

# Journal Pre-proof

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PII: S0013-9351(24)01252-0

DOI: <https://doi.org/10.1016/j.envres.2024.119347>

Reference: YENRS 119347

To appear in: *Environmental Research*

Received Date: 27 February 2024

Revised Date: 30 May 2024

Accepted Date: 4 June 2024

Please cite this article as: Boudreault, J., Lavigne, É., Campagna, C., Chebana, F., Estimating the heat-related mortality and morbidity burden in the province of Quebec, Canada, *Environmental Research*, <https://doi.org/10.1016/j.envres.2024.119347>.

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# 1 **Estimating the heat-related mortality and morbidity** 2 **burden in the province of Quebec, Canada**

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16  
17 Revised manuscript submitted to *Environmental Research*.

18  
19 May 30, 2024

## 20 **Abstract**

21 **Background:** As climate change increases the frequency and intensity of extreme heat  
22 events, there is an urgent need to quantify the heat-related health burden. However, most  
23 past studies have focussed on a single health outcome (mainly mortality) or on specific  
24 heatwaves, thus providing limited knowledge of the total pressure heat exerts on health  
25 services.

26 **Objectives:** This study aims to quantify the heat-related mortality and morbidity burden for  
27 five different health outcomes including all-cause mortality, hospitalizations, emergency  
28 department (ED) visits, ambulance transports and calls to a health hotline, using the  
29 province of Quebec (Canada) as a case study.

30 **Methods:** A two-step statistical analysis was employed to estimate regional heat-health  
31 relationships using Distributed Lag Non-Linear Models (DLNM) and pooled estimates  
32 using a multivariate meta-regression. Heat burden was quantified by attributable fraction  
33 (AF) and attributable number (AN) for two temperature ranges: total heat (above the  
34 minimum mortality/morbidity temperature) and extreme heat (above the 95<sup>th</sup> percentile of  
35 temperature).

36 **Results:** Higher temperatures were associated with greater risk ratios for all the health  
37 outcomes studied, but at different levels. Significant AF ranging from 2–3% for the total  
38 heat effect and 0.4–1.0% for extreme heat effect were found for all health outcomes, except  
39 for hospitalizations that had an AF of 0.1% for both heat exposures. The estimated burden  
40 of all heat (and extreme heat) every summer across the province was 470 (200) deaths, 225

41 (170) hospitalizations, 36 000 (6 200) ED visits, 7 200 (1 500) ambulance transports and  
42 15 000 (3 300) calls to a health hotline, all figures significant.

43 **Discussion:** This new knowledge on the total heat load will help public health authorities  
44 to target appropriate actions to reduce its burden now and in the future. The proposed state-  
45 of-the-art framework can easily be applied to other regions also experiencing the adverse  
46 effects of extreme heat.

47 **Keywords :** hot temperature, extreme heat events, health outcomes, distributed lag non-  
48 linear model, multivariate meta-regression, attributable fraction.

49

## 50 **1. Introduction**

51 It is well established that elevated ambient temperatures (i.e., heat) exacerbate health  
52 conditions, leading to increased mortality and morbidity (Basu, 2009; Campbell et al.,  
53 2018; Kovats & Hajat, 2008; Song et al., 2017). These heat-related health effects can  
54 translate into massive costs borne by the healthcare system and society in general (Callahan  
55 & Mankin, 2022; Wondmagegn et al., 2019). Because of climate change and population  
56 ageing, these impacts will become even more devastating in the future (Curtis et al., 2017;  
57 Huang et al., 2011; 2013). Therefore, there is an urgent need to understand and quantify  
58 the current heat load on health to reduce its effects now and in the future.

59 In the recent years, the quantification of the heat burden became increasingly important  
60 (e.g., Ballester et al., 2023; Burkart et al., 2021; Gasparrini et al., 2015; Vicedo-Cabrera et  
61 al., 2021; Zhao et al., 2021). However, previous studies on temperature-related health  
62 burden have focussed mainly on mortality, as recently reviewed by Cheng et al. (2019).  
63 The same emphasis on mortality has also been observed more generally in the heat-health  
64 literature (Campbell et al., 2018; Cole et al., 2023). Although heat-related mortality is a  
65 major concern, the burden on the healthcare system should also be measured in terms of  
66 morbidity indicators such as hospitalizations, emergency department (ED) visits and  
67 ambulance transports (Wondmagegn et al., 2019), among others. While studies estimating  
68 morbidity-health relationships exist (see Li et al. (2015) and Ye et al. (2012) for reviews),  
69 a more limited number of studies have taken a step further and quantified the morbidity  
70 burden associated with heat (e.g., Bai et al., 2016, 2018; Cheng et al., 2016; Lin et al.,  
71 2012; Liu et al., 2019; Wellenius et al., 2017; Wondmagegn, 2021a, 2021b; and Cheng et

72 al. (2019) for a review). Finally, most of the aforementioned studies were limited to a single  
73 health outcome (i.e., either mortality or one morbidity variable), thus providing a limited  
74 portrait of the overall heat burden.

75 In addition, past research on the heat burden has sometimes focussed on specific episodes  
76 of extreme heat such as “official” heatwaves defined by public health authorities (e.g.,  
77 Adélaïde et al., 2021; Bustinza et al., 2013; Fouillet et al., 2006; Knowlton et al., 2009;  
78 Limaye et al., 2019) or on the additional heatwave effect above average temperatures (e.g.,  
79 Guo et al., 2018; Yin et al., 2018). While these studies are important, they only cover part  
80 of the total heat load. Indeed, heat can have adverse effects at temperatures below heatwave  
81 thresholds, as well as on singular days of hot temperature (i.e., non-consecutive) (Campbell  
82 et al., 2018; Song et al., 2017). Moreover, there exists many different definitions of  
83 heatwaves (e.g., using different temperature thresholds, single or multiple day average,  
84 etc.), in addition to the “official” heatwaves, that can all lead to diverging results when  
85 quantifying the heat burden (e.g., Cheng et al., 2018; Kanti et al., 2022; Kent et al., 2014;  
86 Xu et al., 2016).

87 In this study, a detailed analysis of the total heat burden of 5 health outcomes, including  
88 both mortality and morbidity (i.e., hospitalizations, ED visits, ambulance transports and  
89 calls to a health hotline) was performed, using the province of Quebec (Canada) as a case  
90 study. Heat-health exposure-response functions were developed using Distributed Lag  
91 Non-Linear Models (DLNM) and multivariate meta-regressions. They were then used to  
92 compute attributable fractions and numbers to both total and extreme heat. This study  
93 provides, for one of the first time, a comprehensive portrait of the overall heat burden on

94 various mortality and morbidity indicators using a unified and state-of-the-art  
95 methodology.

## 96 **2. Material and methods**

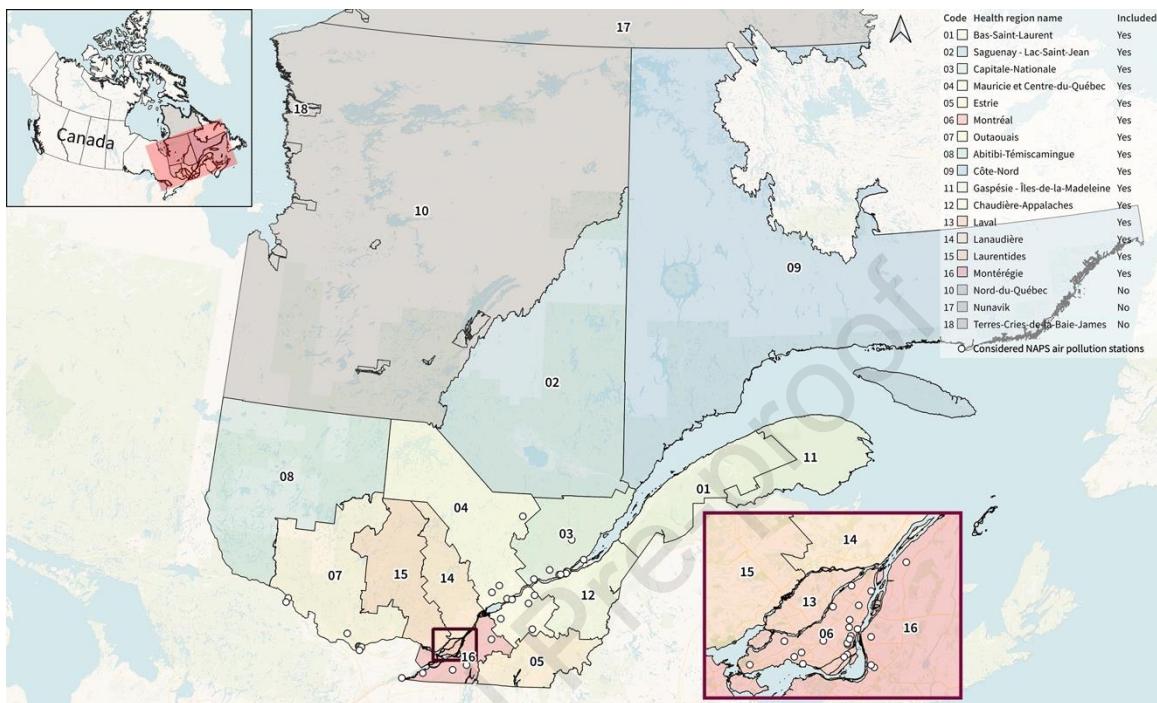
### 97 **2.1. Study design and health outcomes**

98 A time series ecological study in the 15 southernmost health regions (HR) in the province  
99 of Quebec, Canada, was performed (Fig. 1). Quebec has a population of approximately  
100 8.5M inhabitants and is the second most populous province of Canada, with 23% of the  
101 Canadian population. Given the diversity of socioeconomic and climate conditions across  
102 its HR, Quebec is a good candidate for this study. In addition, Quebec's health data was  
103 readily available for multiple health outcomes, which was not systematically the case for  
104 the other provinces of Canada. Finally, most heat burden studies performed in Quebec have  
105 focussed on heatwaves only (e.g., Bustinza et al., 2013; Lebel et al., 2017, 2019). This  
106 project received ethics approval from the Human Research Ethics Committee of the  
107 National Institute of Scientific Research (CER-22-693).

108 The HR spatial resolution was chosen because of its relevance for decision makers (i.e.,  
109 each HR has its own public health team and director). Also, it represents a finer resolution  
110 of previously developed "climate regions" for heat-health studies in Quebec, that are  
111 combinations of multiple HR (e.g., Chebana et al., 2013; Tupinier Martin et al., 2024). As  
112 in previous studies, the northernmost HR #10, #17 and #18 were excluded due to their low  
113 population (i.e., <1% of Quebec population) and their colder climate. Associations between  
114 daily mortality or morbidity and ambient temperature in each HR during the months of  
115 May to September were evaluated.

116  
117  
118

**Fig. 1. Map of the studied health regions (HR) in the province of Quebec (Canada) and location of the considered air pollution stations of the National Air Pollution Surveillance (NAPS) program. HR #10, #17 and #18 were excluded from the present study.**



119

120 Health data was made available by the *Institut national de santé publique du Québec*  
 121 (INSPQ) at the HR level, based on 2021 administrative boundaries for all years of data.  
 122 Studied health outcomes were the daily cases of all-cause: 1) mortality, 2) hospitalizations,  
 123 3) ED visits, 4) ambulance transports and 5) *Info-Santé* 811 calls (a free hotline service for  
 124 non-urgent health issues). All-cause indicators were selected instead of cause-specific (e.g.,  
 125 cardiovascular diseases) as they are more readily available indicators tracked by public  
 126 health authorities in Québec through the health surveillance system of extreme weather  
 127 events called *SUPREME* (Toutant et al., 2011), although they may include some non-heat-  
 128 related conditions. Mortality and hospitalization data were available from 1996, ED visits  
 129 and ambulance transports from 2014 and 811 calls from 2008. All health outcomes were  
 130 considered until 2019, prior to the COVID-19 pandemic.



## 131 2.2. Weather, air pollution and regional data

132 Daily maximum and minimum temperature and water vapour pressure data at 1 km × 1 km  
133 over North America were extracted from NASA's Daymet database for the 1996–2019  
134 period (Thornton et al., 2022). Mean temperature was computed from the average of daily  
135 maximum and minimum temperatures. Relative humidity, dew point and humidex were  
136 computed from vapour pressure and mean temperature using *MetPy* in Python (May et al.,  
137 2022). Humidex is a well-known index of perceived temperature, derived from temperature  
138 and dew point, widely used for heat surveillance in Canada (Smoyer-Tomic & Rainham,  
139 2001). Pixel values of weather data were aggregated at the HR level by weighting each  
140 pixel with the number of units located in each pixel using the *AQgéobâti* database  
141 (Adresses Québec, 2023), resulting in population-weighted weather times series for each  
142 HR. *AQgéobâti* provides the geocoded information of each residential address in Quebec,  
143 along with the number of units (e.g., single-family home = 1 unit, duplex = 2 units, etc.).  
144 The number of units was used as a proxy for the population, as population data at the  
145 address level was not available.

146 Hourly air pollution data was extracted at the stations of the National Air Pollution  
147 Surveillance (NAPS) program of Environment and Climate Change Canada (ECCC). All  
148 stations in the province of Quebec were included, in addition to those within 10 km of  
149 Quebec's border to supplement the database (e.g., Ottawa stations, located ~5 km from HR  
150 #7, were considered to be within this HR). NAPS hourly data was aggregated to daily mean  
151 values. Then, a spatial aggregation of daily data from all stations within the HR boundaries  
152 was performed to get a unique time series of air pollutants for each HR, as commonly done  
153 in similar studies (Boudreault et al., 2024; Lavigne et al., 2023; Masselot et al., 2019).

154 Resulting air pollution time series were complete (with less than 2% of daily missing  
155 values) for ozone concentration ( $O_3$ ) in HR #3, #4, #6, #7 and #16 from 1996 to 2019 and  
156 particulate matter ( $PM_{2.5}$ ) for the 5 same HR, but only from 2003 to 2019. There were  
157 approximately 5 to 15 stations used for the spatial aggregation in each of these five HR  
158 with complete  $O_3$  and  $PM_{2.5}$  data (Figure 1). Other air pollutants ( $PM_{10}$ ,  $NO_2$ ,  $SO_2$  and  $CO$ )  
159 were not available in sufficient regions/years to be included. Remaining missing values in  
160  $O_3$  and  $PM_{2.5}$  time series in these 5 HR were linearly interpolated. While including air  
161 pollution as a confounding effect in heat-mortality relationships is debatable (Buckley et  
162 al., 2014; Reid et al., 2012), its inclusion was considered as a sensitivity analysis to see if  
163 the heat-health relationships change when air pollution is added in the models (e.g.,  
164 Gasparrini et al., 2015; Guo et al., 2014).

165 HR-level information was considered to explain the heterogeneity that can exist across  
166 regions in the multivariate meta-regression (presented in section 2.3), including the  
167 percentage of elderly population (65+), the percentage of women, the material and social  
168 deprivation index (MSDI), the natural logarithm of population density, the normalized  
169 difference vegetation index (NDVI) and the historical summer temperature (Lavigne et al.,  
170 2023; Wang et al., 2021). Demographic data was provided by the *Ministère de la Santé et*  
171 *des Services Sociaux* (MSSS) du Québec from 1996 to 2019 (MSSS, 2022). The MSDI for  
172 its social and material components was available every census year (1996, 2001, 2006,  
173 2011, 2016 and 2021) from INSPQ (INSPQ, 2024). NDVI values during the hottest  
174 summer day without cloud interference in each region (i.e., varying spatially) were  
175 computed from 2013 Landsat-8 data by the *Centre d'enseignement et de recherche en*  
176 *foresterie de Sainte-Foy* and provided by INSPQ (CERFO, 2022) and provided by INSPQ.

177 A single year of NDVI value (i.e., 2013) was used for the whole study period, shown to be  
178 equivalent than using multiple yearly values in such studies (Pascal et al., 2021). Finally,  
179 historical temperature from the past 20 years during summer (May to September) by HR  
180 was computed from Daymet mean temperature values.

### 181 **2.3. Statistical analysis**

182 A two-stage statistical analysis was performed. First, a quasi-Poisson model was fitted in  
183 each HR to assess the association between daily fluctuations in temperature and daily  
184 mortality or morbidity. The Distributed Lag Non-Linear Model (DLNM) was employed to  
185 account for both the non-linear relationship between mean temperature and the health  
186 outcome, as well as the delayed (lagged) temperature effect (Armstrong, 2006; Gasparrini  
187 et al., 2010). Models were fitted with a lag period of 8 days (i.e., lag 0 to lag 7) before the  
188 observed mortality or morbidity (Boudreault et al., 2023; Wang et al., 2021), using a natural  
189 cubic spline with 2 knots placed on the log scale from 0 to 7 days. The non-linear heat-  
190 health association was accounted by using a natural spline with two internal knots placed  
191 at the 50<sup>th</sup> and 90<sup>th</sup> percentiles of each HR's temperature distribution (Vicedo-Cabrera et  
192 al., 2021). Day of week and holidays were considered as factors in the model. Seasonal and  
193 long-term trends were also adjusted with a natural spline of day of year with 4 degrees of  
194 freedom by year and a factor variable for each year (Vicedo-Cabrera et al., 2021). Daily  
195 relative humidity up to 7 days prior to the observed mortality/morbidity was also controlled  
196 with a linear effect.

197 Second, the region-specific estimates obtained from the first step were pooled using  
198 multivariate meta-regression models (Gasparrini & Armstrong, 2013; Sera et al., 2019).

199 This pooling procedure allowed to decrease uncertainty around region-specific heat-health  
200 associations by deriving Best Linear Unbiased Predictions (BLUP). In this study, the Risk  
201 Ratio (RR) of heat-related health outcomes were reported to the Minimum  
202 Mortality/Morbidity Temperature (MMT), the temperature at which the lowest health  
203 effect is observed. This was done by scanning temperature values from the 25<sup>th</sup> and 98<sup>th</sup>  
204 percentiles of temperatures in each HR (Vicedo-Cabrera et al., 2021), excluding potentially  
205 colder temperatures from the MMT value (i.e., if the function is decreasing for colder  
206 temperature, the MMT will refer to the 25<sup>th</sup> percentile of temperature). Sensitivity analyses  
207 were performed to validate the obtained MMT values (see last paragraph of this section).

208 Meta-regression was fitted with a forward stepwise approach using health region  
209 information. Briefly, the meta-regression was first fitted with no predictors. Then, each of  
210 the regional predictors (described in section 2.2) was tested in the model, but only the  
211 predictor minimizing the Akaike Information Criteria (AIC) was added to the model. This  
212 process was repeated until the AIC did not decrease anymore. Heterogeneity in the meta-  
213 regression was assessed using the  $I^2$  statistic and Cochran's Q-test (Lavigne et al., 2023).  
214 All analyses were performed using the R software (version 4.3.0) with the packages *dlnm*  
215 and *mixmeta*.

216 Apart from the main analysis, supplementary analyses included: 1- changing the definition  
217 of summer season to the three hottest months of summer (June–August) instead of May–  
218 September; 2- considering the 5<sup>th</sup> or the 10<sup>th</sup> percentile of temperature as the lower bound  
219 for searching the MMT value (instead of the 25<sup>th</sup> percentile); 3- changing the temperature  
220 exposure variable to minimum temperature, maximum temperature, mean humidex and

221 mean dew point, as recommended for temperature-related burden studies (Cheng et al.,  
222 2019); and 4- adding controls for daily mean concentration of O<sub>3</sub> and PM<sub>2.5</sub> (up to 7 days  
223 prior to observed mortality/morbidity) in regional estimates for which air pollution data  
224 was available for potential changes in the obtained heat-health relationships.

#### 225 **2.4. Heat burden quantification**

226 To quantify the heat burden, attributable fraction (AF) and attributable number (AN) were  
227 computed using the BLUP of the reduced exposure-response functions obtained from meta-  
228 regression for each HR (Gasparrini et al., 2015; Vicedo-Cabrera et al., 2021). AF was  
229 computed from  $AF_t = (RR_t - 1)/RR_t$ , where  $RR_t$  is the risk ratio at a given temperature value  
230  $t$ . AN was computed from  $AN_t = AF_t \times n_t$ , where  $n_t$  is the mortality/morbidity count at  
231 temperature  $t$  (Steenland & Armstrong, 2006).  $AF_t$  and  $AN_t$  for all temperature values were  
232 aggregated across the desired heat range (e.g., all heat or extreme heat only). To account  
233 for the uncertainty in the BLUP, 95% confidence intervals (CI) of AF/AN were derived  
234 from 1000 simulations of AN/AF computed by sampling the coefficients of the exposure-  
235 response curve, assuming a multivariate normal distribution (Vicedo-Cabreba et al., 2021).  
236 Two temperature ranges were considered to quantify the heat burden: 1) *all heat*, which  
237 refers to all the temperature effect above the MMT, and 2) *extreme heat*, which refers to  
238 the effect of extremely hot temperature only, defined as the temperatures above the 95<sup>th</sup>  
239 percentile of the summer temperature distribution by HR during the 1996–2019 period (Liu  
240 et al., 2019; Tupinier Martin et al., 2024; Wang et al., 2021).

241 To supplement results obtained using the regional BLUP, two other attribution methods  
242 were also considered. First, the complete bidimensional functions from regional DLNM

243 were also used (referred to as the “*regional DLNM*” method). The *attrdl* function described  
244 in Gasparrini & Leone (2014) was employed and, again, 95% CI were computed from 1000  
245 simulations of the coefficients. Second, the pooled cumulative exposure-response function  
246 over the whole province (i.e., all HR) obtained from the meta-regression was also used  
247 (referred to as the “*pooled estimate*” method). The AF/AN computation from the pooled  
248 estimate (and its uncertainty) was similar as the one described above for the regional  
249 BLUP.

### 250 **3. Results**

251 Descriptive statistics of health outcomes, weather and air pollution data are first presented  
252 (section 3.1). Then, results of the regional and pooled exposure-response curves are shown  
253 (section 3.2). Finally, the heat-related health burden is presented for the whole province of  
254 Quebec (section 3.3), as well as its temporal and spatial variability (section 3.4).

#### 255 **3.1. Descriptive statistics**

256 During May to September, there were every year approximately 23k deaths, 263k  
257 hospitalizations, 1.5M ED visits, 265k ambulance transports and 662k calls to the 811-  
258 health hotline in the province of Quebec (Table 1). HR #6 (Montréal), #16 (Montérégie)  
259 and #3 (Capitale nationale) had the highest observed health outcomes during the study  
260 period, with respectively 20–26%, 13–16% and 8–11% of the grand total across the  
261 province. HR #8 (Abitibi-Témiscamingue), #9 (Côte-Nord) and #11 (Gaspésie — Îles-de-  
262 la-Madeleine) were the HR with the fewest cases of health outcomes (i.e., <3%). Indeed,  
263 these more rural HR had less population and hence, experienced less of the total mortality

264 and morbidity. Information about socio-economic variables by HR (i.e., regional  
265 information) can be found in Table S1.

266 **Table 1. Descriptive statistics of the studied health outcomes for each health region (HR)**  
267 **and the total for the province of Quebec.** All values are the total count during May to  
268 September period, averaged across all available years of data. The percentage in parentheses is  
269 the proportion over the total for the whole province. ED = Emergency department.

HR	Mortality	Hospitalizations	ED visits	Amb. transports	811 calls
<b>1</b>	735 (3.2%)	8 264 (3.1%)	60 782 (3.9%)	7 580 (2.9%)	16 523 (2.5%)
<b>2</b>	883 (3.9%)	11 887 (4.5%)	73 260 (4.7%)	9 133 (3.4%)	26 460 (4%)
<b>3</b>	2 110 (9.3%)	22 104 (8.4%)	162 564 (10.5%)	26 359 (9.9%)	60 008 (9.1%)
<b>4</b>	1 673 (7.4%)	18 517 (7.1%)	110 657 (7.2%)	18 581 (7%)	51 836 (7.8%)
<b>5</b>	1 320 (5.8%)	16 188 (6.2%)	94 105 (6.1%)	15 023 (5.7%)	45 488 (6.9%)
<b>6</b>	5 871 (25.9%)	59 807 (22.8%)	312 656 (20.2%)	68 247 (25.7%)	144 264 (21.8%)
<b>7</b>	938 (4.1%)	9 139 (3.5%)	71 055 (4.6%)	10 505 (4%)	34 328 (5.2%)
<b>8</b>	380 (1.7%)	5 066 (1.9%)	45 297 (2.9%)	3 142 (1.2%)	10 080 (1.5%)
<b>9</b>	275 (1.2%)	4 545 (1.7%)	47 447 (3.1%)	3 247 (1.2%)	6 147 (0.9%)
<b>11</b>	334 (1.5%)	4 482 (1.7%)	45 051 (2.9%)	3 948 (1.5%)	5 955 (0.9%)
<b>12</b>	1 236 (5.4%)	15 089 (5.7%)	110 144 (7.1%)	13 158 (5%)	32 986 (5%)
<b>13</b>	994 (4.4%)	12 332 (4.7%)	53 714 (3.5%)	12 381 (4.7%)	33 709 (5.1%)
<b>14</b>	1 157 (5.1%)	14 931 (5.7%)	71 666 (4.6%)	15 211 (5.7%)	42 677 (6.4%)
<b>15</b>	1 426 (6.3%)	18 542 (7.1%)	90 935 (5.9%)	16 962 (6.4%)	47 059 (7.1%)
<b>16</b>	3 352 (14.8%)	41 665 (15.9%)	197 621 (12.8%)	41 987 (15.8%)	104 448 (15.8%)
<b>Total</b>	22 684	262 558	1 546 954	265 464	661 968

270

271 Mean temperature values during summer were higher in southernmost HR #6 (Montréal),  
272 #13 (Laval) and #16 (Montréal) with a mean value during summer of ~18°C, while  
273 northernmost HR #1 (Saguenay-Lac-Saint-Jean), #9 (Côte-Nord) and #11 (Gaspésie —  
274 Îles-de-la-Madeleine) were colder with mean summer temperature below 14°C (Table 2).  
275 Generally, maximum temperature was 4–6°C higher than mean temperature, while  
276 minimum temperature and mean dew point were 4–6°C lower. Mean humidex was  
277 approximately 5–6 units higher than mean temperature. Its standard deviation of 7–8 units  
278 was higher than for other temperature metrics that had a standard deviation of 4–5°C. Mean  
279 relative humidity during summer ranged from 65% to 75% (with a standard deviation of

280 approximately 10%) for all HR. Finally, mean daily O<sub>3</sub> and PM<sub>2.5</sub> concentration during  
 281 summer (in HR where air pollution data was available) ranged respectively from 22 to 27  
 282 ppb and from 6 to 9 µg/m<sup>3</sup>.

283 **Table 2. Descriptive statistics of weather and air pollution variables for each health region**  
 284 **(HR)**. All values are the mean daily values during May to September from 1996 to 2019, except  
 285 for PM<sub>2.5</sub> that was only available from 2003 to 2019. O<sub>3</sub> and PM<sub>2.5</sub> were only available for five  
 286 HR. Standard deviation is indicated in parentheses.

HR	Tmean °C	Tmin °C	Tmax °C	Dew point °C	Humidex	Rel. Hum. %	Mean O <sub>3</sub> ppb	Mean PM <sub>2.5</sub> µg/m <sup>3</sup>
1	13.8 (4.2)	8.6 (4.0)	18.9 (4.9)	8.5 (4.1)	18.0 (7.0)	71.0 (8.2)	-	-
2	14.8 (4.7)	8.9 (4.9)	20.8 (5.5)	8.5 (5.2)	19.2 (7.9)	66.6 (11.4)	-	-
3	16.0 (4.5)	10.4 (4.7)	21.6 (5.2)	10.3 (4.7)	21.5 (7.6)	69.4 (9.8)	22.4 (7.8)	7.7 (4.8)
4	16.7 (4.5)	11.1 (4.8)	22.4 (5.0)	10.9 (5.0)	22.7 (7.7)	69.1 (9.7)	25.1 (8.4)	7.8 (5.0)
5	16.6 (4.6)	10.9 (4.9)	22.4 (5.0)	10.8 (5.0)	22.5 (7.8)	68.9 (9.5)	-	-
6	18.2 (4.6)	12.9 (4.7)	23.5 (5.0)	12.7 (4.7)	25.4 (7.7)	70.9 (8.7)	23.6 (8.9)	9.0 (5.5)
7	17.3 (4.6)	11.3 (4.9)	23.4 (5.2)	11.0 (5.1)	23.4 (7.7)	67.3 (10.7)	25.5 (9.2)	6.2 (4.6)
8	14.3 (5.1)	8.0 (5.2)	20.6 (5.8)	7.4 (5.7)	18.0 (8.4)	64.4 (12.0)	-	-
9	12.3 (4.1)	7.7 (4.1)	16.9 (4.5)	7.6 (4.1)	15.9 (7.0)	73.6 (7.8)	-	-
11	13.8 (4.5)	9.1 (4.5)	18.4 (5.0)	9.1 (4.5)	18.4 (7.7)	73.6 (7.9)	-	-
12	15.8 (4.6)	10.0 (4.8)	21.6 (5.2)	10.0 (4.9)	21.1 (7.7)	68.8 (9.9)	-	-
13	18.2 (4.6)	12.9 (4.7)	23.5 (5.0)	12.7 (4.8)	25.4 (7.7)	70.9 (8.9)	-	-
14	17.4 (4.6)	11.7 (4.8)	23.2 (5.1)	11.5 (4.9)	23.8 (7.7)	69.0 (9.8)	-	-
15	17.1 (4.6)	11.3 (4.9)	22.9 (5.1)	11.2 (4.9)	23.3 (7.7)	68.9 (9.7)	-	-
16	18.1 (4.5)	12.5 (4.7)	23.6 (5.0)	12.4 (4.8)	25.0 (7.7)	69.9 (9.2)	26.7 (9.1)	7.8 (5.0)

287

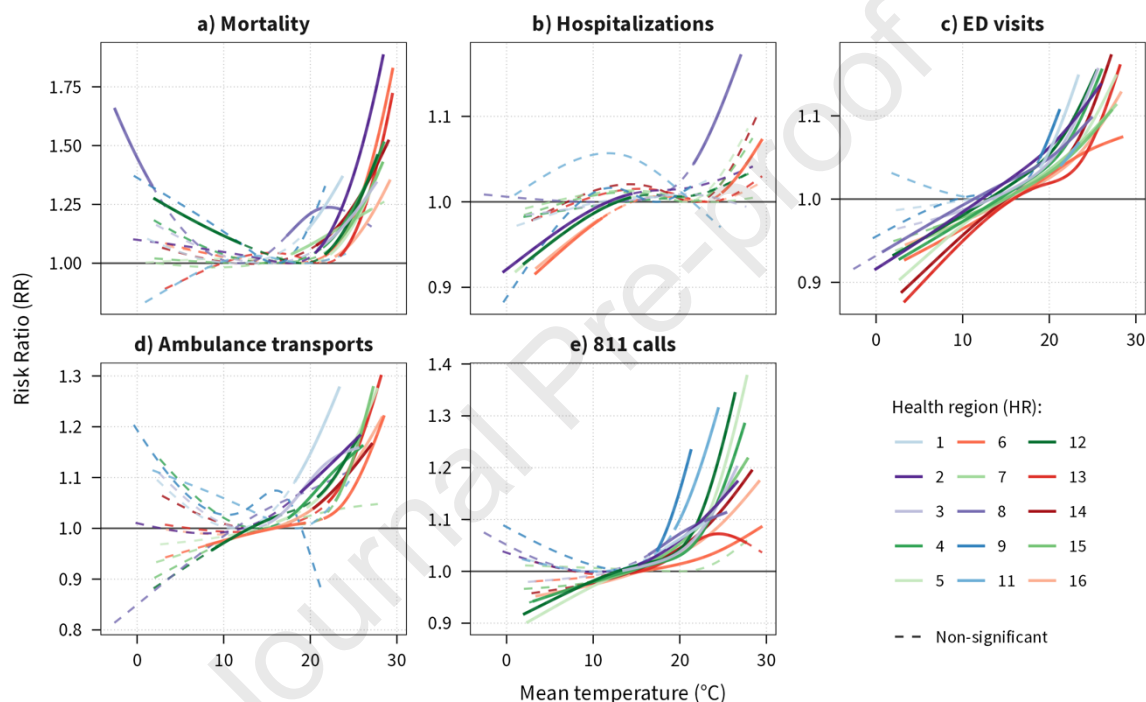
### 288 3.2. Regional and pooled cumulative exposure-response curves

289 Estimated cumulative exposure-response curves over 8 days from DLNM for the 15 HR  
 290 and 5 health outcomes are presented in Fig. 2. Overall, all health outcomes appeared to  
 291 increase with higher temperatures, but there was considerable variability between HR and  
 292 across the different health outcomes. Notably, curves in light blue (i.e., HR #11) were the  
 293 most variables with generally a different shape than the others. An important point is that



294 some of these relationships were not significant for some parts of the curves (as noted with  
 295 dashed lines on the figure), particularly for hospitalizations.

296 **Fig. 2. Cumulative exposure-response functions over 8 days in each health region (HR) from**  
 297 **regional DLNM for a) mortality, b) hospitalizations, c) ED visits, d) ambulance transports**  
 298 **and e) 811 calls.** Dashed lines indicate non-significant parts of the curves, while solid lines  
 299 indicate significant parts. Colours are set according to climate classification from red (warmer  
 300 HRs) to green to purple to blue (colder HRs).

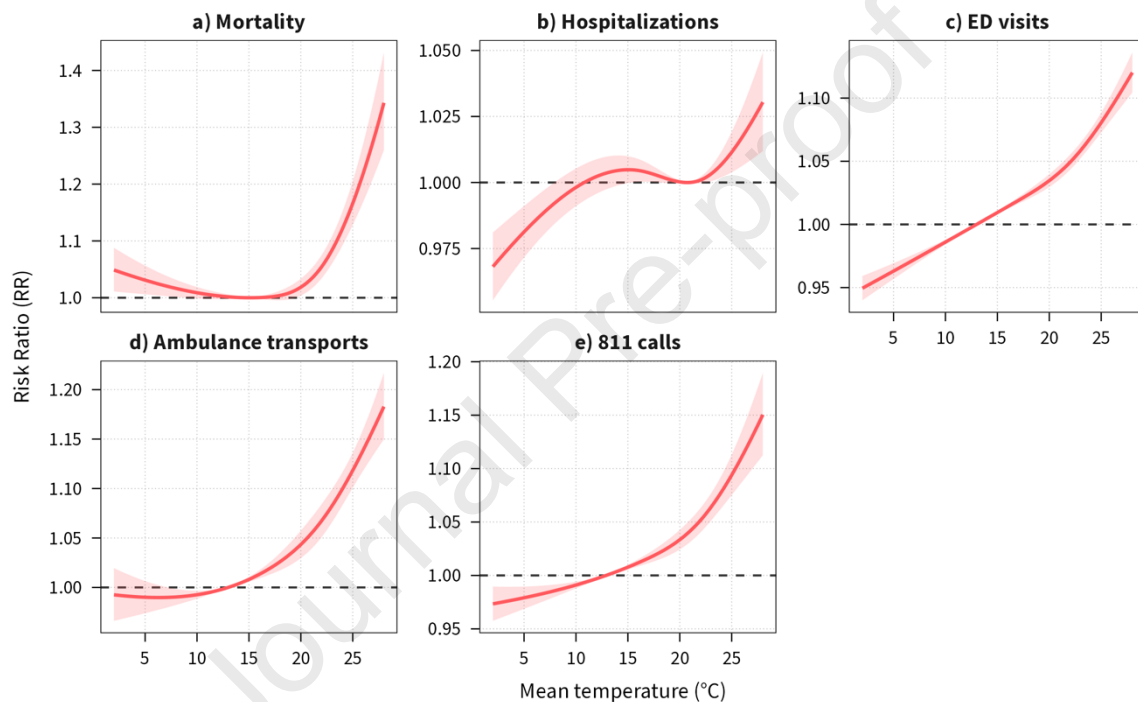


301

302 These regional estimates from DLNM were then combined in a multivariate meta-  
 303 regression model using regional information, resulting in BLUP for each HR (Fig. S1) and  
 304 pooled cumulative exposure-response curves for the whole province of Quebec (Fig. 3).  
 305 Mortality exhibited a U-shaped relationship, with an MMT at around 15°C. The exposure-  
 306 response curve for hospitalizations had a wiggly shape (~), with a potential protective effect  
 307 below 7°C. However, a significant increase in hospitalizations was noted after 22°C. ED  
 308 visits, ambulance transports and 811 calls increased consistently with higher temperatures,

309 with different inflection points. The MMT for these three health outcomes was thus  
 310 assigned to  $\sim 13^{\circ}\text{C}$  (i.e., corresponding the 25<sup>th</sup> percentile of the summer temperature  
 311 distribution, recall section 2.3).

312 **Fig. 3. Pooled cumulative exposure-response functions over 8 days in the province of**  
 313 **Quebec for a) mortality, b) hospitalizations, c) ED visits, d) ambulance transports and e)**  
 314 **811 calls. Light red region indicates the 95% confidence interval in the risk ratio.**



315

316 As seen on Fig. 2, there was some heterogeneity in the regional curves to develop these  
 317 pooled effects, but also when looking at Q-Cochran test and  $I^2$  statistics (Table S2). For  
 318 example, the Q-Cochran test was only significant for 811 calls, while the  $I^2$  statistic was  
 319  $<5\%$  for hospitalizations, ED visits and ambulance transports. Thus, results from meta-  
 320 regression should be interpreted with caution, especially for these three health outcomes.  
 321 In addition, it was found that for all health outcomes, none of the meta-predictors

322 significantly minimized the overall AIC in the meta-regression (Table S2). Hence, none of  
323 the meta-predictors was included in the model.

324 Supplementary analyses for the pooled exposure-response curves showed that considering  
325 summer to span only from June to August led to similar trends at high temperatures  
326 (Fig. S2). For most of the studied health outcomes, the effect of colder temperature had  
327 wider CI. MMT values were consistent for mortality and hospitalizations, but slightly  
328 higher (i.e.,  $\sim 5^{\circ}\text{C}$  higher) for the other health outcomes (e.g., ED visits, ambulance  
329 transports and 811 calls) when using only June to August data. When testing lower  
330 percentiles for searching the MMT value (i.e., 5<sup>th</sup> and 10<sup>th</sup>), lower MMT values (i.e.,  $\sim 5^{\circ}\text{C}$   
331 lower) were found for all health outcomes, except for mortality (Fig. S3). Exposure-  
332 response curves with other temperature metrics showed similar shapes than when using  
333 mean temperature (Fig. S4). Finally, adjusting models for  $\text{O}_3$  and  $\text{PM}_{2.5}$  (in HR for which  
334 air pollution data was available) did not change the shape of the pooled cumulative curve  
335 for most health outcomes, only slightly for hospitalizations when both  $\text{O}_3$  and  $\text{PM}_{2.5}$  were  
336 controlled for (Fig. S5).

### 337 **3.3. Heat burden in the province of Quebec**

338 AF and AN for all health outcomes derived from regional BLUP were computed across the  
339 whole province (Table 3). Supplementary analyses (e.g., with other temperature metrics,  
340 air pollution controls, etc.) were not pursued in this section as pooled exposure-response  
341 curves presented in section 3.2 did not differ significantly in these analyses. Approximately  
342 2–3% of summer mortality, ED visits, ambulance transports and 811 calls were attributable  
343 to heat, but only 0.1% for hospitalization. Less than 1% of each health outcome was

344 attributable to extreme heat (i.e., temperatures above the 95<sup>th</sup> percentile). AF/AN to heat  
 345 and extreme heat were all statistically significant. In terms of AN, 471 and 203 deaths every  
 346 summer were attributable respectively to all heat and extreme heat exposures, 226 and 170  
 347 hospitalizations, 36 273 and 6 228 ED visits, 7 177 and 1 517 ambulance transports, and  
 348 15 058 and 3 3306 calls to the 811-health hotline (Table 3). Differences between all heat  
 349 and extreme heat exposures were greater for ED visits, ambulance transport and 811 calls  
 350 (with factors of 4X–6X for all heat vs. extreme heat), while for deaths and hospitalizations,  
 351 the difference was much smaller (1.5X–2.5X).

352 **Table 3. Attributable number (AN) and fraction (AF) in the province of Quebec computed**  
 353 **from regional BLUP for all heat and extreme heat exposures.** AN are presented as the yearly  
 354 number of cases every summer across the studied period. ED = Emergency department.

	All heat		Extreme heat	
	AN	AF	AN	AF
<b>Mortality</b>	471 (251 – 693)	2.1% (1.1–3.1)	203 (138 – 267)	0.9% (0.6–1.2)
<b>Hospitalizations</b>	226 (138 – 318)	0.1% (0.1–0.1)	170 (118 – 226)	0.1% (0.0–0.1)
<b>ED visits</b>	36 273 (32 328 – 40 060)	2.3% (2.1–2.6)	6 228 (5 589 – 6 870)	0.4% (0.4–0.4)
<b>Ambulance transports</b>	7 177 (4 594 – 9 592)	2.7% (1.7–3.6)	1 517 (1 326 – 1 700)	0.6% (0.5–0.6)
<b>811 calls</b>	15 058 (9 270 – 20 990)	2.3% (1.4–3.2)	3 306 (2 413 – 4 206)	0.5% (0.4–0.6)

355

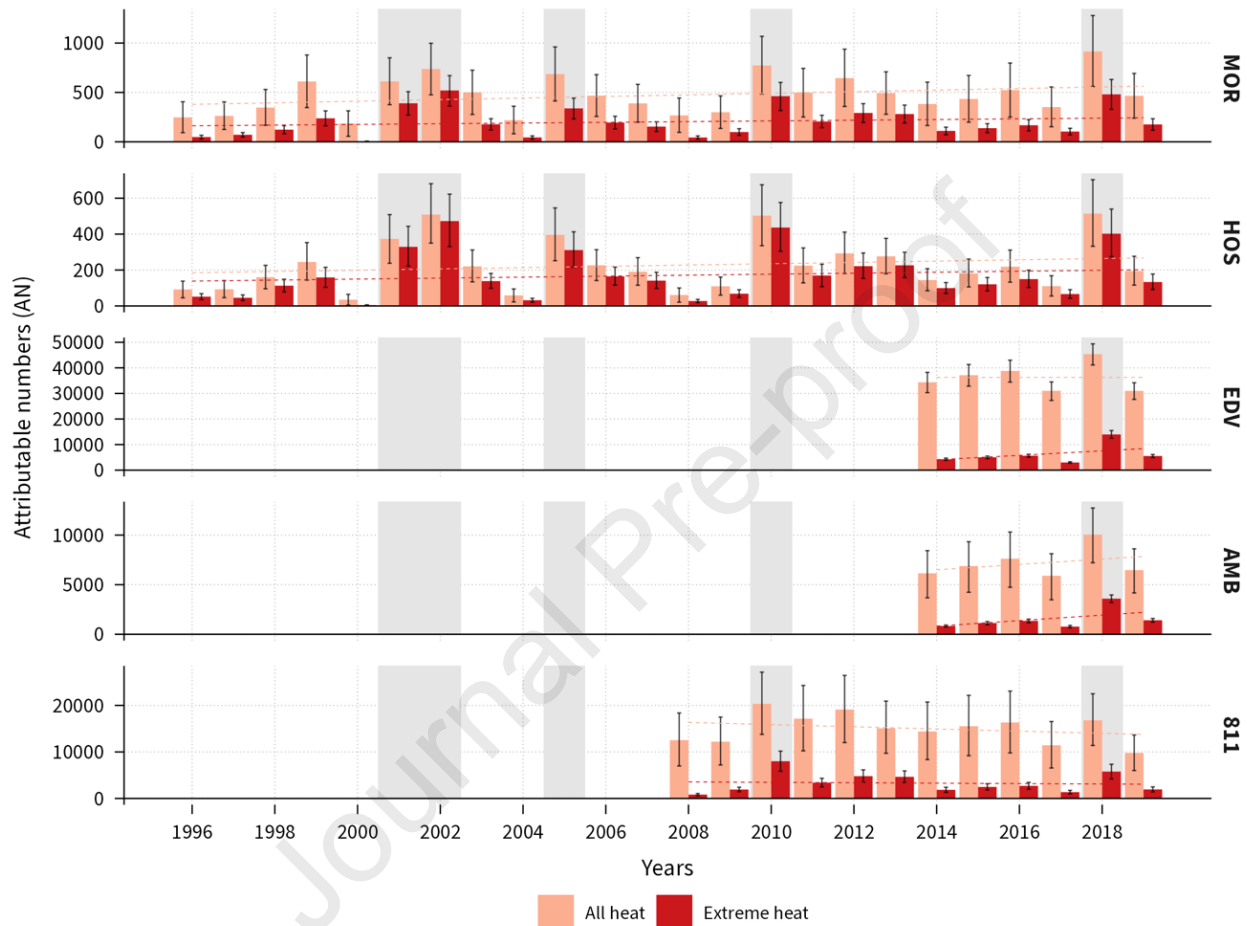
356 Comparison of results using regional DLNM, regional BLUP and pooled estimates  
 357 generally showed consistent values (within the same CI) for AN (Fig S6a) and AF (Fig  
 358 S7a) for the total heat exposure. However, higher AN/AF for hospitalizations were found  
 359 when using regional DLNM, but these values were not statistically significant anymore.  
 360 For extreme heat exposure, AN (Fig. S6b) and AF (Fig. S7b) from pooled estimates yielded  
 361 much lower values (1.5-2.0X smaller) than from regional DLNM and BLUP that were  
 362 consistent. CI were generally wider with regional DLNM than with BLUP, except for

363 mortality and 811 calls. Numerical values of AN/AF computed from the regional DLNM  
364 and pooled estimates are available in Table S3.

### 365 **3.4. Temporal and spatial variability of the heat burden**

366 AN were also computed separately for each year of observations (Fig. 4 and Table S4 for  
367 numerical results of the last 10 years of data). Years with higher AN were respectively  
368 2001, 2002, 2005, 2010 and 2018. For mortality, almost half the heat related AN was due  
369 to extreme heat (i.e., the 5% hottest days). This proportion was even higher for  
370 hospitalizations. For all health outcomes (except 811 calls), there were linear increasing  
371 trends in AN to total and extreme heat over the studied period. During summer of 2018  
372 (during which all health outcomes data was available), approximately 480 deaths, 400  
373 hospitalizations, 14 000 ED visits, 3 600 ambulance transports and 5 800 calls to the 811-  
374 health hotline were attributable to extreme heat (Table S4).

375 **Fig. 4. Attributable numbers (AN) to all heat and extreme heat exposures during 1996–2019**  
 376 **based on regional BLUP across Quebec.** Error bars represent the 95% CI. Dotted lines are  
 377 linear trends over the studied period. Grey regions are years with major heatwaves. Data was not  
 378 available for 1996–2013 for EDV and AMB, and 1996–2007 for 811 calls. MOR = Mortality.  
 379 HOS = Hospitalizations. EDV = Emergency department visits. AMB = Ambulance transports.  
 380 811 = Health hotline calls.

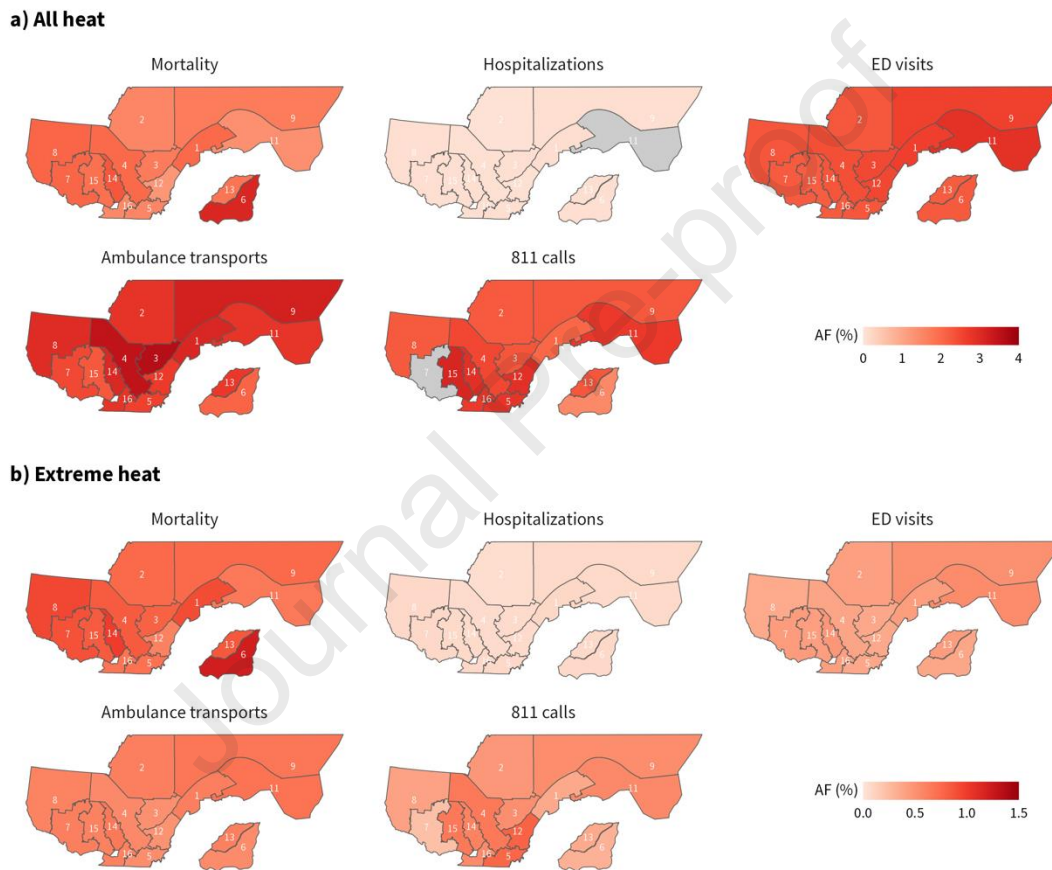


381

382 AF were also computed separately by HR (Fig. 5 and Table S5 for numerical results of AF  
 383 and AN). Significant AF due to all heat ranged from 1.5–3.0% for mortality, 0.1–0.1% for  
 384 hospitalizations, 2.0–3.0% for ED visits, 2.0%–3.5% for ambulance transports and 1.5–  
 385 3.0% for 811 calls (Fig. 5). These AF were all <1.5% when looking only at extreme heat  
 386 exposure. AF to all heat exposure in HR #11 for hospitalizations and in HR #7 for 811 calls  
 387 was not significant. The example of the HR with the highest population (Montreal, #6)

388 showed statistically significant AN of 180 (71) deaths, 53 (41) hospitalizations, 6 995  
 389 (1 171) ED visits, 1 415 (368) ambulance transports and 2 157 (446) calls to the 811 due to  
 390 all heat (and extreme heat) every summer in average (Table S5).

391 **Fig. 5. Attributable fraction (AF) in each health region (HR) computed from regional BLUP**  
 392 **for a) all heat and b) extreme heat exposures.** Non statistically significant AF are in grey. Refer  
 393 to Table S5 for numeric results of AF (and AN).



#### 394 4. Discussion

395 This study considered a two-stage statistical analysis with DLNM and multivariate meta-  
 396 regression to estimate the heat burden for five health outcomes, including both mortality  
 397 and morbidity, in the province of Quebec (Canada). While there was some notable  
 398 variability and heterogeneity in the regional exposure-response functions (recall Fig. 2 and

399 Table S2), the pooled estimates showed that all studied health outcomes had higher RR at  
400 hotter temperatures, but at different levels. ED visits, ambulance transports and 811 calls  
401 did not exhibit the classical U- or J-shape relationship as observed with mortality, but rather  
402 monotonic increases in RR with higher temperatures. These less familiar relationships  
403 highlighted the need to correctly assign the MMT value, which will ultimately affect the  
404 calculation of the heat burden. MMT values were slightly higher when focussing on the  
405 three hottest months of summer (Fig. S2), but lower when using lower percentiles of  
406 temperature (e.g., 5<sup>th</sup> or 10<sup>th</sup>) to extract MMT values (Fig. S3). Thus, the reference MMT  
407 values used for these three health outcomes should be seen as conservative in our study.  
408 Further work is therefore recommended to that end for morbidity variables, as well as  
409 explaining the mechanisms between lower temperatures and these health impacts, that was  
410 out-of-scope for the current study focussing on heat. For mortality and hospitalizations,  
411 MMT values were consistent in the main and supplementary analyses.

412 Exposure-response functions obtained from BLUP were then used to compute the health  
413 burden associated with exposure to all heat (i.e., temperatures above the MMT) and  
414 extreme heat (i.e., temperatures above the 95<sup>th</sup> percentile) for the whole province, across  
415 years and regions. Results from BLUP were compared to regional DLNM (without meta-  
416 regression adjustment) and pooled estimates across the province. For the total heat  
417 exposure, the associated burden had consistent results for the three tested methods, except  
418 for hospitalizations when using regional DLNM (i.e., non-significant AN/AF). For extreme  
419 heat, both regional methods (BLUP and DLNM) had higher values (1.5-2X) than pooled  
420 estimates for the whole province, which could be due to a poorer representation of extreme  
421 temperatures in a unique pooled estimate for the whole province compared to regional



422 estimates (recall Fig. S6 and Fig. S7). To the best of our knowledge, this study is one of  
423 the first to provide such a detailed portrait of the heat burden at a national scale combining  
424 multiple health outcomes, employing a robust and state-of-the-art methodology, as well as  
425 testing for different summer definitions, temperature exposures, air pollution adjustments  
426 and methods for the heat burden quantification.

427 In terms of mortality, the burden for the whole province was equivalent to AF of 2.1% of  
428 all summer mortality, representing AN of ~470 deaths each summer. When only the 5%  
429 hottest days of summer were analyzed, AF and AN were respectively of 0.9% and ~200  
430 deaths. In Montreal (Canada), Benmarhnia et al. (2014) found 62 deaths each summer  
431 attributable to heat during June to August months (years 1990–2007), while we found 180  
432 deaths during the May–September period (years 1996–2019) for this location (recall  
433 Table S5). In Canada, Hebbern et al. (2023) found AF of 0.41% of annual mortality due to  
434 heat during the 2010–2019 period. In the multi-city multi-country study of Gasparri et al  
435 (2015), the AF in Canada was 0.54% for all heat (0.68% in Montreal only) and 0.26% for  
436 extreme heat (above the 97.5<sup>th</sup> percentile), again reported over annual mortality. These last  
437 two results are consistent with ours, given that our AF were reported over summer mortality  
438 only.

439 During the heatwaves of 2010 and 2018 in Quebec, public health authorities reported  
440 respectively an excess of 280 (Bustinza et al., 2013) and 210 (Lebel et al., 2019) deaths  
441 across the province. We found that 462 and 480 deaths were associated to extreme heat  
442 during these two years, respectively (recall Fig. 4 and Table S4). In addition, there was  
443 almost no significant impact on mortality during 2011–2015 official heatwaves in Quebec

444 (Lebel et al., 2017), while we have found significant AN due to heat and extreme heat  
445 during all those years. These differences are mainly due to the definition of heatwaves used  
446 in the aforementioned studies (i.e., three consecutive days above thresholds), whereas  
447 extreme heat was defined as all days with temperatures above the 95<sup>th</sup> percentile in our  
448 study. Globally, the review of Cheng et al. (2019) reported that AF to heat were <2.0% for  
449 mortality (when only summer months were analyzed), which is consistent with our findings  
450 for the province of Quebec (i.e., AF of ~2.0%).

451 The AF for hospitalizations was found to be 0.1% (to both all heat and extreme heat  
452 exposures), representing AN of ~225 hospitalizations to heat and ~170 to extreme heat  
453 every summer across the province. In Ontario (Canada), Bai et al. (2016, 2018) found that  
454 1.4% of hypertension, 11.2% of diabetes, 1.2% of coronary heart disease and 1.8% of  
455 stroke hospitalizations were due to heat. These AF are much higher than our results, but  
456 focussed on specific diseases that are known to be temperature-related (e.g., Liu et al.,  
457 2022; Phung et al., 2016). Indeed, focussing on all-cause hospitalizations could dilute the  
458 heat-hospitalization relationship, making the heat effect harder to detect. Studying cause-  
459 specific morbidity (or mortality) is left for future research. Furthermore, as our AF was  
460 computed over all hospitalizations, the obtained value will be lower than if it were  
461 calculated only for cause-specific hospitalizations as in Bai et al. (2016, 2018). Around the  
462 world, the study of Lin et al. (2012) found that ~100 hospitalizations for respiratory causes  
463 were attributable to extreme heat each year in the state of New York (~20M hab.).  
464 Wondmagegn et al. (2021a) found that ~500 hospitalizations were attributable to heat every  
465 year in Adelaide, an Australian city of ~1.3M inhabitants. Finally, Liu et al. (2019) noted  
466 that 0.8% of hospitalizations in the 0–19 age group were related to moderate heat in the

467 Minneapolis/St. Paul region (USA), while AF for other age groups were not statistically  
468 significant.

469 The heat burden on ED, ambulance services and health hotline was considerable in our  
470 study with AN of respectively 36 000, 7 000 and 15 000 cases due to heat every summer in  
471 Quebec, representing AF of 2.0–3.0%. During the 2010 heatwave in Quebec, Bustinza et  
472 al. (2013) noted an excess of 3 400 ED visits. While no ED data was available in 2010 in  
473 our study, we found ~14 000 ED visits associated with extreme heat in 2018, a comparable  
474 year to 2010 in terms of heatwave intensity. Kegel et al. (2021) studied the relationship  
475 between ED visits and summer temperatures in two hospitals of Montreal (Canada) and  
476 found that higher temperatures were associated with a higher number of patients in ED  
477 (same as in our study), but they did not report any AF nor AN. There was no other burden  
478 quantification study found in the literature for these morbidity variables in Quebec or  
479 Canada. Around the world, Wondmagegn et al. (2021b) found that ~290 ED visits were  
480 heatwave-related every year in Adelaide while Wellenius et al. (2017) estimated ~7 200  
481 ED visits attributable to heat in 15 locations of New England, USA (2.7 M hab.). Cheng et  
482 al. (2016) reported AF of 2.2% for extreme heat-related ambulance transports in the city of  
483 Huainan (China) while Li et al. (2021) found AF of 11.7% to high temperatures for  
484 ambulance calls in 11 Chinese cities. In its review of temperature-related health burden,  
485 Cheng et al. (2019) reported that AF to heat was <1.4% for all-cause morbidity globally  
486 (though without differentiating by health outcomes), whereas higher figures were reported  
487 in our study (i.e., AF of 2–3% for ED visits, ambulance transports and calls to a health  
488 hotline).

489 Even though the above comparisons are interesting, they should be interpreted with  
490 caution, particularly for studies performed in other regions given the differences that exist  
491 in population size, demography, climate and health service management, among others  
492 (Gasparrini et al., 2015). In addition, in all studies cited above, there is not a single and  
493 universal temperature metrics used (e.g., mean, maximum, minimum temperature, heat  
494 index, etc.), no consistent definition of the total and extreme heat exposure (e.g., above  
495 different MMT values or percentiles such as 90<sup>th</sup>, 95<sup>th</sup> or 97.5<sup>th</sup>) and multiple heatwaves  
496 definitions (e.g., intensity, duration, temperature metrics), leading to even greater  
497 challenges in comparing the heat-related burden (Cheng et al., 2019). Thus, it highlights  
498 the need for more robust and comparable studies using a unified methodology that  
499 simultaneously quantify multiple health outcomes, as we have presented in this study and  
500 applied to Quebec, Canada.

501 By examining total and extreme heat exposures on a broader scaler than only during  
502 “official” heatwaves as was previously done in Quebec (e.g., Bustinza et al., 2013; Lebel  
503 et al., 2017; 2019), a new perspective on the heat load has been provided, which can be  
504 used to improve surveillance and protection activities by public health authorities. In  
505 addition, the variability of the heat burden between years (temporal) and health regions  
506 (spatial) was another insight that was revealed by our study. On the temporal side, years  
507 2001, 2002, 2005, 2010 and 2018 were found to have the greatest health burden (recall  
508 Fig 4. and Table S4). The two last years (2010 and 2018) are also the ones during which  
509 recent major heatwaves occurred in Quebec. On the spatial side, the variability found in  
510 AF across health regions can be associated with different vulnerability factors such as  
511 demography, deprivation, built environment and adaptation to heat, among others (recall

512 Fig 5. and Table S5). These results can therefore help prioritize the most effective actions  
513 to reduce the heat burden locally and in a timely manner. However, as meta-regressions  
514 did not detect any significant predictors (recall Table S2), future work is therefore  
515 recommended to explore the vulnerability factors associated with heat in Quebec at a finer  
516 scale, as well as to better understand the heterogeneity that was found between HR across  
517 the different studied health outcomes.

518 Our study has several strengths. First, it used a well-known and state-of-the-art framework  
519 combining DLNM and meta-regression to derive the heat-related health burden. Second, 5  
520 health outcomes (including both mortality and morbidity) were included, most of them  
521 only seldom studied in Quebec and Canada, but also in the literature more generally. Third,  
522 our study was performed for the 15 southernmost health regions of Quebec (i.e., 99% of  
523 Quebec's population of 8.5M inhabitants), a geographic scale representing a trade-off  
524 between statistical power and relevance for decision makers. Fourth, the computed AF and  
525 AN from regional BLUP were consistent compared to estimates from regional DLNM or  
526 pooled functions across the province, though some differences were also noted. Finally,  
527 supplementary analyses including a shorter definition of summer, different percentiles for  
528 the MMT value, four other temperature exposure variables and adjusting for air pollution,  
529 showed generally consistent results compared to the main analysis.

530 Some limitations must also be discussed. First, no stratification by sex/gender, age or cause  
531 of disease were performed. Our goal was to quantify the overall heat burden, not to assess  
532 the differences in heat-health relationships among population strata. This is left for further  
533 research, as well as the study of smaller geographical units that could be more suitable than

534 the HR spatial scale for stratified analysis of some vulnerability factors such as built  
535 environment. Second, only 6 years of data was available for ED visits and ambulance  
536 transports (2014–2019), while longer time series (>10 years) were available for all other  
537 health outcomes. This limited the comparison of our results for ED visits and ambulance  
538 transports with the ones observed during the 2010 heatwave in Quebec. Third, the burden  
539 was only estimated for all heat (above the MMT) and extreme heat (above the 95<sup>th</sup>  
540 percentile) exposures, without studying, for example, the excess burden during heatwaves  
541 or the exposure to cold temperatures. Fourth, daily air pollution data was only fully  
542 available in 5 HR for the sensitivity analysis, while the low spatio-temporal resolution of  
543 air pollution data in Quebec prevented us from using population-weighted time series (as  
544 for weather data) in this analysis. Finally, while results were provided for other temperature  
545 exposure metrics, these analyses could be deepened in a future work.

## 546 **5. Conclusion**

547 By representing a state-of-the-art quantification of the heat burden, this study can serve as  
548 a reference for further studies in other parts of the world also experiencing the adverse  
549 effect of heat, drawing on the same detailed methodology and a comparable comprehensive  
550 database (i.e., several health outcomes in multiple regions combined with various  
551 temperature, air pollution and regional variables). Applied to the province of Quebec  
552 (Canada), all studied health outcomes had higher risk ratios at hotter temperatures, but at  
553 different levels. In terms of burden, heat translated into a massive load on the health system  
554 every summer in Quebec, with significant numbers of 470 (200) deaths, 225 (170)  
555 hospitalizations, 36 000 (6 200) ED visits, 7 200 (1 500) ambulance transports and 15 000

556 (3 300) calls to a health hotline due to heat (and extreme heat). Results also revealed a great  
557 deal of temporal (across years) and spatial (across health regions) variability. This new  
558 information will help public health authorities to target appropriate interventions  
559 depending on the health outcome, region and type of heat exposure, to reduce the heat-  
560 related health burden now and in the future.

## 561 **Acknowledgments**

562 The authors would like to thank Denis Hamel and Louis Rochette for their help with the  
563 health data extraction, Antoine Saint-Amand, Mathieu Tandonnet and Nathalie Gravel for  
564 their help with the spatial data extraction, Félix Lamothe, Magalie Canuel, Ray Bustinza  
565 and Annabel Ruf for their comments on this work, and Vaibhavi Mayya for her preliminary  
566 work on this project. The authors would also like to thank the editor, the associate editor  
567 and three anonymous reviewers for their comments that helped improve this manuscript.

## 568 **Funding**

569 The first author has received funding from the Natural Sciences and Engineering Research  
570 Council of Canada (Vanier Scholarship, #CGV-180821), the Canadian Institute of Health  
571 Research (Health System Impact Fellowship, #IF1-184093), Ouranos (Real-Décoste  
572 Excellence Scholarship, #RDX-317725) and the National Institute of Public Health of  
573 Quebec (no grant number).

## 574 **Credit authorship contribution statement**

575 **Jérémie Boudreault:** Conceptualization, Methodology, Data Curation, Formal analysis,  
576 Visualization, Software, Writing—Original Draft, Review and Editing, Funding

577 acquisition. **Éric Lavigne:** Conceptualization, Methodology, Software, Writing—Review  
578 and Editing. **Céline Campagna:** Conceptualization, Writing—Review & Editing,  
579 Supervision, Project administration, Funding acquisition. **Fateh Chebana:**  
580 Conceptualization, Writing—Review & Editing, Supervision, Project administration,  
581 Funding acquisition.

## 582 **Declaration of competing interest**

583 The authors declare that they have no known competing financial interests or personal  
584 relationships that could have appeared to influence the work reported in this paper.

## 585 **Data availability**

586 Authors do not have permission to share health data. Socioenvironmental data are freely  
587 available from the following organization: NASA for weather data, Environment and  
588 Climate Change Canada (ECCC) for air pollution data, Ministère de la Santé et des  
589 Services Sociaux (MSSS) for demographic data and Institut national de la santé publique  
590 du Québec (INSPQ) for socio-economic and built environment data. A code example to  
591 reproduce the conducted analyses with synthetic health data is available on the first  
592 author's Github page ([https://github.com/jeremieboudreault/paper\\_heat\\_burden\\_qc](https://github.com/jeremieboudreault/paper_heat_burden_qc)).

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802



**Highlights :**

- The heat-related burden of five health outcomes (HO) quantified in Quebec, Canada
- DLNM and meta-regression used to compute attributable fraction (AF) and number (AN)
- All HO showed increased risk ratios at higher temperatures, but at different levels
- AF to heat were 2–3% for all HO except for hospitalizations (0.1%), all significant
- Depending on the HO, AN to heat ranged from 100s to 10 000s each summer in Quebec

**Declaration of interests**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Journal Pre-proof