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## Seasonal patterns in greenhouse gas emissions from thermokarst lakes

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#### **Abstract**

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In the ice-rich permafrost area of Central Yakutia (Eastern Siberia, Russia), climate warming and other natural and anthropogenic disturbances have caused permafrost degradation and soil subsidence, resulting in the formation of thermokarst (thaw) lakes. These lakes are hotspots of greenhouse gas emissions, but with substantial spatial and temporal heterogeneity across the Arctic. We measured dissolved CO2 and CH4 concentrations in thermokarst lakes of Central Yakutia and their seasonal patterns over a yearly cycle. Lakes formed over the Holocene (alas lakes) are compared to lakes that developed over the last decades. The results show striking differences in dissolved greenhouse gases (up to two orders of magnitude) between lake types and seasons. Shallow lakes located in hydrologically closed alas depressions acted as CO<sub>2</sub> sinks and strong sources of diffusive CH<sub>4</sub> during some seasons (ebullition was not assessed). Recent thermokarst lakes were moderate to extremely high sources of diffusive CO<sub>2</sub> and CH<sub>4</sub>, with considerable accumulation of greenhouse gas under the ice cover (winter) or in the deepest water layers (summer), highlighting the need to include spring and autumn as critical periods for integrated assessments. The water column was stratified in winter (all lake types) and especially in summer (thermokarst lakes), generating anoxia in bottom waters and favoring  $CH_4$  production and storage, particularly in the most organic-rich lakes. The diffusive fluxes measured from thermokarst lakes of this typical taiga alas landscape of Central Yakutia are among the highest presented across Arctic and subarctic regions.

Permafrost occupies more than 20 million square kilome-35 ters and represents 24% of land cover within the northern 36 hemisphere (Brown et al. 1998). It is especially abundant in 37 Siberia, Alaska, and Canada, and its spatial extent, thickness, 38 and ground ice content can vary widely across landscapes (Grosse et al. 2013; Strauss et al. 2017). Studies on permafrost landscape dynamics during the Holocene and in recent decades have shown that areas dominated by ice-rich perma-42 frost are very sensitive to changes in temperature and other 43 local disturbances (Grosse et al. 2013; Ulrich et al. 2019). Ris-<sup>44</sup> ing air and ground temperatures result in permafrost thaw, which can have widespread implications for local and regional hydrology (Biskaborn et al. 2019). Permafrost thaw can release substantial amounts of organic and mineral matter to aquatic ecosystems causing profound changes in their biogeochemistry and their role in the global carbon cycle (Vonk et al. 2015

and references therein). The current rate and magnitude of 88 temperature rise in the Arctic is disproportionately high compared to global averages, with mean annual air temperature predicted to increase by as much as 5.4°C within the next century in the absence of significant and directed global effort to reduce greenhouse gas emissions (IPCC 2019). This will likely herald a period of dynamic changes within permafrost landscapes.

A prevalent pathway of permafrost degradation in areas of ice-rich permafrost is the initiation of thermokarst processes, which eventually result in the formation of numerous lakes in regions where topography is flat (Grosse et al. 2013). Thermokarst processes begin when disturbances such as warming or forest removal cause deepening of the active layer (the layer which thaws each summer), which results in either thaw subsidence or thermal erosion depending on the relief of the terrain and the ground ice content. In areas that are dominated by low relief terrain, such as Central Yakutia, thermokarst subsidence often causes ponding and shallow depressions (French 2017). The coalescence and expansion

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Additional Supporting Information may be found in the online version of 54 this article.

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1 (both laterally and vertically) of these ponds eventually results 2 in the development of larger and deeper lakes, with a portion 3 of unfrozen water remaining under the ice cover in winter 4 (Bouchard et al. 2020). Once formed, these lakes profoundly 5 change the local ground thermal regime, sometimes increas-6 ing surrounding sediment temperatures by as much as 10°C 7 above the mean annual air temperature 8 et al. 2004). Thermokarst processes, especially the presence of 9 a lake or pond, increase the rate of permafrost thaw signifi-10 cantly compared to what would be predicted from increases in 11 air temperature alone (Brouchkov et al. 2004; Schuur 12 et al. 2015). Continued lateral lake expansion can cause slope 13 slumping (so-called "retrogressive thaw slumps") or further 14 ground surface subsidence and erosion (Séjourné et al. 2015; 15 Bouchard et al. 2017). Expansion and deepening continue 16 until the lake depth surpasses the depth of ice-rich permafrost 17 or thaw propagation beneath the lake is impeded by lake sedi-18 ment insulation properties, halting further deepening. After 19 this phase, lake size and depth are controlled mostly by the 20 ratio between precipitation and evaporation (Soloviev 1973; Grosse et al. 2013; French 2017). Although some of the mech-22 anisms involved are not well understood, drainage, evapora-23 tion and infilling eventually result in lake disappearance 24 (Soloviev 1973; Desyatkin et al. 2009).

25 It is estimated that soils in northern permafrost landscapes 26 contain twice as much carbon as currently exists in the atmosphere (Hugelius et al. 2014). Specifically, thermokarst lakes 28 are considered as biogeochemical hotspots that play a pivotal 29 role in the processing of permafrost organic matter and there-30 fore potentially on the global climate (Walter Anthony 31 et al. 2016). Anoxic sediments at the bottom of these lakes are 32 sites of CH<sub>4</sub> production that can be released through ebulli-33 tion and diffusion. The CH<sub>4</sub> released though ebullition in 34 shallow lakes has little opportunity to be oxidized within the 35 water column, so it typically represents a direct flux to the 36 atmosphere (Bastviken et al. 2004; Bouchard et al. 2015). 37 However, the CH<sub>4</sub> released through diffusive flux, or through 38 ebullition in deeper stratified lakes, remains in the water col-39 umn long enough so that a significant fraction can be oxi-40 dized to CO<sub>2</sub> by methane-oxidizing bacteria (Vachon et al. 2019). In winter, the CH<sub>4</sub> emitted through both pro-42 cesses can be oxidized under the ice cover (Matveev 43 et al. 2019). Significant release of CO<sub>2</sub> from thermokarst lakes northern 44 in latitudes has also been documented 45 (e.g., Abnizova et al. 2012). There are three main pathways by 46 which CO<sub>2</sub> can be released from thermokarst lakes, besides 47 when it originates from CH<sub>4</sub> oxidation: (1) respiration within 48 the oxic water column and sediments (MacIntyre et al. 2018), 49 (2) lateral transfer of dissolved inorganic carbon from sur-50 rounding soils by groundwater flow and carbonate weathering 51 (Zolkos et al. 2018), and (3) photo-oxidation of dissolved 52 organic carbon (Ward et al. 2017). Lateral transport of CH<sub>4</sub> 53 from the active layer was also shown to represent a significant

allochthonous source in Toolik Lake, Alaska (Paytan et al.

2015). Overall, the seasonal release of greenhouse gases pro- 55 duced from these multiple sources is largely controlled by the 56 ice cover and the mixing regime of lakes (Sepulveda-Jauregui 57 et al. 2015; Matveev et al. 2019).

Little is known about the seasonality of greenhouse gas 59 emissions from thermokarst lakes developed within perma- 60 frost. This is particularly true for winter and spring seasons 61 that present logistical constrains (Matveev et al. 2019). Here 62 we report on dissolved greenhouse gas concentrations, and 63 derived diffusive fluxes, in thermokarst lakes of different 64 development stages in Central Yakutia, Eastern Siberia, a 65 region experiencing rapid degradation of ice-rich permafrost. 66 The limnological properties and greenhouse gas concentra- 67 tions were quantified four times throughout a full annual 68 cycle (2018-2019). We tested the two-fold hypothesis that 69 (1) marked local landscape heterogeneities (permafrost condi-70 tions, lake morphology and water balance) strongly affect 71 greenhouse gas concentrations and fluxes in thermokarst lakes 72 of Central Yakutia, and (2) these spatial patterns are further 73 influenced by strong seasonal changes in lake water temperature, mixing regime, and dissolved oxygen.

#### Study area

Central Yakutia is characterized by extreme subarctic continental climate with long, cold and dry winters (January mean 80 temperature around -40°C) and warm summers (July mean 81 temperature around +20°C), resulting in notably strong sea- 82 sonal variability. The winter season (defined by the presence 83 of an ice cover) usually lasts from late September until early 84 May. The low annual precipitation (190-230 mm) is generally 85 restricted to summer. Average snow depth for winter months 86 (January to April) ranges from 24 cm in January to a maximum of 30 cm in March, and then decreasing to 10 cm at the 88 end of April (1980-2020 recorded values from Yakutsk 89 weather station). Yearly evaporation rates exceed total precipi- 90 tation in this region (Fedorov et al. 2014b). Central Yakutia is no exception to the observed trend of high latitude regions 92 warming disproportionately quicker than lower latitudes. 93 Between 1996 and 2016, the mean annual air temperature of 94 Central Yakutia has increased by 0.5–0.6°C per decade (Gorokhov and Fedorov 2018).

The study site (62.554°N; 130.982°E) is located near the 97 rural village of Syrdakh, which lies on the lowland plain between the Lena River to the west and the Aldan River to the 99 east, approximately 130 km north-east of Yakutsk (Fig. 1). The region is predominantly covered by late Pleistocene sediments, including silty clays and sandy silts of fluvial, lacustrine or eolian origin (Ivanov 1984). The activity of the Lena 103 and Aldan rivers, as well as their smaller tributaries, has 104 resulted in the formation of numerous fluvial terraces during 105 the Pleistocene (Soloviev 1959). Lakes later developed within 106 the late Pleistocene-aged fluvial terraces: the Tyungyulyu ter- 107 race, which covers the western section of the study area,

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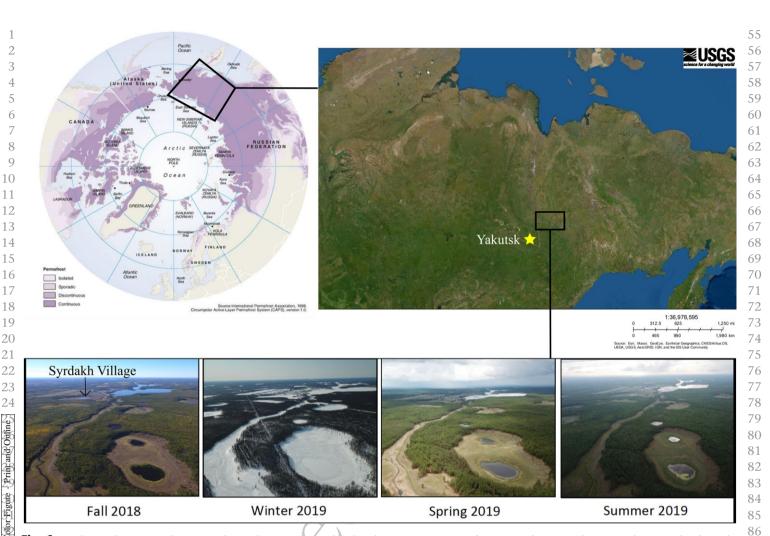


Fig. 1. Study area location and context. The study area is located within the continuous permafrost zone of eastern Siberia (a), about 120 km from the city of Yakutsk (b). Sampling sites (lakes) were visited during four different seasons (Syrdakh village appears in the background) (c).

36 50-200 m above sea level (asl), dated 14-22 kyr BP, and the higher Abalakh terrace in the eastern sector of the study area, 38 200-280 m asl, dated 45-56 kyr BP (Soloviev 1959, 1973; 39 Ivanov 1984). This region is characterized as a middle taiga 40 landscape regime (Fedorov et al. 2014a) and is dominated by larch, pine and birch forests (Ulrich et al. 2017a). Grasslands 42 are abundant in unforested areas, such as land previously 43 cleared for farming, ranching or in the remnant depressions of 44 old thaw lakes known as "alases" (see below). They consist of 45 halophytic steppe-like and bog plant communities (Ulrich 46 et al. 2017b).

Permafrost in this region is continuous (Fig. 1a), thick 48 (> 500 m deep), and the upper 30-50 m (Pleistocene-age flu-49 vial and eolian sediments called "Yedoma") can be extremely 50 rich in ground ice (50-90% by volume) (Ivanov 1984). The 51 amount of organic carbon stored in Yedoma varies widely. For 52 example, deep cores from Northern Siberia and Alaska yielded  $\mathbf{m}$  organic carbon pool estimates of approximately 10 + 754 -6 kg m<sup>-3</sup> (Strauss et al. 2015). A 22 m deep core in Central

Yakutia (Yukechi) on the Abalakh terrace yielded a much 90 lower value of organic carbon content of  $\sim 5 \text{ kg m}^{-3}$ (Windirsch et al. 2020), while another Central Yakutian study 92 (Spasskaya Pad/Neleger site) of a shallow core (2 m) showed a 93 notably higher organic carbon content of 19 kg m<sup>-3</sup> for the 94 top two meters of larch forest-covered Yedoma deposits 95 (Siewert et al. 2015).

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Our study site is underlain by Yedoma silty loams, which 97 are common to the Lena-Aldan interfluve, with abundant gro- 98 und ice in the form of 1.5–3 m-wide ice wedges. Active layer 99 depth in the region generally ranges between ~ 1 m below for- 100 ested areas to > 2 m in exposed grassland areas (Desyatkin 101 et al. 2009). Zones of unfrozen ground (or taliks) exist under- 102 neath major rivers and lakes deeper than the ice-cover thick- 103 ness. Nearly half of the landscape has been affected by 104 thermokarst since the early Holocene, resulting in the formation of thousands of partly drained alas depressions 106 (Soloviev 1959; Brouchkov et al. 2004). However, recent 107 thermokarst activity related to natural landscape evolution,

1 increasing air temperatures and/or human-induced landscape 2 modifications (agriculture, clear-cutting, and infrastructure) is 3 also occurring in the region, as shown by the presence of 4 numerous small and young, fast-developing lakes and retro-5 gressive thaw slumps along lake shores (Fedorov et al. 2014b; 6 Séjourné et al. 2015). The landscape is thus highly dynamic; yet the competing driving factors (climate vs. local geomor-8 phology or vegetation development) and the timing of such 9 changes (gradual vs. rapid/threshold) are complex to charac-10 terize and quantify at the regional scale. 12 Lake types To understand the spatial and temporal heterogeneities in

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greenhouse gas concentrations from thaw lakes in the area, it was first necessary to classify the lakes at the study site (Fig. 2). Differentiation was done based on field observations, past radiocarbon dating of lake sediments, geochemical signatures of lake waters, and a multiple-stage development model (Soloviev 1973; Desyatkin et al. 2009). Relevant statistics and properties for all lake types are given in Table 1. Based on the compiled information, the lake types are:

Unconnected alas lakes: residual water bodies located within endoreic or hydrologically closed basins (Desyatkin

et al. 2009). Most of these lakes likely formed during the 55 transition between the Pleistocene and Holocene, approxi- 56 mately 8–10 thousand years ago (Ulrich et al. 2017b). How- 57 ever, the Holocene Thermal Maximum (~ 6700–5000 cal. yr 58 BP) was also a time of prolific formation and continued 59 development of new and existing alas lakes (Biskaborn 60 et al. 2012; Ulrich et al. 2017b). These lakes can be up to a 61 few meters deep but are typically very shallow (1 m deep or 62 less) and are thus generally frozen to the bottom in winter. 63 Despite a lack of liquid water in the winter, these frozen 64 lakes have heat fluxes that can be one order of magnitude 65 higher than those of the surrounding permafrost (Boike 66 et al. 2015). The ancient lake depressions surrounding the 67 small residual lakes of this type can be up to several kilome- 68 ters wide and several meters deep. These alas lakes have 69 already undergone much of the thermokarst processes and 70 very little ground ice typically remains beneath the residual 71 lake. Therefore, the thaw potential and resulting input of 72 stored carbon to these lakes is low compared to recently 73 formed thermokarst lakes (Ulrich et al. 2019). Yakutian peo- 74 ple have used these depressions for agricultural purposes for 75 centuries (Crate et al. 2017).

Connected alas lakes: lakes connected hydrologically to the 77 watershed by streams or rivers. These lakes are consistently

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Fig. 2. Distribution of lake types in a subset of the study area, including unconnected alas lakes (a; blue), recent thermokarst lakes (b; red), and connected alas lakes (c; purple).

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1 **Table 1.** Depths and surface areas of sampled lakes in central Yakutia (eastern Siberia). Limnological point samples indicated by a star. 55 Limnological profile samples indicated by an arrow. Greenhouse gas samples indicated by +. ND = no data available.

	Lake name	Geographical coordinates	Geomorphological terrace	Lake depth (m)	Surface area (ha)	Fall 2018	Winter 2019	Spring 2019	Sumei 2019
Recent thermokarst	Lake 1	62.5575°N 130.9208°E	Tyungyulyu	2.9	0.31	* +	* ↓ +	* +	* ↓ +
	Lake 4	62.5674°N 131.0203°E	Tyungyulyu	ND	1.99	ND	ND	* +	* +
	Lake 7	62.5452°N 130.9683°E	Tyungyulyu	2.9	0.34	*	*↓+	* +	* +
	Lake 8	62.5489°N 130.9240°E	Tyungyulyu	2.0	0.26	*	*↓+	* +	*   +
	Lake 9	62.5591°N 130.9393°E	Tyungyulyu	4.6	0.94	* +	*↓+	* +	* +
	Lake 18	62.5627°N 130.9191°E	Tyungyulyu	2.3	1.26	*+	*↓+	* +	* +
	Lake 34	62.6112°N 131.1173°E	Abalakh	1.2	0.32	*+	*↓+	* +	* +
	Lake 35	62.6192°N 131.1472°E	Abalakh	3.5	0.37	*+	+ ↓ الأخب	* +	* +
	Lake 37	62.55492°N 130.917°E	Tyungyulyu	1.5	0.14	* +	*↓+	* +	* +
	Lake 38	62.5453°N 130.9684°E	Tyungyulyu	2.2	0.27	*	*↓+	* +	* +
	Lake 39	62.5485°N 130.9178°E	Tyungyulyu	1.5	0.13	*	*↓+	* +	* +
	Lake 40	62.5535°N 130.9175°E	Tyungyulyu	ND	0.17	,	ND	* +	* +
	Lake 41	62.5748°N 130.8580°E	Tyungyulyu	ND	0.37	*	ND	* +	* +
	Lake 42	62.5746°N 130.8547°E	Tyungyulyu	ND	0.18	*	ND	* +	* +
	Lake 43	62.5647°N 130.9033°E	Tyungyulyu	3.6	0.31	*	*↓+	* +	* +
	Mean			2.6	0.49				
Connected alas	Lake 2	62.5598°N 130.9655°E	Tyungyulyu	8.8	64.64	* ↓ +	* ↓ +	* +	* ↓ +
	Lake 11	62.5909°N 131.1871°E	Abalakh	4.9	59.35	* +	*↓+	* +	* +
	Lake 36	62.5631°N 130.8998°E	Tyungyulyu	3.6	179.46	* +	*↓+	* +	* +
	Mean			5.7	101.15				
Unconnected alas	Lake 3	62.5507°N 130.9236°E	Tyungyulyu	2.0	1.68	* +	* ↓ +	* +	* +
	Lake 5	62.5615°N 131.0110°E	Tyungyulyu	ND	4.54	ND	ND	* +	* +
	Lake 10	62.5507°N 130.9226°E	Tyungyulyu	ND	0.88	* +	ND	* +	* +
	Lake 12	62.5844°N 131.1910°E	Abalakh	1.4	40.72	* +	*   +	* +	* +
	Lake 19	62.5585°N 130.9510°E	Tyungyulyu	ND	2.22	ND	ND	* +	* +
	Lake 20	62.5702°N 130.917°E	Tyungyulyu	ND	5.35	ND	ND	* +	* +
	Lake 44	62.5660°N 130.9012°E	Tyungyulyu	3.1	1.93	*	*↓+	* +	* +
	Lake 45	62.5753°N 130.9142°E	Tyungyulyu	ND	0.54	*	ND	* +	* +
	Lake 46	62.5625°N 130.9269°E	Tyungyulyu	ND	1.81	ND	ND	* +	ND
	Lake 47	62.5467°N 130.9660°E	Tyungyulyu	ND	3.32	ND	ND	* +	* +
	Lake 48	62.5663°N 130.9401°E	Tyungyulyu	ND	3.07	ND	ND	* +	ND
	Lake 49	62.5478°N 130.9587°E	Tyungyulyu	ND	2.44	ND	ND	*	* +
	Lake 50	62.5618°N 130.9859°E	Tyungyulyu	ND	1.35	ND	ND	*	*
	Lake 100	62.5864°N 131.0380°E	Tyungyulyu	ND	4.54	ND	ND	ND	* +
	Lake 101	62.5510°N 130.9347°E	Tyungyulyu	ND	4.55	ND	ND	* +	* +

larger (several hundreds of meters across) and deeper (up to ~ 10 m). Most of them were probably formed during the late Holocene, approximately 5-3.5 thousand years ago, although detailed chronology about their inception is still incomplete (Soloviev 1973; Ulrich et al. 2017b). Local people currently use some of these lakes for fishing (e.g., Syrdakh Lake; Fig. 1c).

Recent thermokarst lakes: thaw lakes formed over the last several decades mostly from human activities (e.g., forest fire and forest removal for agriculture, pipelines, or road construction) and rising temperature (Fedorov et al. 2014b). These lakes are generally small (meters to tens of meters across) and relatively shallow (generally one to two meters deep), and are still expanding downwards and laterally due to active layer deepening and thermokarst processes. Their waters can be notably rich in dissolved organic carbon

(DOC), with concentrations typically ranging from ~ 100 to 93  $300 \text{ mg L}^{-1}$  (Hughes-Allen et al. 2020).

#### Methods

#### Physicochemical characteristics of lake water

In 2018–2019, data were collected during four field cam- 99 paigns: September 2018 (fall), March–April 2019 (winter), May 100 2019 (spring), and August 2019 (summer) (Fig. 1; Table 1; 101 Supp. Fig. S1). Physicochemical measurements were con- 102 ducted during each of the four field campaigns using a YSI Pro 103 DSS multiprobe sensor, including profiles in temperature 104 (accuracy of  $\pm$  0.2°C), dissolved oxygen (accuracy of  $\pm$  1% of 105 reading or 1% saturation, whichever is greater), specific con- 106 ductivity (accuracy of  $\pm$  1% of reading) and pH (accuracy of 107  $\pm$  0.2 pH units). Vertical profiles of the water column were 108

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1 obtained from 17 lakes during the winter through a hole 2 drilled in the ice cover using a hand-held auger, and from 3 3 lakes during the summer from an inflatable boat (Table 1). 4 All other data were collected from the lake surface as discrete 5 point measurements, mostly from the lakeshores.

#### Dissolved greenhouse gas measurements

In spring, summer, and autumn, the water was mostly sampled from the lakeshore at the surface. Summer samples were 10 also collected from the pelagic section of the lake at the surface and bottom using a horizontal van Dorn sampler (integrating 12 15 cm depth) for recent thermokarst Lake 1 only. In winter, sam-13 ples were collected from the center of each lake through the ice 14 cover using a vertical van Dorn sampler (integrating ~ 40 cm 15 depth), including a few supplementary samples taken from bot-16 tom waters. Water was quickly transferred into a 2-liter gas 17 exchange LDPE bottle (if not directly sampled from this bottle), 18 and a headspace of 30 mL of atmospheric air was created (bottle 19 with an inlet at the bottom, connected to a syringe). The bottle 20 was shaken vigorously for 3 min, and 20 mL of the gaseous headspace was then transferred to a 12 mL pre-vacuumed gastight Exetainer glass vial previously flushed with nitrogen (Hesslein et al. 1991). Samples were later analyzed by gas chro-24 matography (Thermo Trace 1310, TRI-Plus Head-Space auto-sampler, HSQ  $80/100 \, 1.6 \, \text{mm} \times 30.5 \, \text{m}$  in series with MS 5A26 1.6 mm × 18.3 m, thermal conductivity and flame ionization detectors). The dissolved greenhouse gas concentration  $(C_{sur})$ was calculated using Henry's Law, and departure from saturation was obtained subtracting the gas concentration in the water at equilibrium with the atmosphere ( $C_{eq}$ ) using global annual average values of 410 ppm for CO<sub>2</sub> and 2 ppm for CH<sub>4</sub> (IPCC 2019).

Diffusive fluxes of dissolved greenhouse gas were estimated 33 as in Bouchard et al. (2015) based on a gas transfer coefficient 34 (k<sub>600</sub>) standardized to a Schmidt number (Sc) of 600 (Wanninkhof 1992), and calculated with the wind-based model of Vachon and Prairie (2013):

$$k_{600} = 2.51 + 1.48 U_{10} + 0.39 U_{10} \log_{10} \text{LA}$$

where  $U_{10}$  is the wind speed at 10 m above the ground from the closest meteorological station (at Yakutsk) and LA is the lake surface area (in  $km^2$ ). The flux (in mmol  $m^{-2}$   $d^{-1}$ ) was then calculated by applying the equation:

$$Flux = k \left( C_{sur} - C_{eq} \right)$$

where k is the transfer coefficient for a given gas calculated as:

$$k = k_{600} \left( \frac{Sc}{600} \right)^{-0.5}$$

Due to generally high pH (>9) and low wind turbulence 53 during the ice-free seasons, chemical enhancement could fur-54 ther increase the high levels of CO<sub>2</sub> uptake estimated for some

lakes (Wanninkhof and Knox 1996). The exchange rate of 55 gases from the atmosphere through the boundary layer of a 56 liquid surface is regulated by molecular diffusion and acceler- 57 ated by increasing turbulence (wind). However, the exchange 58 rate of CO<sub>2</sub> molecules from the atmosphere through the 59 boundary layer of a lake surface is also regulated by water temperature, pH and ionic strength. When favorable conditions 61 are met (high pH and temperature, low winds), diffusion 62 across the boundary layer can be enhanced by the hydration 63 of CO<sub>2</sub> to HCO<sub>3</sub><sup>-</sup>. The CO<sub>2</sub> uptake is therefore a function of 64 both diffusion and chemical reaction. The chemical enhance- 65 ment factor for  $CO_2$  ( $\alpha$ ) is defined as:

$$\alpha = k_{\rm enh}/k$$

where  $k_{enh}$  is the enhanced gas transfer velocity, and k is the gas transfer velocity for a nonreactive gas with the same diffusion coefficient as CO<sub>2</sub> as calculated above (Wanninkhof and Knox 1996).

#### Statistical analysis

The dissolved greenhouse gas data were not found to be normally distributed and did not meet the assumption of homoscedasticity based on the Shapiro-Wilk normality test. Typical data transformations to improve normality, such as a Box-Cox and logarithmic transformations, were attempted with unsuccessful results. The small sample sizes likely impede the ability to achieve normality, and we did not expect the data to be inherently normal. Therefore, we used Kruskal-Wallis one-way ANOVA to test the relationships between dissolved CO2 and CH4 concentrations vs. lake types and seasons. Significance between groups was tested for dissolved CO2 and CH4 concentrations as follows: the values for dissolved greenhouse gas concentrations for each lake type were compared across seasons (e.g., dissolved CO2 concentrations for unconnected alas lakes between fall, winter, spring, and summer) and between lake types for a specific season (e.g., dissolved CO<sub>2</sub> concentrations for unconnected alas lakes, connected alas lakes, and recent thermokarst lakes during the fall season). The null hypothesis that the medians of the compared groups are equal was rejected if p < 0.05. All of the statistical analysis was completed using the Python programming language (Python Software Foundation, https://www.python. org/).

#### Results

#### Seasonal conditions

Sampling began in September 2018, at the end of the icefree season when the depth of the active layer is at its maximum. During this period, nighttime air temperatures drop near 0°C, initiating lake water mixing (fall overturn). Although limnological profiles were not done during the fall sampling season, water surface temperatures (ranging from 4.7 to 12.5°C depending on the lake; Hughes-Allen et al. 2020)

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1 indicate that some lakes may have been sampled before the 2 water column had completely mixed to bottom. During the 3 winter, lakes were covered by 0.7–1.2 m of ice and 30–50 cm snow (measured in March-April 2019). Shallow lakes, 5 mostly unconnected alas lakes, were frozen completely to the 6 bottom during the winter, whereas conditions were met for 7 inverse stratification in deeper lakes (defined as the water bod-8 ies keeping liquid water during winter, i.e., of a maximal 9 depth > 1.2 m). Sampling occurred late in winter (< 3 April), 10 but well before the thaw season, which usually begins in May 11 in this region. Many lakes (larger and/or deeper) were still cov-12 ered in ice at the beginning of May when the site was visited 13 for the spring sampling, but the ice cover melted progressively 14 until surface water reached 4°C and spring overturn started. 15 The surface temperature of most lakes was already above 4°C 16 at the time of sampling in spring (ranging from 2.8°C to 17 17.5°C; Hughes-Allen et al. 2020), with the smaller uncon-18 nected alas lakes showing consistently higher surface tempera-19 tures (mean 10.9°C) than the connected alas lakes (5.4°C) or 20 recent thermokarst lakes (6.8°C). Therefore, spring sampling 21 likely occurred after the initiation of the thermal stratification. 22 The warming of surface waters during the summer initiated 23 the thermal and chemical stratification of the water column 24 that may have lasted throughout the summer for deeper lakes.

#### 26 Physicochemical characterization of lake water 27 Broad trends and seasonal averages

The hydrochemical characteristics of the sampled lakes dif-29 fered notably between lake types and seasons (surface values presented in Fig. 3). Regardless of the season, unconnected 31 alas lakes generally had the highest electrical conductivity. 32 This inter-lake difference was particularly marked during the 33 winter, when the mean conductivity value for unconnected 34 alas lakes was  $5711 \,\mu\text{S cm}^{-1}$ . The conductivity for uncon-35 nected alas lake #3 reached 8000  $\mu$ S cm<sup>-1</sup>, four times higher 36 than the average conductivity for connected alas lakes. Con-37 nected alas lakes showed the lowest values (winter mean = 56338  $\mu$ S cm <sup>-1</sup>), with recent thermokarst lakes being intermediate 39 (winter mean = 4228  $\mu$ S cm<sup>-1</sup>) (Fig. 3a–d). As expected, con-40 ductivity reached much lower values (averages < 1000  $\mu$ S cm<sup>-1</sup>) for all lakes in the springtime, during and shortly after 42 ice-cover breakup. Conductivity increased slightly during the 43 summer, reaching mean values similar to the previous fall 44 (~ 1000–2000  $\mu$ S cm<sup>-1</sup>) and showing comparable values 45 between recent thermokarst and unconnected alas lakes.

The percent saturation of dissolved oxygen for all three lake 47 types was lowest in the wintertime, with mean values at lake surface between 0% and 20% (Fig. 3e-h). Unconnected alas 49 lakes maintained the highest levels of dissolved oxygen during 50 fall, spring, and summer compared to the other two lake types (with more pronounced differences in spring), reaching values 52 above 200% in summer (Fig. 3h). Recent thermokarst lakes 53 maintained the lowest concentrations of dissolved oxygen, 54 rarely reaching saturation levels even in the springtime.

Connected alas lakes followed a similar trend as the other two 55 lake types, but with intermediate values.

The pH values exhibited less seasonal variability compared 57 to the other hydrochemical properties analyzed (Fig. 3i-l). 58 Unconnected alas lakes consistently had the highest pH levels 59 relative to the other two lakes types, maintaining an average 60 pH between 9.4 and 10 for all four seasons. Connected alas lakes and recent thermokarst lakes exhibited similar seasonal 62 trends, with average pH values in fall, spring and summer near 63 8.8, and with lower values during the winter season (average 8.7).

#### Seasonal profiles

Although temperature and dissolved oxygen profiles are not available for all lakes for each season, analysis of the profiles for selected representative lakes (unconnected alas lake #3, connected alas lake #2, recent thermokarst lake #1) (see lake locations in Supp. Fig. S1) illustrates some broad seasonal trends. Profiles of temperature and dissolved oxygen show that the deeper lakes (i.e., connected alas and some recent 74 thermokarst lakes) were stratified during winter and summer (Figs. 4, 5). Inverse stratification was observed in winter, with bottom water ranging from near zero (at 2.1 m in unconnected alas #3) to 4°C (at 8.2 m in connected alas #2). During 78 this period, dissolved oxygen saturation levels ranged between  $\sim 10\%$  and 40% in surface waters (0–50 cm), while it decreased 80 to values < 5% in bottom waters (Fig. 4a,c,e). Conductivity 81 and pH profiles showed smaller differences between surface and bottom waters during winter, except for recent thermokarst lake #1 showing a decrease of nearly 2000 µS cm<sup>-1</sup> between surface and bottom (Fig. 4b,d,f). These profiles illustrate the broad range of conductivity (496–8066  $\mu$ S cm<sup>-1</sup>) and pH (7.4-9.4) that can be found between individual lakes 87 during the winter on such a lake-rich landscape.

In August, available temperature and oxygen profiles for recent thermokarst lakes (lakes #1, #8, and #18) show strong stratification (Fig. 5). At the time of sampling (0945–1200), the difference between surface and bottom temperature was 7.1°C (lake #1), 5.8°C (lake #8), and 7.3°C (lake #18). Dissolved oxygen saturation levels ranged between ~ 28% and 94 54% in surface waters and between ~ 4% and 11% in bottom waters (Fig. 5). All three recent thermokarst lakes showed similar profiles in pH and conductivity. Profiles showed decreasing 97 pH values with depth (ranging from approximately 9 at the 98 surface to 7.5 in bottom waters). Conductivity profiles showed an increase with depth from approximately 2500  $\mu$ S cm<sup>-1</sup> at the surface to  $6000 \,\mu\text{S} \,\text{cm}^{-1}$  in bottom waters for all three sampled recent thermokarst lakes.

#### Dissolved greenhouse gas concentrations

Surface concentrations of dissolved greenhouse gas, expressed as departure from saturation, varied strongly between lake types and seasons (Fig. 6). Broadly, unconnected alas lakes and recent thermokarst lakes had stronger seasonal

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1 differences in CO<sub>2</sub> concentrations, while connected alas lakes 2 showed less seasonal difference in CO<sub>2</sub> concentration. Winter 3 CO<sub>2</sub> concentrations for all lake types were generally signifi-4 cantly different from all other seasons. Differences between 5 unconnected alas lakes and recent thermokarst lakes were gen-6 erally significant for most seasons (Table 2; Table 3). Overall, recent thermokarst lakes consistently showed the highest CO<sub>2</sub> 8 concentrations between lakes types, yet values varied consid-9 erably among the 15 lakes sampled and between seasons, with 10 departure from saturation averages ranging from slight CO<sub>2</sub> sink during spring sampling ( $-10 \mu M$ ) to very strong  $CO_2$ 12 source during winter sampling (+ 1440  $\mu$ M).

During the fall sampling period, unconnected alas lakes 14 were under-saturated in CO<sub>2</sub>, and thus acting as CO<sub>2</sub> sinks rel-15 ative to the atmosphere, whereas recent thermokarst lakes 16 were generally super-saturated and therefore sources of CO<sub>2</sub> (Fig. 6a). The CO<sub>2</sub> concentrations in connected alas lakes were close to equilibrium with the atmosphere and thus neither a 55 source nor a sink at time of sampling. A remarkable increase 56 in CO<sub>2</sub> concentrations occurred under the ice cover for all lake 57 types. For example, recent thermokarst lake #34 had 20 times 58 more  $CO_2$  (1295  $\mu$ M) in the winter than in the previous fall 59  $(62 \,\mu\text{M})$  and connected alas lake #11 had 10 times more CO<sub>2</sub> 60 in the winter compared to the previous fall (190  $\mu$ M in winter 61 vs. 27  $\mu$ M in fall) (Fig. 6c). Bottom water concentration of dissolved CO<sub>2</sub> was only sampled in the connected alas lake #2 63 during winter, showing a concentration similar than at the 64 surface (155  $\mu$ M in bottom water vs. 168  $\mu$ M at the surface).

During spring sampling, nearly all lakes (28 out of 33) had 66 CO<sub>2</sub> concentrations below saturation, with relatively low variability (Fig. 6e). During late summer however (August), satura- 68 tion levels at the lake surface returned to similar values as 69 observed during the previous fall (early fall, September); 70 unconnected alas lakes reverted to slight CO<sub>2</sub> sinks, connected 71

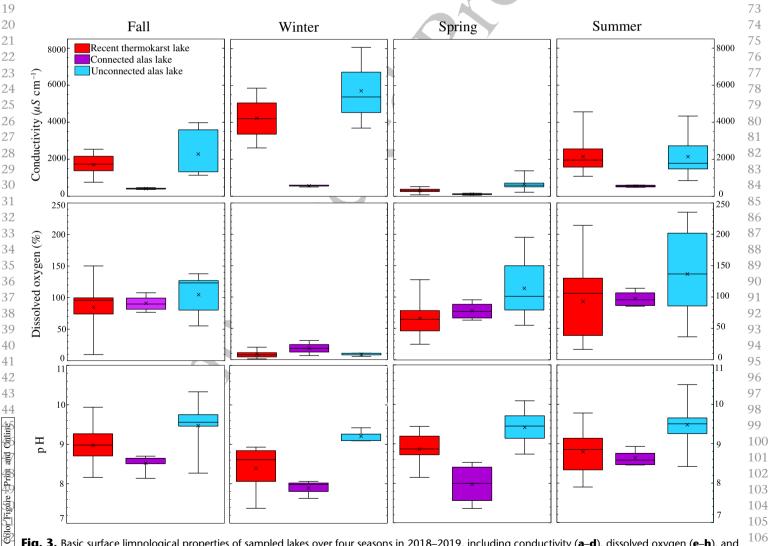


Fig. 3. Basic surface limnological properties of sampled lakes over four seasons in 2018–2019, including conductivity (a-d), dissolved oxygen (e-h), and pH (i-l). Number of lakes sampled for each lake type for each season is presented in Table 1. The box indicates the interguartile range. The horizontal line within the box is the median and the small x is the mean. The whiskers indicate minimum and maximum.

1 alas lakes returned to equilibrium, and recent thermokarst 2 lakes became the largest CO<sub>2</sub> sources again, with one particu-3 lar lake (#35) showing a notably high value of nearly 500 μM 4 (Fig. 6g) (Supp. Fig. S1). During summer sampling, the bottom 5 waters of recent thermokarst lake #1 (Supp. Fig. S1) showed 6 extremely high saturation levels (> 1500  $\mu$ M), nearly two orders of magnitude higher than at the surface, representing a substantial storage of CO<sub>2</sub>. Differences between surface and 9 bottom water were not assessed in other lakes.

General trends in the concentrations of CH<sub>4</sub> showed that 11 recent thermokarst lakes typically had stronger seasonal differ-12 ences in CH<sub>4</sub> concentrations compared to the other two lake 13 types. The summer sampling season showed the highest vari-14 ability in CH<sub>4</sub> concentration between lake types than the other three seasons (Table 2; Table 3).

During fall sampling, unconnected alas lakes were the 17 strongest sources of CH<sub>4</sub>, while the other lake types were small or negligible sources (Fig. 6b). During winter, as for CO<sub>2</sub>, dis-19 solved CH<sub>4</sub> showed a large increase in concentration for 20 recent thermokarst and connected alas lakes (by one order of magnitude) compared to the previous fall (Fig. 6d). Bottom 22 water concentration of dissolved CH<sub>4</sub> was only sampled for 23 connected alas lake #2. The bottom water concentration was 24 one order of magnitude higher than surface concentration 25 during the winter sampling period (133  $\mu$ M in bottom water 26 vs. 15  $\mu$ M at the surface).

During spring sampling, all lakes (excluding a small num-28 ber of outliers) were moderately saturated in CH<sub>4</sub> compared to 29 the values observed during winter sampling (nonoutliers  $30 < 60 \,\mu\text{M}$ ), with similar saturation levels between lake types (Fig. 6f). During summer sampling, observed CH<sub>4</sub> saturation 32 levels at the surface were lower ( $\sim 10 \mu M$ ), with unconnected 33 alas lakes as the largest CH<sub>4</sub> sources compared to the other 34 lake types, which is a similar trend to the previous fall 35 (Fig. 6h). As observed for CO<sub>2</sub> during summer, recent 36 thermokarst lake #35 showed notably higher surface CH<sub>4</sub> con-37 centrations (saturation levels > 700  $\mu$ M, identified by an arrow 38 in Fig. 6h) compared to all other sampled lakes. Bottom waters 39 of recent thermokarst lake #1 (Supp. Fig. S1) showed 40 extremely high saturation levels during summer sampling  $(>40 \mu M)$ , nearly 130 times higher than at the surface 42 ( $\sim 0.4 \mu M$ ), representing a substantial storage of CH<sub>4</sub>. Differ-43 ences between surface and bottom water were not assessed in 44 other lakes.

### 46 Diffusive greenhouse gas fluxes

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Considering only the ice-free season (from spring to fall; 47 Table 4), the diffusive CO<sub>2</sub> fluxes ranged from a minimum of  $49 - 13.6 \text{ mmol m}^{-2} \text{ d}^{-1}$  (or approximately  $-600 \text{ mg C m}^{-2} \text{ d}^{-1}$ ) 50 recorded in spring for unconnected alas lake #3 to a maximum 51 of  $\sim 355 \text{ mmol m}^{-2} \text{ d}^{-1}$  (or  $\sim 1200 \text{ mg C m}^{-2} \text{ d}^{-1}$ ) recorded in 52 summer for recent thermokarst lake #18. The ice-free season 53 median CO<sub>2</sub> fluxes were relatively low for connected alas lakes negative for unconnected alas lakes

- 4.8 mmol m<sup>-2</sup> d<sup>-1</sup>, respectively), and slightly positive for 55 recent thermokarst lakes (1.8 mmol m<sup>-2</sup> d<sup>-1</sup>). The diffusive 56 CH<sub>4</sub> fluxes were highest from recent thermokarst lakes in sum- 57 mer, reaching  $\sim 560 \text{ mmol m}^{-2} \text{ d}^{-1}$  (or 6720 mg C m<sup>-2</sup> d<sup>-1</sup>). It 58 is noteworthy that the average ice-free season CH<sub>4</sub> flux for all 59 studied lakes was substantially higher than the median value 60  $(14 \text{ mmol m}^{-2} \text{ d}^{-1} \text{ compared to } 3.1 \text{ mmol m}^{-2} \text{ d}^{1})$ , reflecting 61 the skewing effect of exceptionally high fluxes in summer 62 from a few recent thermokarst lakes. Unconnected alas lakes 63 were regularly large CH<sub>4</sub> sources during the ice-free season 64 (median flux of  $\sim 7.0 \text{ mmol m}^{-2} \text{ d}^{-1}$ , compared to 1.0 mmol m<sup>-2</sup> d<sup>-1</sup> for recent thermokarst lakes). It should be 66 noted that these results are estimations based on wind data from the only available meteorological station in the region (Yakutsk Airport), located ~ 120 km from our study site.

It is possible that the chemical enhancement effect contributed to increased CO<sub>2</sub> uptake rates by some lakes during the 71 ice-free seasons. Unconnected alas lakes consistently had the 72 highest α values (mean value of 9.9 for unconnected alas 73 lakes, reaching up to 27 for lake #20 in August). These large 74 values occurred when pH was particularly high, lake water was 75 warm and wind speeds were low. Alpha values calculated for 76 the other two lakes presenting a CO<sub>2</sub> uptake during August 77 were lower (connected alas lake #11  $\alpha$  = 1.6; recent 78 thermokarst lake #37  $\alpha$  = 5.3). During September, there were 79 only two connected alas lakes and two recent thermokarst 80 lakes that were acting as  $CO_2$  sinks (mean  $\alpha$  1.4), while the 81 unconnected alas lake #10 presented the highest  $\alpha$  value (3.8). 82 The spring sampling season showed low variability in the calculated  $\alpha$  values between lakes types, although unconnected 84 alas lakes still had the highest mean value (2.8, compared to 85 mean values ~ 1.0 for connected alas lakes and recent 86 thermokarst lakes). It is important to note that during the 87 spring season, most lakes (26 out of 31) were acting as CO<sub>2</sub> sinks and the magnitude of CO<sub>2</sub> uptake was highest compared 89 to the other seasons (Fig. 6). However, the highest chemical 90 enhancement of CO2 flux was calculated for the summer season, when CO<sub>2</sub> uptake is less prevalent and smaller in magnitude (Fig. 6).

#### Discussion

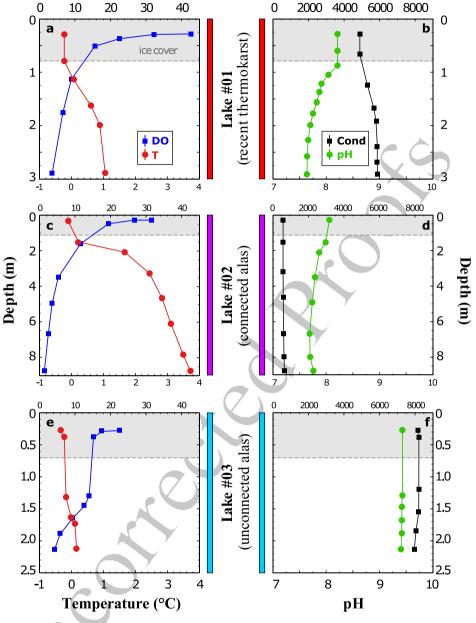
The study area represents the typical taiga alas landscape found in Central Yakutia, a region encompassing approximately 17,000 km<sup>2</sup> in the Lena-Aldan interfluve (Ulrich et al. 2017a). Nearly 10 % of this landscape is comprised of lakes at varying stages of thermokarst development (Ulrich et al. 2017a). Lake morphology is directly linked to the evolution stage within the thermokarst sequence (Soloviev 1973), 103 with rates of change substantially lower for older lakes (Desyatkin et al. 2009). The morphology of unconnected alas lakes has remained relatively stable during the last decades, 106 with little lateral expansion or deepening. In contrast, recent 107 thermokarst lakes have been notably dynamic, experiencing

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Conductivity ( $\mu S \text{ cm}^{-1}$ )



Dissolved oxygen (%)

**Fig. 4.** Winter limnological profiles (temperature, dissolved oxygen, conductivity, and pH) in recent thermokarst lake #1 (**a** and **b**), connected alas lake #2 (**c** and **d**), and unconnected alas lake #3 (**e** and **f**).

subsidence and lateral expansion consistent with the rapid landscape evolution observed since the mid-20th century (Fedorov et al. 2014a; Ulrich et al. 2017a). Connected alas lakes, while they did not notably expand or deepen in the recent past, have shown some signs of thermokarst activity such as thaw slumps (Séjourné et al. 2015). This spatial heterogeneity resulting from the above chronological sequence can contribute to notable differences in lake water physicochemical characteristics, such as timing of spring ice-cover breakup, water temperature and lake water mixing regime, electrical conductivity, and dissolved oxygen concentration.

Depending on the lake type, we also found substantial seasonal variations in the amount of dissolved greenhouse gas in lake water, spanning two orders of magnitude throughout the year. Since most greenhouse gas studies are limited to summer quantification (e.g., Abnizova et al. 2012; Bouchard et al. 2015), such seasonal assessments are precious (Langer et al. 2015; Matveev et al. 2019). The wide seasonal ranges observed here underscore the need to consider such heterogeneities when attempting to estimate global greenhouse gas emissions from thermokarst lakes and permafrost landscapes.



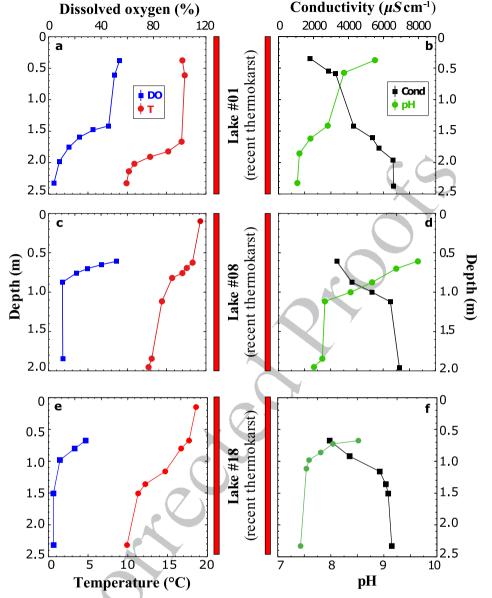
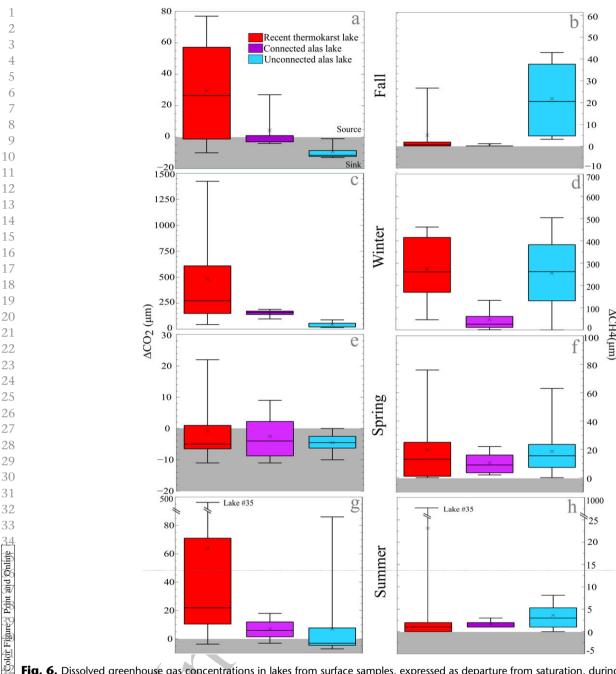


Fig. 5. Summer limnological profiles (temperature, dissolved oxygen, conductivity, and pH) in recent thermokarst lakes #1 (a and b), #8 (c and d), and #18 (e and f).

#### 43 Developmental stage as a driving factor on lake greenhouse gas concentrations and fluxes

We measured notable differences in physicochemical prop-46 erties (Fig. 3) and dissolved greenhouse gas concentrations (Fig. 6) between thermokarst lakes of different development stages. The clear (statistically significant) difference between 49 unconnected alas lakes and the other two types of lakes under-50 lines the unique role played by these shallow lakes, acting as CO<sub>2</sub> sinks most of the year and substantial CH<sub>4</sub> sources, espe-52 cially in fall. Several studies have concluded that unconnected 53 alas lakes are generally not sites of current thermokarst activity 54 (e.g., Brouchkov et al. 2004; Ulrich et al. 2017a), which is

consistent with our observations. Little ground ice remains 97 below alas lake depressions, which significantly reduces the potential for input of stored carbon from surrounding perma- 99 frost (Ulrich et al. 2019). They are stable both laterally and 100 vertically, and any variations in lake volume can be attributed 101 to changes in the hydrological regime. As they are hydrologi- 102 cally isolated, the interannual water balance in these lakes is 103 entirely dependent upon the ratio between annual precipita- 104 tion and evaporation (Brouchkov et al. 2004; Ulrich 105 et al. 2017a). Based on these characteristics, evaporative 106 enrichment during summer and substantial solute exclusion 107 under winter ice cover are likely to play a role in the high 108



**Fig. 6.** Dissolved greenhouse gas concentrations in lakes from surface samples, expressed as departure from saturation, during fall 2018 (**a** and **b**), winter 2019 (**c** and **d**), spring 2019 (**e** and **f**), and summer 2019 (**g** and **h**). Scales for  $CO_2$  concentrations appear on the left, and for  $CH_4$  concentrations on the right. Number of lakes sampled for each lake type for each season is presented in Table 1. The box indicates the interquartile range. The horizontal line within the box is the median and the small x is the mean. The whiskers indicate minimum and maximum.

 $^{48}$  concentration of nutrients and minerals observed within these  $^{49}$  lakes, such as found in Western Siberia (Manasypov  $^{50}$  et al. 2015). The notably high organic carbon concentrations  $^{51}$  (generally DOC >  $^{100}$  mg  $^{1}$  except in spring) are characteristic of these lakes (Hughes-Allen et al. 2020).

In the open-water seasons, unconnected alas lakes are hot-54 spots of biological activity, sometimes resulting in the entire surface of the lake being covered by floating aquatic plants 102 (e.g., *Lemna* sp.) (Malyschez and Peschkova 2001). These lakes 103 are particularly shallow (< 2 m), generally warmer than the 104 other two lake types, and rich in nutrients. These characteristics favor high rates of photosynthetic activity, causing unconnected alas lakes to be efficient CO<sub>2</sub> sinks (or negligible CO<sub>2</sub> 107 sources) during the open-water season. Negative fluxes are 108

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1 **Table 2.** Results of the Kruskal–Wallis significance tests which compared dissolved greenhouse gas concentrations for each lake type 55 2 between seasons. The results for CO<sub>2</sub> and CH<sub>4</sub> are grouped by lake type. The left-most and right-most columns indicate which seasons are being compared (e.g., winter  $CO_2$  was larger than fall  $CO_2$  for unconnected alas lakes). Significance (p < 0.05) is indicated by > or < signs, while = signs indicate that differences are not significant.

5 6	Unconnect	ed alas lake	Connected alas lake		Recent therm	okarst lake	:
7	CO <sub>2</sub>	CH₄	CO <sub>2</sub>	CH₄	CO <sub>2</sub>	CH₄	
Winter	>	=	>	>	>	>	Fall
Winter	>	=	>	=	>	>	Spring
Winter	>	=	>	=	>	>	Summer
Spring	=	=	=	>	=	>	Fall
<sup>2</sup> Spring	<	>	=	=	<	) =	Summer
<sup>3</sup> Summer	=	=	=	=	>	>	Fall

most prominent during the spring and fall seasons, probably linked to the more turbulent conditions over these periods leading to an efficient CO<sub>2</sub> transfer from the atmosphere to the lake water. It should be noted that the CO<sub>2</sub> uptake rates for these lakes might be underestimated. Chemical enhancement calculations indicate that CO<sub>2</sub> uptake could be much <sup>22</sup> higher during the summer (on average by a factor of 10) when 23 conditions are met for this effect to be significant (high pH, high temperature, and low winds). These lakes are presenting the highest pH values (generally above 9 and up to 10.5).

Unconnected alas lakes also act as CH<sub>4</sub> sources throughout <sup>27</sup> the year, particularly in fall with flux reaching ~ 16 mmol m <sup>28</sup> d<sup>-1</sup> (Table 4). These fluxes are generally much higher than <sup>29</sup> what has been reported elsewhere across the Arctic (e.g., Abnizova et al. 2012; Bouchard et al. 2015; Matveev  $^{31}$  et al. 2016). High production of CH<sub>4</sub> throughout the open-32 water season is likely linked to the high levels of primary pro-33 ductivity fueling microbes with recent, labile organic carbon 34 to the system. Recently produced, autochthonous organic car-35 bon may be more readily decomposed under anoxic condi-<sup>36</sup> tions than organic carbon from a terrestrial source (Grasset et al. 2018). Consequently, we hypothesize that CH<sub>4</sub> produc-<sup>38</sup> tion is positively linked to the presence of autochthonous organic carbon in these lakes.

Unconnected alas lakes are shallow water bodies less strongly stratified during winter compared to the other lake

types (deeper and more humic). Although limnological profiles were not collected during summer for unconnected alas lakes, it is likely that they are not as stratified as other lake types in summer as well due to their very shallow depth and susceptibility to wind-induced mixing. These conditions likely contribute to an earlier release of the greenhouse gas stored in the water column or sediments in spring and during summer. The mean surface temperature of unconnected alas lakes in May was 11°C compared to 5°C (connected alas lakes) and 7°C (recent thermokarst lakes), indicating that summer stratification started earlier for these lakes (Hughes-Allen et al. 2020) despite they are likely polymictic lakes. It is possible that they were presenting positive CO<sub>2</sub> flux and higher CH<sub>4</sub> flux earlier in spring, but since most unconnected alas lakes freeze to bottom in winter, storage flux would be mainly through greenhouse gas stored in the ice itself, or in the interstitial water of the sediment. In summer, periods of wind-induced mixing would likely generate regular venting of the greenhouse gas produced in summer by these shallow lakes, although the very high CH<sub>4</sub> concentrations observed in fall indicate either some release of CH<sub>4</sub> stored at the bottom or very high production rates during this period. Other studies on thermokarst lakes have documented that the timing and magnitude of greenhouse gas storage flux can vary between lakes even when they are subject to the same local climate conditions (Matveev et al. 2019).

Table 3. Results of the Kruskal-Wallis significance tests which compared dissolved greenhouse gas concentrations of the three lake types during one season. The leftmost and rightmost columns indicate the lake types being compared. Significance (p < 0.05) is indi-46 cated by either > or < signs, depending on which lake type has the highest dissolved greenhouse gas concentration for that particular season, while = signs indicate that differences are not significant.

)	Fall 2018		Winter 2019		Spring 2019		Summer 2019		
	CO <sub>2</sub>	CH₄	CO <sub>2</sub>	CH₄	CO <sub>2</sub>	CH <sub>4</sub>	CO <sub>2</sub>	CH₄	
Unconnected alas lake	=	>	<	=	=	=	=	=	Connected alas lake
Unconnected alas lake	=	=	<	=	<	=	<	<	Recent thermokarst lake
Connected alas lake	=	=	=	<	=	=	<	<	Recent thermokarst lake

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1 **Table 4.** Diffusive fluxes (in mmol m<sup>-2</sup> d<sup>-1</sup>) of CO<sub>2</sub> and CH<sub>4</sub> over four sampling seasons in 2018–2019 in lakes of different develop- 55mental stages (age) in central Yakutia (eastern Siberia). N is the number of lakes sampled in each category.

1		CO <sub>2</sub> flux (m	$mol m^{-2} d^{-1}$	)	CH <sub>4</sub> flux (mmol m <sup>-2</sup> d <sup>-1</sup> )				
Туре	N	Min	Med	Mean	Max	Min	Med	Mean	Max
Fall 2018									
Recent thermokarst	6	-5.7	19.8	17.7	39.8	0.1	0.3	3.9	21.7
Connected alas	4	-4.0	1.5	6.7	27.6	0.2	0.4	0.6	1.4
Unconnected alas	4	-10.6	-7.3	-6.4	-0.7	2.6	12.0	12.4	22.9
All lake types	14	-10.6	0.4	7.7	39.8	0.1	0.6	5.4	22.9
Spring 2019									
Recent thermokarst	23	-9.4	-3.3	-1.5	21.7	0.2	6.6	12.7	74.9
Connected alas	5	-6.9	-0.1	0.3	8.6	4.0	4.7	8.5	21.0
Unconnected alas	17	-13.6	-4.9	-5.2	-0.2	0.2	11.6	15.0	56.8
All lake types	45	-13.6	-3.6	-2.7	21.7	0.2	8.2	13.1	74.9
Summer 2019									
Recent thermokarst	16	-3.3	20.8	51.4	355.4	0.3	0.6	36.2	566.4
Connected alas	4	-3.7	15.5	13.6	27.0	0.3	1.1	1.5	3.2
Unconnected alas	12	-7.4	-3.2	8.2	98.1	0.2	3.8	3.9	9.5
All lake types	32	-7.4	14.5	30.5	355.4	0.2	1.1	19.8	566.4
lce-free seasons (2018–20	19)								
Recent thermokarst	45	-9.4	1.8	19.9	355.4	0.1	1.1	19.9	566.4
Connected alas	13	-6.9	1.5	6.4	27.6	0.2	1.4	3.9	21.0
Unconnected alas	33	-13.6	-4.8	-0.5	98.1	0.2	6.9	10.6	56.8
All lake types	91	-13.6	-2.3	10.6	355.4	0.1	3.1	14.2	566.4

Between the three lake types, recent thermokarst lakes generally showed the highest variability in dissolved greenhouse gas concentrations among lakes of this type (Fig. 6). This potentially results from the substantial diversity in morphology and thermokarst developmental stage for this lake type, with some lakes being much deeper and larger than others (depths from 1.2 to 4.6 m, surface areas from 0.1 to 2.0 ha) (Table 1). Morphology has been suggested to affect the mixing regime, and thus the seasonal patterns in greenhouse gas production vs. consumption, storage and emissions (Matveev et al. 2019). For example, the spring overturn of shallower lakes may represent a more complete outgassing of all greenhouse gas stored under the ice in winter, while deeper lakes may only experience outgassing of the greenhouse gas stored at the lake surface, with deeper storage only venting out in the fall (Matveev et al. 2019). Morphology as thus been identified as a major characteristic controlling seasonal emission patterns, and deserving attention in future studies.

High concentrations of both CO<sub>2</sub> and CH<sub>4</sub> were measured in recent thermokarst lakes, especially under the ice cover (in winter) but also during spring and summer in some cases (Fig. 6). Notably, thermokarst lake #35 (Supp. Fig. S1), currently developing within a forested ground and away from any apparent human influence, had the highest surface concentrations observed in summer (> 450 µM of CO<sub>2</sub> and ~ 1000  $\mu$ M of CH<sub>4</sub>) (Fig. 6g,h). Although a limnological profile

is not available for this lake, strong summer stratification is expected such as is observed in other recent thermokarst lakes (Fig. 5). Finding such high concentrations of CO<sub>2</sub> and CH<sub>4</sub> at the surface of a strongly stratified lake would indicate that greenhouse gas production was elevated in the pelagic portion of the lake, or that lateral transfer of littoral benthic production was efficient. It is also possible that this specific sampling day was following a partial mixing event bringing greenhouse gas enriched waters at the surface. The wind speed during the 2 d prior to greenhouse gas sampling reached sustained speeds up to 10 m s<sup>-1</sup>, which is much higher than typical wind speeds in this area, usually near 1 m s<sup>-1</sup> (Yakutsk Airport Metrological Station). These conditions could have contributed to partial mixing of the lake waters and the observed high concentrations of greenhouse gas in the surface waters. Overall, recent thermokarst lakes were particularly high CO<sub>2</sub> emitters in summer and fall (Fig. 6), and lower CH<sub>4</sub> emitters than unconnected alas lakes (particularly in fall). In fact, CO<sub>2</sub>/CH<sub>4</sub> molar ratios stayed relatively high over the summer and fall in recent thermokarst lakes (> 40) (Table 5), potentially indicating efficient methanotrophy in these systems (Matveev et al. 2019). It is also possible that such elevated ratios were caused by a lateral transfer of CO<sub>2</sub> produced in soils under aerobic conditions (Campeau et al. 2018) or from the chemical weathering of carbonate (Zolkos et al. 2018), sources that would be particularly significant under active thermokarst erosion.

The greenhouse gas production in recent thermokarst lakes 2 is likely fueled by the thawing of ice-rich Pleistocene perma-3 frost. Recent thermokarst lakes experience high rates of lateral 4 and vertical expansion (Fedorov et al. 2014a; Ulrich 5 et al. 2017a), and once initiated, this expansion occurs rela-6 tively continuously. However, the rate of expansion is strongly dependent on vegetation cover type, the presence of 8 disturbed landscape surface (i.e., clearing of land for agricul-9 ture or infrastructure), and local permafrost conditions (Lara 10 et al. 2015; Ulrich et al. 2017a). The expansion of these lakes 11 into surrounding permafrost results in the mobilization of 12 organic and mineral matter which was sequestered for thou-13 sands of years by freezing temperatures (Hugelius et al. 2014; 14 Strauss et al. 2017). For example, DOC values ranging from 15 8 to 838 mg L<sup>-1</sup> were measured in 2018–2019 (Hughes-Allen 16 et al. 2020). Previous work in the region suggests that old car-17 bon is released within these lakes in the form of century- to 18 millennia-old dissolved inorganic carbon, which is resulting 19 from variable mixtures of late Pleistocene (20 kyr BP) and 20 more recent or modern carbon (Soloviev 1959, 1973; 21 Ivanov 1984). Radiocarbon dating and stable isotopic signa-22 tures of gases, lake sediments and surrounding soils should 23 provide more insight about the sources and pathways of the 24 greenhouse gas emitted from recent thermokarst lakes of this 25 region (Bouchard et al. 2015).

#### 27 Seasonal variations in greenhouse gas concentrations

Seasonal differences in dissolved greenhouse gas concentra-29 tions were particularly striking at the studied site. The most notable seasonal change was the substantial increase (up to two orders of magnitude) in both CO2 and CH4 during the 32 winter season (Fig. 6). Increasing concentrations under the ice 33 cover have been noted in other Arctic, subarctic and boreal 34 regions (Langer et al. 2015; Matveev et al. 2019). In productive 35 systems, the isolation of the water mass under an ice cover 36 quickly generates anoxic conditions favoring methanogenesis 37 and suppressing methanotrophy. Our studied lakes showed 38 median CO<sub>2</sub>/CH<sub>4</sub> molar ratios that were much lower in winter 39 (2.2 for all lakes; ranging from 0.2 for unconnected alas lakes 40 to 10 for connected alas lakes) compared to summer (33 for all lakes; ranging from 4.4 for unconnected alas lakes to 56 for 42 recent thermokarst lakes) (Table 5). This suggests a strong pro-43 duction of CH<sub>4</sub> and the predominance of hydrogenotrophy 44 (consuming CO<sub>2</sub> for methanogenesis) in the anoxic condi-45 tions of winter, such as proposed in other thermokarst lakes of 46 subarctic peatlands (Matveev et al. 2019).

The greenhouse gases accumulated in the water column in 48 late winter will be partly or completely released in spring, as 49 soon as the ice cover degrades, a period when sampling is par-50 ticularly challenging logistically. The spring overturn period has been shown to be particularly important for greenhouse 52 gas release from lakes to the atmosphere (Langer et al. 2015). 53 Phelps et al. (1998) found that greenhouse gases released dur-54 ing spring overturn accounted for as much as half of the total

yearly emissions from Alaskan lakes. Our sampling occurred in 55 May, likely after the major venting period, since the ice cover 56 had already disappeared in many lakes (especially in uncon- 57 nected alas lakes, which were all ice-free in early May) and sur- 58 face waters had already started to warm (26 lakes were 59 presenting temperatures above 5°C, reaching up to 17.5°C). 60 This likely explains why lake water did not present high satu- 61 rations levels during this sampling period despite the high 62 concentrations observed 1 month earlier under the ice cover. 63 In fact, CO<sub>2</sub> concentrations were mainly below saturation in 64 May, indicating that primary producers had already started to 65 be active.

In summer, several recent thermokarst lakes were strongly stratified (examples given in Fig. 5), isolating bottom waters where oxygen depletion occurs and greenhouse gases accumu- 69 late. During August sampling, sufficient time had passed after 70 initiation of summer stratification to observe a significant 71 accumulation of greenhouse gases in bottom waters; for exam-72 ple, more than  $1500 \,\mu\text{M}$  of  $CO_2$  and  $50 \,\mu\text{M}$  of  $CH_4$  were measured at the bottom of recent thermokarst lake #1. This 74 summer storage of greenhouse gas will likely be released to the atmosphere during the fall overturn period, in addition to 76 greenhouse gas produced during the fall in sediments and 77 water column, which can be significant considering the time 78 lag before water and sediment temperatures cool down after 79 air temperature has dropped (Serikova et al. 2019). This could 80 explain the super-saturation in CO<sub>2</sub> (recent thermokarst lakes) and CH<sub>4</sub> (unconnected alas lakes) measured in fall (Fig. 6). Fall 82 sampling may have been too early for the highly stratified 83 thermokarst lakes, and thus the cumulated CH<sub>4</sub> at their bottom would be venting later. If this is the case, surface concen-85 tration (and emissions) of CO2 would also be higher at the 86 surface later in fall for these lakes. By contrast, connected alas 87 lakes are relatively well mixed during the summer, as they are 88 large water bodies influenced by wind-induced turbulence and water flows from regional streams and rivers (Fig. 2). These characteristics likely lead to the continuous release of CO<sub>2</sub> and CH<sub>4</sub> produced within the lake (and intermediate gas concen- 92 trations, Fig. 6), rather than an isolated release of the storage 93 flux during the fall overturn period. It is important to note that the intervals between sampling periods were not symmetrical throughout the year, and particularly short between summer (mid-late August) and fall (early September), explaining their similarities in greenhouse gas concentrations. These differing patterns underline the importance in placing automated sensors in lakes to better identify critical periods of greenhouse gas emissions (Matveev et al. 2019).

#### Diffusive greenhouse gas fluxes: Comparison across highlatitude regions

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Diffusive greenhouse gas fluxes from lakes studied across 105 Arctic and subarctic regions are generally limited to the icefree season, with only a few studies providing estimations from wintertime or over a year cycle (Langer et al. 2015;

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**Table 5.** The CO<sub>2</sub> to CH<sub>4</sub> molar ratios for all lakes sampled in central Yakutia along the seasons in 2018–2019. N is the number of lakes 55 sampled in each category.

		CO <sub>2</sub> /CH <sub>4</sub> ratio							
Туре	N	Minimum	Median	Mean	Maximum				
Fall 2018									
Recent thermokarst	6	2.5	52	86	219				
Connected alas	4	16	79	78	139				
Unconnected alas	4	0.3	1.0	1.6	3.9				
All lake types	14	0.3	27	60	219				
Winter 2019									
Recent thermokarst	11	0.2	1.4	2.8	8.8				
Connected alas	4	1.4	10	61	224				
Unconnected alas	2	0.2	0.2	92	275				
All lake types	17	0.2	2.2	31	275				
Spring 2019									
Recent thermokarst	23	0.3	1.8	16	118				
Connected alas	5	1.2	1,4	4.8	12				
Unconnected alas	17	0.2	1.3	6.7	73				
All lake types	45	0.2	1.8	11	118				
Summer 2019									
Recent thermokarst	16	0.6	56	83	383				
Connected alas	4	5.2	48	58	134				
Unconnected alas	12	1.0	4.4	21	148				
All lake types	32	0.6	33	57	383				

29 Matveev et al. 2019). Our estimations for diffusive greenhouse 30 gas fluxes are based on four discrete periods over a full year 31 cycle, including spring and fall, which have been previously 32 identified as critical periods for greenhouse gas evasion 33 (Jammet et al. 2015; Matveev et al. 2019). The diffusive CO<sub>2</sub> 34 fluxes estimated for this study site of Central Yakutia (range 35 from – 14 to 350 mmol m<sup>-2</sup> d<sup>-1</sup>; Table 4) are clearly outside 36 the range reported for lakes and ponds across the Arctic. The 37 most comparable fluxes are measured from ponds on Bylot 38 Island (Bouchard et al. 2015) and on the Mackenzie Delta 39 (Tank et al. 2008). The estimated diffusive CH<sub>4</sub> fluxes are also 40 one order of magnitude higher (up to 560 mmol m<sup>-2</sup> d<sup>-1</sup>) than published values across the Arctic (maximum values 42 below 20 mmol m<sup>-2</sup> d<sup>-1</sup> for different types of lakes, including 43 thermokarst lakes; (Wik et al. 2016) and almost as high as low 44 seep ebullition fluxes from erosive margins of thermokarst 45 lakes in Alaska and Siberia (Walter Anthony et al. 2010). An 46 important point to consider here is that our estimations only 47 include diffusive fluxes, while many studies have shown that 48 ebullition can represent a significant mode of CH<sub>4</sub> evasion 49 (Bastviken et al. 2011; DelSontro et al. 2016), particularly from 50 Arctic lakes (Walter et al. 2007). Some studies report that ebul-51 lition fluxes can be equivalent to diffusive fluxes, if not largely 52 dominant in certain conditions (Walter Anthony et al. 2010; 53 Wik et al. 2016), while others found that diffusion dominates 54 in shallow thermokarst lakes (Matveev et al. 2016). The

diffusive fluxes presented here already being in the upper 83 range of published values for thermokarst lakes across the cir- 84 cumpolar North, they can only increase when ebullition is 85 included, underscoring the need to consider such landscapes 86 in global assessments.

#### **Conclusions**

Our study site in Central Yakutia is within an area of thick, 91 ice rich permafrost, with the potential for substantial carbon 92 mobilization in the face of continued climate change. There is 93 considerable landscape variability and the thermokarst lakes 94 in this region vary in age from early Holocene to the last 95 decades. We observed temporal and spatial heterogeneity 96 (up to two orders of magnitude) in the concentrations and 97 fluxes of dissolved greenhouse gases from the three develop- 98 ment stages of thermokarst lakes found in the region. There 99 were also large differences in the relative importance of CO<sub>2</sub> 100 and CH<sub>4</sub> emissions, which is fundamental for assessing their 101 global warming potential. These differences are related in part 102 to the mixing regime that is controlled by lake morphology 103 and leading to variable seasonal patterns. They are also linked 104 to variable contributions from primary producers and ther- 105 mokarstic erosion shaping the light and oxygen availability. 106 We show that diffusive fluxes of both CO<sub>2</sub> and CH<sub>4</sub> from 107 thermokarst and alas lakes of Central Yakutia are among the 108

1 highest presented across Arctic and subarctic regions. With 2 their extreme bottom greenhouse gas concentrations and large 3 but variable emissions, recent thermokarst lakes need to be 4 closely considered as they likely involve the mineralization of 5 ancient organic carbon that can contribute to the amplifica-6 tion of the greenhouse effect. On the other hand, uncon-7 nected alas lakes act as active CO<sub>2</sub> sinks where primary 8 production fuels methanogeny, likely dampening total emis-9 sions from the region depending on the overall greenhouse 10 gas budget. The high levels of temporal and spatial heteroge-11 neity between seasons and lake types illustrate the complexi-12 ties of such permafrost landscapes and their responses to 13 changes in temperature, precipitation, and anthropogenic 14 influences. The fate of the large amounts of organic carbon 15 stored in permafrost landscapes of Central Yakutia needs to be 16 assessed to identify in which cases lakes contribute to its trans-17 fer to the atmosphere, especially considering the continued 18 influence of climate change on these sensitive environments. 19

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#### **Conflict of Interest**

None declared.

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