## Tables

**Table 1** Means  $\pm$  standard deviations and ranges are given for fork lengths, total weights, ages, somatic condition,  $\delta^{13}\text{C}$ , % carbon (C),  $\delta^{15}\text{N}$ , and % nitrogen (N) of Arctic charr used from the summer 2016 and post-winter 2017 sampling seasons. Sex ratios are also noted. Significant seasonal differences are denoted with \* ( $p < 0.05$ ), \*\* ( $p < 0.001$ ), and \*\*\* ( $p < 0.0001$ )

Season of Capture	Sample Size	Fork length (mm)	Total weight (g)	Age (Years)	Sex	Condition (K)	$\delta^{13}\text{C}$ (‰)	% C	$\delta^{15}\text{N}$ (‰)	% N
Summer 2016	49	$483.5 \pm 107.1$	$1380 \pm 819^*$	$9.6 \pm 2.4$	67% Female	$1.06 \pm 0.28^{***}$	$-19.42 \pm 0.99^{***}$	$47.36 \pm 3.06$	$13.06 \pm 1.11$	$14.00 \pm 0.82^{***}$
		143.0; 689.0	20; 3300	5; 15	33% Male	0.57; 2.38	-22.97; -16.82	35.72; 52.19	8.12; 15.00	11.22; 15.73
Post-Winter 2017	51	$449.4 \pm 130.2$	$1031 \pm 824$	$8.9 \pm 2.9$	73% Female	$0.91 \pm 0.11$	$-20.55 \pm 1.00$	$48.30 \pm 2.58$	$13.33 \pm 1.01$	$12.92 \pm 1.05$
		221.0; 698.5	99.8; 3230	5; 20	27% Male	0.71; 1.46	-22.75; -16.89	44.03; 57.51	11.23; 15.22	8.60; 14.64

**Table 2** Means  $\pm$  standard deviations and ranges of dorsal muscle % lipid values, caloric densities, and THg concentrations of anadromous Arctic charr captured in the summer 2016 and post-winter 2017 sampling periods. \* ( $p < 0.05$ ), \*\* ( $p < 0.001$ ), and \*\*\* ( $p < 0.0001$ ) indicate significant seasonal variation

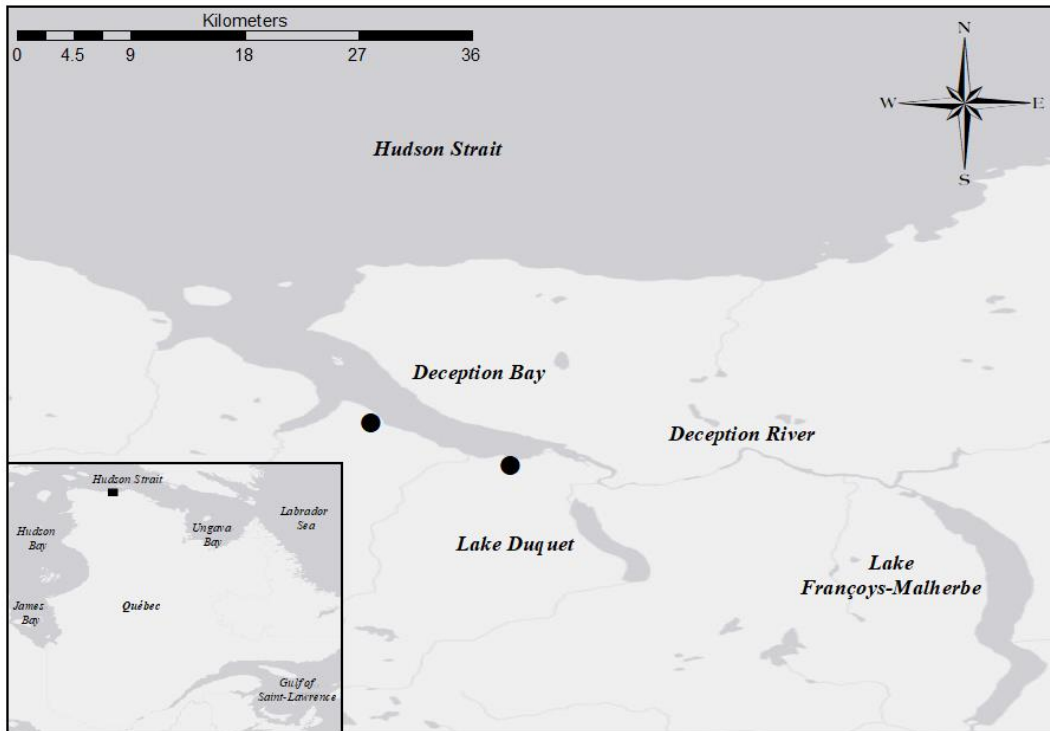
Location of Capture	Sample Size	Lipids (%)	Sample Size	Caloric Density (cal·g <sup>-1</sup> )	Sample Size	THg (mg·kg <sup>-1</sup> )
Summer 2016	40	4.08 $\pm$ 2.00 2.15; 11.76	30	1327.6 $\pm$ 51.6 1180.8; 1409.8	49	0.12 $\pm$ 0.05* 0.06; 0.26
Post-Winter 2017	36	8.34 $\pm$ 4.85*** 2.66; 27.98	27	1545.8 $\pm$ 88.5*** 1371.8; 1723.8	51	0.09 $\pm$ 0.05 0.05; 0.35

**Table 3** Sample size corrected Akaike information criterion (AICc) rankings of the models\* that best described variation in  $\log_{10}$  % lipid values (top)  $\log_{10}$  caloric densities ( $\text{cal} \cdot \text{g}^{-1}$ ) (middle) and  $\log_{10}$  THg concentrations ( $\text{mg} \cdot \text{kg}^{-1}$ ) (bottom) from anadromous Deception River Arctic charr. Only models within  $\Delta_i = 2$  are included in the table as these were considered equivalent to the best model. Also given are the number of model fitted parameters ( $k$ ), residual sums of squares (RSS), delta values ( $\Delta_i$ ) defining model the difference between the estimated model AICc and the best model AICc, the Akaike weight ( $w_i$ ) and evidence ratio ( $ER_i$ ) for each model

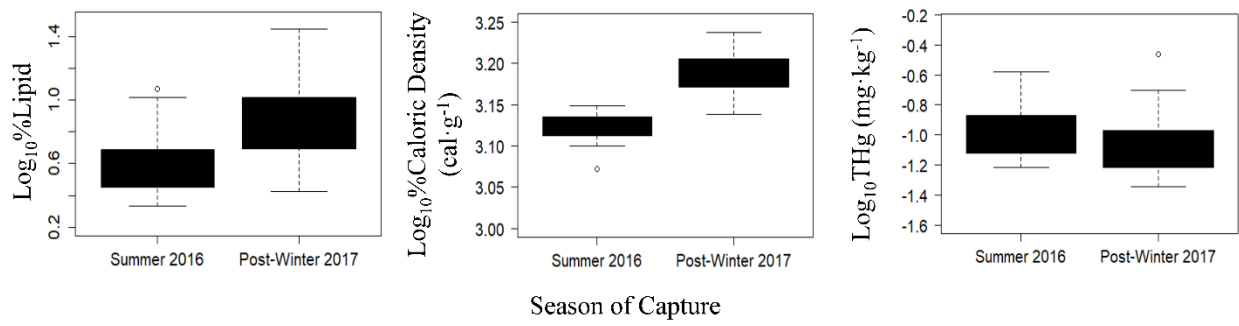
Model	$k$	RSS	AICc	$\Delta_i$	$w_i$	$ER_i$
Variations in $\log_{10}$ % lipids						
Season, % C, % N	5	2.32	-38.78	0.00	0.26	1.00
Season, % N	4	2.42	-37.79	0.99	0.16	1.63
Variations in $\log_{10}$ caloric density						
Season, Age, Condition, % C, % N	7	0.01	-296.32	0.00	0.28	1.00
Season, Fork length, Condition, % C, % N	7	0.01	-296.20	0.12	0.26	1.06
Season, Total weight, Condition, % C, % N	7	0.01	-294.69	1.64	0.12	2.27
Variations in $\log_{10}$ THg						
Total weight, Condition, $\delta^{13}\text{C}$	5	2.40	-78.34	0.00	0.27	1.00
Total weight, Condition, $\delta^{13}\text{C}$ , $\delta^{15}\text{N}$	6	2.37	-77.73	0.61	0.20	1.36
Season, Total weight, Condition, $\delta^{13}\text{C}$	6	2.39	-76.49	1.85	0.11	2.52

\* General linear models were estimated using the following candidate variables: season (summer 2016 and post-winter 2017), fork length (mm), total weight (g), age (years), somatic condition (K),  $\delta^{13}\text{C}$  (‰), % carbon (C),  $\delta^{15}\text{N}$  (‰), % nitrogen (N), and significant two-way interactions. Explanatory variables that were highly correlated, or known to have biological redundancy (e.g., fork length and total weight) were not included in the same model

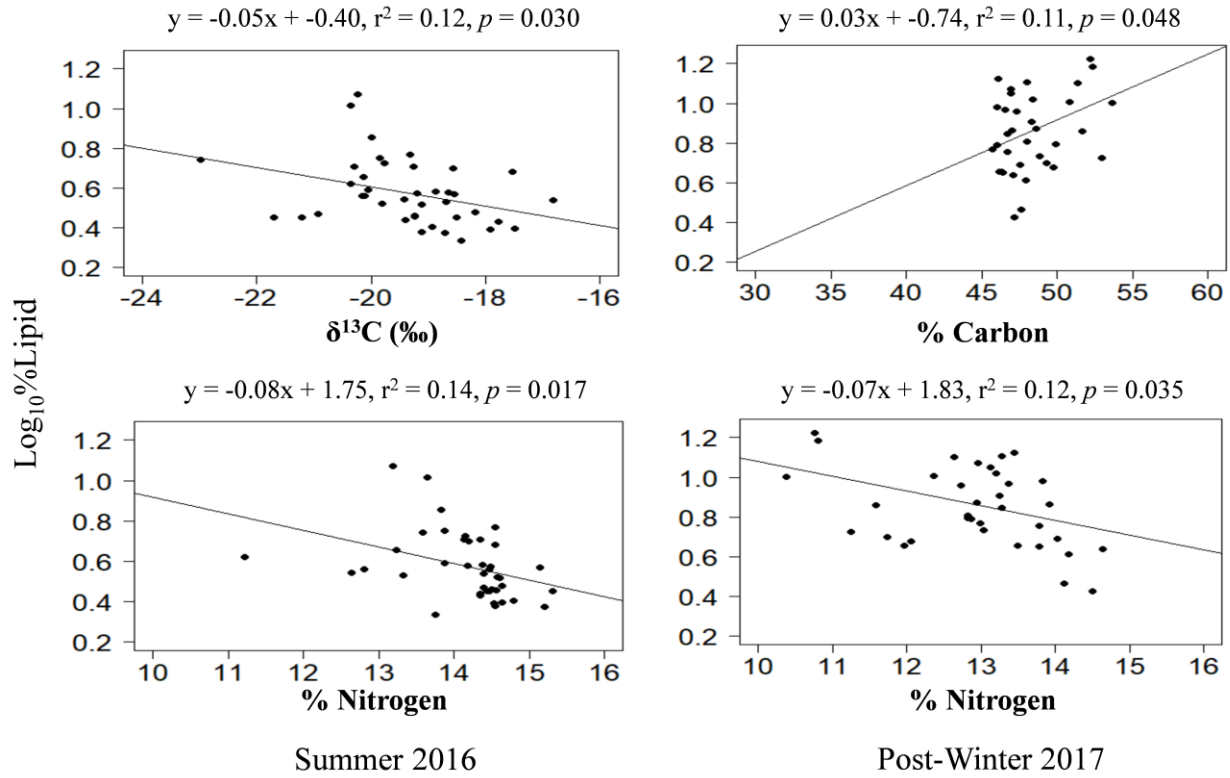
## Figures



**Fig. 1** Map of the Deception Bay, the Deception River, and the two over-wintering lakes, Lake Duquet and Lake François-Malherbe, from which Arctic charr were sampled for this study. Black circles represent mining operations present in the area, while the black square represents the sampling locations in relation to the province of Québec in eastern Canada

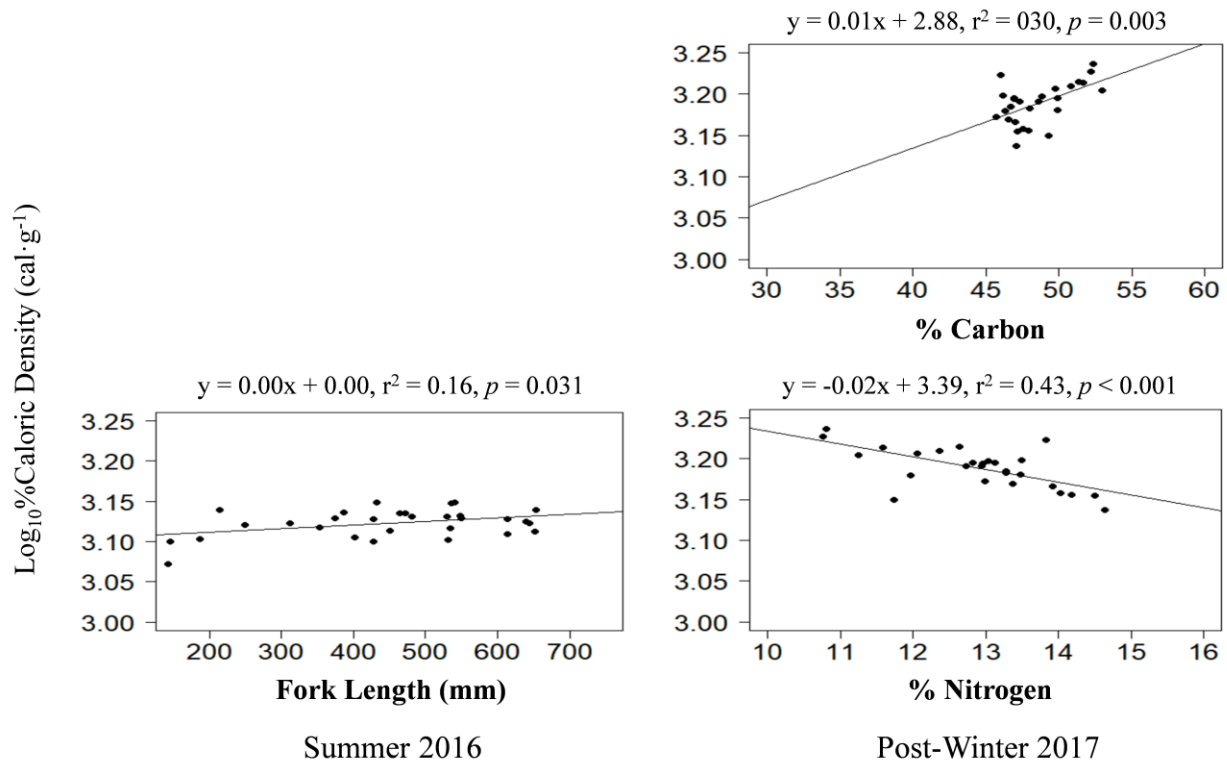


**Fig. 2** Seasonal variation in dorsal muscle log<sub>10</sub> % lipid values, log<sub>10</sub> caloric densities, and log<sub>10</sub> THg concentrations from summer 2016 and post-winter 2017 sampled Deception River anadromous Arctic charr

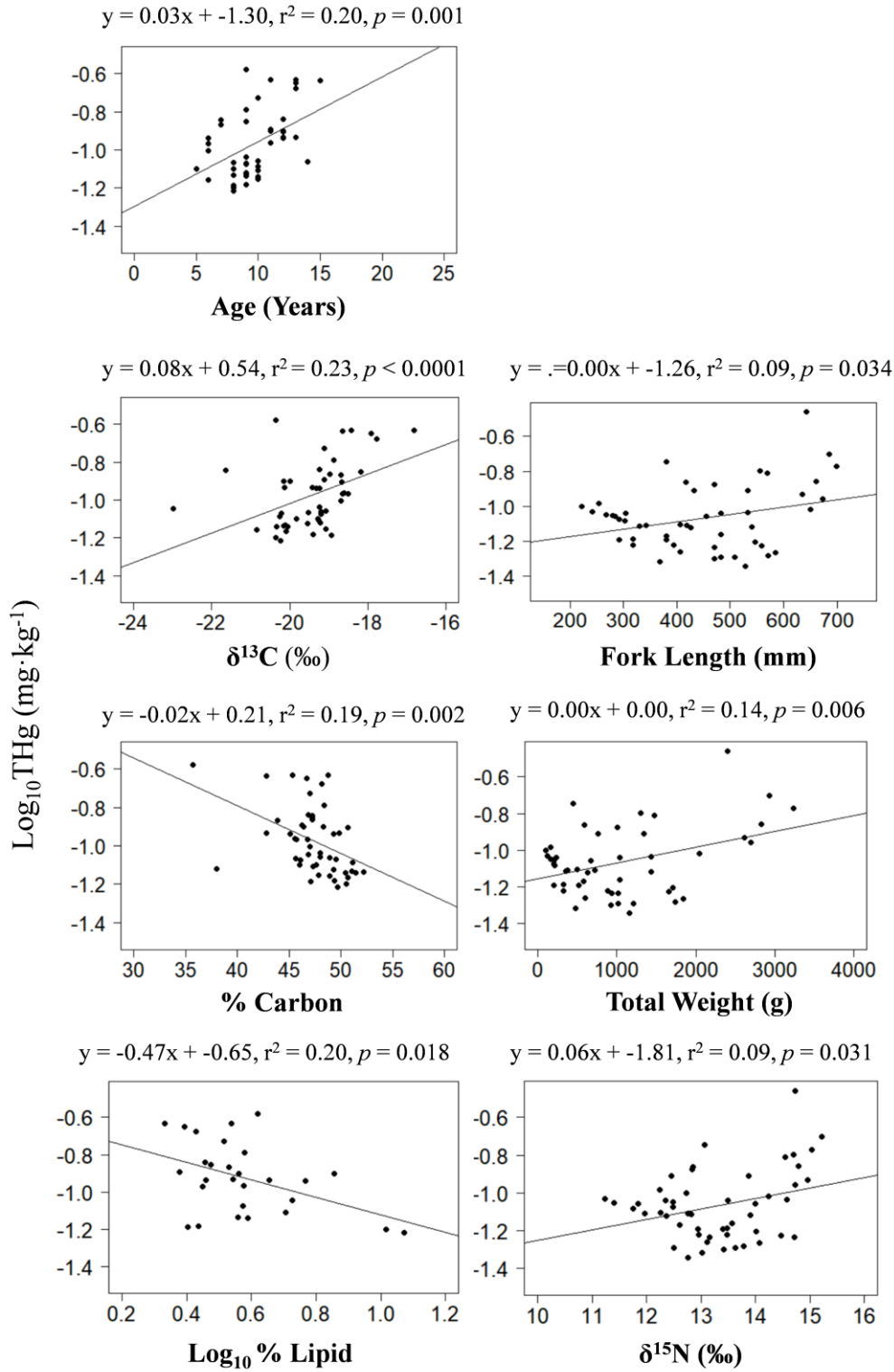


**Fig. 3** Relationships between dorsal muscle  $\text{log}_{10}$  % lipid values and biological variables (fork length, total weight, fish age, and somatic condition), as well as stable isotopes ( $\delta^{13}\text{C}$ , % carbon,  $\delta^{15}\text{N}$ , and % nitrogen)





**Fig. 4** Relationships between collected dorsal muscle  $\text{log}_{10}$  caloric densities and biological variables (fork length, total weight, fish age, and somatic condition), as well as stable isotopes ( $\delta^{13}\text{C}$ , % carbon,  $\delta^{15}\text{N}$ , and % nitrogen)



Summer 2016

Post-Winter 2017

**Fig. 5** Relationships between sampled dorsal muscle  $\log_{10}$  THg concentrations and biological variables (fork length, total weight, age, and somatic condition), stable isotopes ( $\delta^{13}\text{C}$ ,  $\delta^{15}\text{N}$ , and % carbon, and % nitrogen), and fish condition measures ( $\log_{10}$  % lipid values)