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1 **Numerical modeling of a regional groundwater flow system to assess**
2 **groundwater storage loss, capture and sustainable exploitation of the**
3 **transboundary Milk River Aquifer (Canada – USA)**

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11

12 **ABSTRACT**

13 Groundwater capture and storage loss play a major role in the sustainable exploitation of a
14 regional aquifer. This study aimed to identify the impact of major and long-term groundwater
15 exploitation on a regional aquifer system to understand the processes controlling the sustainable
16 exploitation of the transboundary Milk River Aquifer (MRA). The MRA extends over 26,300 km²,
17 being a major water resource across southern Alberta (Canada) and northern Montana (USA).
18 Concerns about the sustainability of the MRA were raised as the century-old exploitation has led
19 to important drawdowns and the local loss of historical artesian conditions. A steady-state
20 numerical model of the regional flow system was developed and calibrated against hydraulic
21 heads, groundwater fluxes, and the area with flowing artesian wells. Four groundwater
22 abstraction scenarios were simulated: 1) natural flow conditions without exploitation; 2) the mean
23 abstraction rate over the last 108 years; 3) the historical maximum global abstraction rate of the

24 MRA; and 4) a theoretical high abstraction rate based on the maximum rate estimated for each
25 MRA exploitation zone. The numerical model agrees with the previously formulated conceptual
26 model and supports its hydraulic plausibility. Results show that MRA exploitation has led to a
27 major change in flow patterns to sustain groundwater abstraction. The MRA water balance under
28 exploitation indicates that more recharge and reduced seepage to bedrock valleys compensate
29 groundwater withdrawals. Based on its impact on regional discharge and the reduction in MRA
30 storage, the mean historical level of exploitation of the MRA appears sustainable. Larger
31 exploitation rates would significantly reduce groundwater discharge to surface seepage locations
32 and lead to a larger reduction in groundwater storage in the MRA. Modeling also illustrates that
33 the MRA is an internationally shared resource. This situation would justify a joint management of
34 the aquifer system between Canada and USA; especially in the area comprised between the
35 recharge area in Montana and the Canadian reach of the Milk River.

36

37

38 **Keywords:** Regional aquifer, Sustainable exploitation, Numerical modelling, Storage changes,
39 Capture, Transboundary aquifer

40 **1. Introduction**

41 Numerical groundwater flow models have proven to be efficient tools to address a variety of
42 water related issues. Applications includes the assessment of sustainable groundwater
43 exploitation, transport of groundwater contaminants, the determination or control of saltwater
44 intrusion (Alwathaf and Mansouri, 2012; Bordeleau et al., 2008; Gaur et al., 2011; Giambastiani
45 et al., 2007; Islam et al., 2016; Meyer, 2014; Sowe, 2017).

46 At the basin scale, numerical models can provide a better understanding of regional
47 groundwater flow systems, including cross-formational flow through the aquitards, following the
48 pioneering work of Toth (1963) and Freeze and Witherspoon (1966a, 1966b). Numerous
49 numerical modelling studies have provided examples of regional groundwater flow (Janos et al.,
50 2018; Lavigne et al., 2010; Michael and Voss, 2009; Rabelo and Wendland, 2009; Voss and
51 Soliman, 2014; Zhou and Li, 2011). To assess the conditions required for the sustainable
52 exploitation of a regional aquifer, it is necessary to determine the impact of exploitation on the
53 flow system, which is controlled in part by the hydraulic connections between the
54 hydrogeological units and surface water features composing the regional flow system. As initially
55 stated by Theis (1940) and more recently by Konikow and Leake (2014), during groundwater
56 withdrawal part of the groundwater is derived from storage, but over time the aquifer system
57 adjusts to pumping through an increase in recharge or a decrease in discharge, which is called
58 capture. Groundwater capture and storage loss are key concepts related to the sustainability of
59 aquifer exploitation (Konikow, 2015; Konikow and Leake, 2014) and should be quantified while
60 investigating the sustainable exploitation of a regional aquifer system.

61 In that perspective, the objective of the present study is to develop a numerical model to assess
62 the regional groundwater flow dynamics of a major aquifer under both natural and exploitation
63 conditions to determine the conditions required for the sustainable management of groundwater
64 resources. More specifically, the role of groundwater capture and storage loss will be evaluated
65 at the regional scale.

66 The transboundary Milk River Aquifer (MRA) constitutes a perfect example of a regional aquifer
67 comprised in a large groundwater flow system. It is also part of a worldwide inventory of
68 transboundary aquifers whose characterization and management are the objects of recent
69 international initiatives such as ISARM or TWAP (IGRAC, 2015; Rivera and Candela, 2018;
70 TWAP, 2012). The MRA extends 26,300 km² over southern Alberta (Canada) and northern

71 Montana (USA) in a semi-arid region where water shortages are an important issue
72 (Government of Alberta, Alberta water for life, 2006). Concerns have been voiced for many
73 decades about the sustainability and mismanagement of groundwater exploitation from the
74 MRA, with indications of locally significant drawdowns (up to 30 m) and the loss of artesian
75 conditions in some areas (AGRA Earth and Environmental Limited, 1998; Borneuf, 1976;
76 Meyboom, 1960). A steady-state numerical model of the regional groundwater flow system
77 comprising the MRA was developed with the objective of understanding the impact of major and
78 long-term exploitation on the entire aquifer system and to identify the processes controlling the
79 sustainable exploitation of the MRA. The numerical model was based on a previously developed
80 3D geological model (Pétre et al., 2015) and a hydrogeological conceptual model (Pétre et al.,
81 2016) of the aquifer system integrating the MRA.

82 A simulation of conditions without groundwater exploitation and three scenarios of groundwater
83 extraction in the MRA are simulated to assess the impact of the MRA exploitation. This model
84 must first verify the plausibility of the previously formulated conceptual model and then
85 determine whether the regional flow system can adjust to the MRA exploitation so that a new
86 sustainable flow pattern is established. Knowing that direct recharge to the MRA does not
87 compensate groundwater abstraction (Pétre et al., 2016), this study also involves the
88 determination of groundwater capture and long-term storage loss.

89 This paper presents first the study area and the conceptual hydrogeological model of the MRA.
90 The Materials and Methods section describes the numerical model design, calibration criteria
91 and the groundwater use evaluation. Then, the simulation and water budget results are
92 presented, followed by a discussion covering the limitations of the model and the implications for
93 groundwater management.

94

95 2. Study area and conceptual hydrogeological model

96 2.1 Study area

97 The MRA is located in a semi-arid region, spanning southern Alberta and northern Montana (Fig.
98 1). The mean annual precipitation is between 250 and 450 mm/y and the potential
99 evapotranspiration ranges from 550 to 578 mm/y (Climate Canada, 2015; NOAA, 2015). The
100 topographic highs in the region are the Sweet Grass Hills and Bears Paw Mountains in Montana,
101 and the Cypress Hills and Milk River Ridge in Alberta. The Sweetgrass Hills are an ensemble of
102 three buttes south of the international border. The hydrography of the region includes the
103 transboundary Milk River, the shallow Pakowki Lake and several valleys with intermittent
104 streams called “coulees” (e.g., Etzikom, Chin and Forty Mile coulees).

105 The stratigraphic sequence in the study area, from the base to the surface, is as follows: the
106 500-m thick regional aquitard of the Colorado Group underlies the study area. The shales of the
107 Colorado Group contain several thin sandstone beds, the most significant being the 25-m thick
108 Bow Island Sandstone. The Colorado Group is overlain by the Milk River Formation (called
109 Eagle Formation in Montana), which is subdivided into three members in Alberta: the basal
110 Telegraph Creek Member, the middle Virgelle Member and the upper Deadhorse Coulee
111 Member. The Virgelle Member constitutes the MRA as it is the most important aquifer unit within
112 the Milk River Formation. The Milk River Formation (about 100 m thick) subcrops or outcrops
113 near the international border in Alberta, in rings around the Sweetgrass Hills and also following
114 two branches on both sides of the Sweetgrass Arch (the outcrop area is shown on Fig. 1). The
115 Milk River Formation is overlain by the low-permeability shales of the Pakowki/Claggett
116 Formation (about 130 m thick). The Belly River Group (Judith River Formation in Montana)
117 overlies the Pakowki/Claggett aquitard and is also considered as an aquifer. With the exception

118 of the topographic highs and coulees, the study area is covered by glacial drift which consists
119 mainly of low-permeability till, typically less than 2 m in thickness (Hendry and Buckland, 1990).

120 Buried valleys (bedrock channels) are present across the study area (Fig. 1). These buried
121 valleys are preglacial stream valleys buried under glacial drift (Cummings et al., 2012a). In
122 southern Alberta, the Medicine Hat, Skiff and Foremost buried valleys are up to 10 km wide and
123 are incised up to 30 m into bedrock (HCL consultants, 2004; Hendry and Buckland, 1990).
124 Buried valleys locally constitute productive aquifers where the fill material is predominantly sand
125 and gravel (Cummings et al., 2012a; Farvolden et al., 1963; HCL consultants, 2004).

126 The extent of the numerical model described in this paper follows the hydrogeological limits of
127 the MRA previously defined by Pétré et al. (2016) in the west, north and east (Fig. 1). The north-
128 eastern hydrogeological limit of the MRA corresponds to a low permeability facies hosting the
129 Medicine Hat natural gas fields. In the south, the Marias River and Cut Bank Creek have been
130 chosen as the physiographic limits of the model, although the MRA may extend farther south in
131 Montana. In the south-east corner, the numerical model is limited by the extent of the geological
132 model developed by Pétré et al. (2015) (at longitude -110°), which is the basis of the numerical
133 model.

134

135 **2.2 Conceptual hydrogeological model**

136 The conceptual hydrogeological model of the MRA developed by Pétré et al., (2016) was used
137 as a basis for the development of the numerical groundwater flow model. Only a brief description
138 of the conceptual model is presented here since details can be found in Pétré et al. (2016).
139 Figure 2 shows a cross-section through the MRA from its outcrop area at high elevation to its
140 downgradient limit (location of line AA' shown on Fig. 1). The MRA is a typical confined aquifer,
141 radially dipping from the outcrop/subcrop areas (Fig. 2). Direct recharge of the MRA occurs

142 mainly in the outcrop or subcrop areas of the Milk River Formation where unconfined conditions
143 and modern groundwaters are present, as indicated by the presence of tritium (Pétre et al.,
144 2016). Groundwater inflow into the MRA also occurs through subsurface vertical inflow from
145 overlying geological units in the topographic highs of the study area. Groundwater flow diverges
146 from the Sweetgrass Hills to the north, east and southeast. West of the Sweet Grass Arch,
147 groundwater flows south-west and north from a groundwater divide located north of Cut Bank.
148 Two transboundary flowpaths were defined on the basis of a potentiometric map (Pétre et al.,
149 2016): (1) an eastern flowpath from the Sweet Grass Hills to the north and (2) a western flow
150 path from the northern part of Cut Bank to the north.

151 An abrupt change in the horizontal hydraulic head gradient indicates that the Milk River and part
152 of the Verdigris Coulee intercept a large proportion of the groundwater flowing to the north from
153 direct MRA recharge areas. As no other natural surface discharge feature has been identified,
154 vertical leakage through the confining units was inferred to be another important natural discharge
155 mechanism. Cross formational flow is enhanced along the talwegs of buried valleys which are
156 acting as drains, as inferred by Toth and Corbet (1986) and shown by Pétre et al. (2016).
157 Indeed, the buried valleys have eroded the upper bedrock (Belly River /Judith River and
158 Pakowki/Claggett formations), thus reducing the thickness of the bedrock between the MRA and
159 surficial sediments. Under the confined conditions found north of the Milk River, the MRA
160 contains a fossil groundwater, not significantly renewed by modern recharge. This fossil
161 signature is demonstrated by the absence of radiocarbon and a $^{36}\text{Cl}/\text{Cl}$ ratio indicating a
162 groundwater residence time reaching 2 My further north of the MRA (Pétre et al., 2016).

163

164

165 3. Model simulation and calibration

166

167 3.1 Model design and boundary conditions

168 The basis of the numerical groundwater flow model of the aquifer system encompassing the
169 MRA is the three-dimensional (3D) geological model that was previously developed with the
170 software Leapfrog Hydro (Pétre et al., 2015). This software was then used to convert the
171 geological model in a finite element grid compatible with the numerical groundwater flow
172 simulator FEFLOW (Diersch, 2014). A 2D areal finite element mesh was first created in
173 FEFLOW and was then applied to the layers of the 3D geological model, resulting in a 3D
174 numerical model grid comprising 15 layers (Fig. 3). The geometry and thicknesses of the
175 geological layers remained unchanged in the numerical model.

176 The numerical model covers a surface area of 26,300 km² with a volume of approximately
177 30,000 km³. The top surface of the groundwater flow model corresponds to ground level,
178 represented by the Digital Elevation Model (DEM) of the study area (pixel size is 500 m x 500
179 m). The domain is discretized into 329,825 triangular prismatic mesh elements and 165,587
180 nodes per slice. The finite element mesh was locally refined along the Milk River where a steep
181 horizontal hydraulic gradient was expected. The lateral sizes of the elements in the mesh vary
182 from 100 m, where the grid was refined, to 650 m in the remainder of the domain.

183 The model considers seven hydrostratigraphic units (Fig. 3): surficial sediments, bedrock
184 valleys, the Belly/Judith River Formation, the Pakowki/Claggett Formation, the Milk River
185 Formation, the Colorado Group and the Bow Island Sandstone.

186

187

188

189 Boundary conditions in the model domain are summarized in Fig. 4d.

190 Specified heads with a seepage constraint were assigned to the streams to precludes any inflow
191 into the model from the stream nodes in accordance with the conceptual model showing losing
192 streams.

193 A hydraulic head corresponding to ground elevation was assigned to the nodes along the
194 streams (Fig. 4a). The basal layer of the model was defined as a no flow boundary. Along the
195 outer boundary of the domain on layer 14 (corresponding to the Bow Island Sandstone), a
196 specified head boundary condition was set to 750 m. This value is based on the potentiometric
197 map of the Bow Island Sandstone produced by Swanick (1982) in Alberta (Fig. 4b). The
198 recharge rate was set to 0 mm/y where the Colorado Group aquitard outcrops; since this
199 aquitard underlies the MRA, the MRA is absent in these areas. In the outcrop/subcrop area,
200 which constitutes the direct recharge area of the MRA, the recharge rate was set to 10 mm/y,
201 based on previous estimates (Pétre et al., 2016). Elsewhere, the recharge rate was adjusted to 1
202 mm/y during model calibration (Fig. 4c).

203

204 **3.2 Hydraulic conductivity of the hydrostratigraphic units**

205 The K value of the MRA was calculated on the basis of the transmissivity and thickness maps of
206 the Milk River Formation derived from the geological model (Pétre et al., 2016, 2015). The K
207 values assigned to the MRA range from 8.1×10^{-9} m/s to 9.4×10^{-4} m/s. The spatial distribution of
208 K (ESM1) was applied to the elements corresponding to the Milk River Formation. Hydraulic
209 conductivity (K) of the hydrostratigraphic units are summarized in Table 2. Except for the MRA,
210 all hydrostratigraphic units were assigned a uniform value based on the limited number of points

211 estimates available. K estimates for most of units were adjusted during calibration given the
212 large uncertainties related to their spatial variability (Table 2).

213

214 **3.3 Groundwater use evaluation in southern Alberta**

215 Information about groundwater use is limited in the study area because the water right holders
216 do not have the statutory requirement to measure their groundwater withdrawals. An
217 assessment of the historical groundwater use in the MRA for the period 1908-2015 was carried
218 out in southern Alberta by Pétré (2016) using 1655 available wells. An extraction rate was
219 assigned to each well according to their intended use (domestic, stock, municipal or industrial
220 use). For this assessment, six exploitation zones were delineated (Fig. 5) on the basis of the well
221 density and the periods at which these wells were installed in the MRA. Those exploitation zones
222 are represented in the numerical model as diffusive sinks, but the flow rate of the eight main
223 municipal wells were considered as local sinks. To model the impact of MRA, the model was first
224 run without pumping in the MRA to establish natural conditions and then three MRA exploitation
225 scenarios based on the estimated historical exploitation were simulated from the natural
226 conditions scenario (Table 1):

- 227 • Scenario 1, mean exploitation rate: this scenario uses the mean exploitation rate in each
228 zone over the 108 years considered, which also corresponds to the overall mean
229 historical exploitation rate of the MRA;
- 230 • Scenario 2, maximum exploitation rate: over the years, the exploitation rate of the MRA
231 had steadily increased from 1908 to the mid-1990's and has been in decline since then
232 (Pétré 2016). Thus, this scenario considers the maximum exploitation rate of the mid-
233 1990's, which is spatially distributed over the exploitation zones based on the mean
234 contribution of each zone over the exploitation period;

235 • Scenario 3, ultimate exploitation rate: MRA exploitation was not uniformly distributed
 236 among the exploitation zones, which have had maximum exploitation rates over different
 237 periods. This last scenario thus considers an exploitation rate of the MRA that
 238 corresponds to the maximum exploitation rate historically reached within each
 239 exploitation zone. The overall exploitation rate of the MRA for this scenario is thus larger
 240 than the historically reached maximum rate (scenario 2).

241

242

243 **Table 1 Groundwater extraction rate assigned to each exploitation zones for the three exploitation**
 244 **scenarios simulated with the numerical model. The mean and maximum rates correspond to the**
 245 **overall historical level of MRA exploitation, whereas the ultimate rate corresponds to the maximum**
 246 **historical rate within each zone.**

MRA exploitation zone	Scenario 1 (Mm ³ /y)	Scenario 2 (Mm ³ /y)	Scenario 3 (Mm ³ /y)
Zone 1	0.311	0.598	0.603
Zone 2	0.019	0.026	0.056
Zone 3	0.111	0.211	0.227
Zone 4	0.220	0.389	0.655
Zone 5	0.028	0.041	0.064
Zone 6	0.471	0.805	0.888
Municipal wells	0.16	0.17	0.17
Total extraction rate in Domain	1.160	2.070	2.493
Total extraction rate in subdomain	0.963	1.564	1.954
Total extraction over 108 years (Mm3)-Domain	142.1	241.9	287.6
Total extraction over 108 years (Mm3)-Subdomain	104	169.8	211.2

247

248 3.4 Numerical model calibration criteria

249 Calibration of the numerical model was based on a set of quantitative and qualitative criteria.

250 Thus, hydraulic heads measured in wells, conceptual direction and magnitude of groundwater

251 fluxes as well as the extent of artesian areas were used to guide model calibration as discussed
252 below.

253 There is very limited information related to the pre-development state of the aquifer. The steady-
254 state calibration dataset consists of hydraulic heads (or calibration targets) produced from
255 historic potentiometric surfaces of the MRA that are considered as representative of a steady-
256 state condition not significantly affected by exploitation. Moreover, qualitative information
257 describing the state of the aquifer at the beginning of the MRA exploitation was used. In
258 Montana, the available potentiometric surface is from Pétré (2016). This map is actually a
259 composite map of historical potentiometric maps from Levings (1982) in south-east Montana,
260 Tuck (1993) in the Sweet Grass Hills area and Zimmerman (1967) in south-west Montana.
261 Although these maps have been drawn after years of groundwater exploitation, it was assumed
262 that they represented a condition similar to predevelopment. Indeed, the south-west area in
263 Montana appears to have stable groundwater levels based on monitoring wells completed in the
264 MRA. Besides, this area is likely under the influence of active recharge due to its higher
265 transmissivity and closer proximity to the outcrop of the aquifer where direct recharge can occur
266 (Figs. 4c and ESM1). In south-east Montana, the investigation carried out in the present study
267 indicates that the magnitude of water use for oil and gas activity is low, except in the extreme
268 south-east corner of the study area (Folnagy A.J.B., Montana Department of Natural Resource
269 Conservation, personal communication). Eleven water level measurements from monitoring
270 wells, mostly located in the outcrop area of the MRA in Montana, were added to the calibration
271 dataset.

272 In Alberta, the same steady-state hypothesis was formulated in the area south of the Milk River,
273 where it is assumed that the groundwater use is minor compared to what is found north of the
274 river. Furthermore, this area receives recharge from the outcrop and subcrop areas that is not
275 intercepted yet by the Milk River. The potentiometric maps used in Alberta are from Meyboom,

276 (1960) and Toth and Corbet, (1986), who reinterpreted Meyboom's data by better considering
277 the surface topography and the potential effect of buried valleys. Simulated heads will be
278 compared to both potentiometric interpretations. Such model calibration to groundwater levels
279 could not be done in the area north of the Milk River, which has been subjected to intensive
280 exploitation and does not have significant renewal from recharge in the outcrop/subcrop of the
281 MRA due to the interception of groundwater by the Milk River.

282 A dataset of observation points was defined by randomly selecting points from the interpolated
283 potentiometric surfaces that were interpolated beforehand within the model domain. As shown in
284 Fig. 6, 132 observation points were defined in northern Montana and 80 points in Alberta (south
285 of the Milk River). These observation points are assigned in the model within the layer at the
286 centre of the Milk River Formation.

287 The discharge mechanism of the MRA through cross-formational flow has been highlighted in
288 the conceptual model of the aquifer Pétré et al. (2016). The direction and magnitude of these
289 cross-formational flows previously estimated are thus used here as a calibration criterion. More
290 specifically, the upward flow component from the MRA to surficial sediments in the vicinity of the
291 bedrock valleys was estimated to be between 4.0×10^2 to 4.0×10^5 m³/y. Another cross-
292 formational flow directed downward from the MRA through the Colorado Group and to the Bow
293 Island Sandstone was also defined and estimated to be between 8.0×10^3 to 8.0×10^5 m³/y.

294 Qualitative information on the occurrence of artesian conditions in the MRA provides an
295 indication of the state of the pre-exploitation system. Previous studies indicate that nearly all the
296 wells drilled in the MRA in southern Alberta were flowing in the pre-exploitation system (Borneuf,
297 1976; Hendry et al., 1991; Phillips et al., 1986). (Dowling, 1917) defined the flowing artesian limit
298 in southern Alberta (Fig. 6). The steady-state model should thus represent this flowing artesian
299 area and be consistent with Dowling's (1917) delineation. Although the magnitude of artesian
300 conditions was not quantified in the past, four recent pressure measurements in the flowing area

301 are available Pétré et al. (2016). It is assumed that under pre-development conditions, the
302 hydraulic heads at these locations would have been greater than the present-day observations.

303

304 **3.5 Adjustments of boundary conditions and hydraulic parameters**

305 During the calibration process, the following modifications were made to improve the match
306 between observed and simulated hydraulic heads. The density of the surface drainage was first
307 adjusted locally during the calibration process. It was increased in a poorly drained area in the
308 central part of southern Alberta and decreased north of the Etzikom coulee to better represent
309 the observed spatial distribution of the flowing artesian area. These modifications imply that
310 groundwater discharge to surface drainage from the MRA and through the overlying aquitard
311 exerts an important control on hydraulic heads (and artesian conditions) in the MRA. Such
312 discharge, while being relatively diffuse, could still be largely controlled by the presence of
313 surface drainage representing low topography areas. Conversely, this could also mean that
314 surface drainage partly reflects the discharge of groundwater originating from the MRA. This
315 adjustment is in agreement with the observation from (Meyboom, 1960) who identified strong
316 flowing wells along the main coulees.

317 In the Milk River canyon area, the value of the specified head boundary condition was adjusted
318 as the resolution of the DEM representing the model surface elevation did not allow the proper
319 representation of the 150-m deep and 1500 m-wide canyon (Beaty, 1990). The specified head
320 was thus set to 30 m below land surface (as defined for the DEM and numerical model) to
321 correctly represent the incision of the Milk River canyon and to increase its effect on
322 groundwater interception, which led to a better representation of observed potentiometric
323 conditions. This adjustment provided a better fit with the calibration target, suggesting that the
324 Milk River Canyon is an important feature in the discharge mechanism of the MRA.

325 Table 2 summarizes the initial ranges of K values and vertical anisotropy for each
 326 hydrostratigraphic units and their final values that provided the best match between the
 327 numerical model and calibration criteria. A few new hydraulic tests recently conducted in the
 328 south-eastern part of the study area in Montana indicate that K of the Milk River/Eagle Formation
 329 is higher than indicated by ESM1 (A. Fohnagy, Montana Department of Natural Resource
 330 Conservation, personal communication). To take into account these new observations, K of the
 331 MRA in the south-east corner of the study area was locally increased by a multiplication factor,
 332 adjusted during the calibration process. The numerical model should thus better represent K of
 333 the MRA. Miller and Norbeck (1996) give a K value for the Pakowki/Claggett aquitard in the East
 334 Butte area (the eastern Butte of the Sweetgrass Hills) that is much higher than the uniform value
 335 used for this geological unit. This higher value of 9×10^{-8} m/s was assigned to the
 336 Pakowki/Claggett Formation in the East Butte area. It provided a better match of hydraulic heads
 337 in this area, suggesting that a zonation of K of the Pakowki/Claggett aquitard is necessary. This
 338 modification indicates that this aquitard probably does not have spatially uniform hydraulic
 339 properties over the study area but available data do not allow the definition of the spatial
 340 distribution of K.

341 **Table 2 Estimated ranges of hydraulic conductivity and vertical anisotropy for each**
 342 **hydrostratigraphic units and their final (calibrated) values**

Hydro- stratigraphic units	Horizontal hydraulic conductivity K_x (m/s)	K anisotropy (K_x/K_z)	Sources	Final values	
				Horizontal hydraulic conductivity K_x (m/s)	K anisotropy (K_x/K_z)
Surficial sediments (till)	7×10^{-8}	1	Robertson (1988); Hendry and Buckland (1990)	7×10^{-8}	1
Belly/Judith River Formation	9×10^{-8} - 8.8×10^{-7}	10-10 ⁴	Hendry and Buckland (1990); Anna (2011)	5×10^{-7}	10 ³
Pakowki/ Claggett Formation	10^{-9} - 10^{-11}	5	Swanick (1982); Hendry and Schwartz (1988)	1×10^{-9} (9×10^{-8} in East Butte area)	5
Milk River Formation	Spatial distribution from	10	Pétre et al. (2016)	Spatial distribution in	10

	Fig.5			ESM1 (factor 50 in south-east Montana)	
Colorado Group	10^{-9} - 10^{-12}	10	Toth and Corbet (1986); Robertson (1988); Hendry and Buckland (1990)	1×10^{-10}	10
Bow Island Sandstone	10^{-6} - 10^{-8}	10	(Schwartz et al., 1981)	5×10^{-7}	10
Bedrock valleys	10^{-3} - 10^{-4}	1	Cummings et al. (2012b)	1×10^{-6}	1

343

344 **4. Results**345 **4.1 Hydrogeological model calibration**

346 Model calibration was carried out by trial-and-error with the objective of obtaining the best fit
 347 between simulated and observed heads at the 212 calibration targets as well as a proper
 348 representation of the artesian conditions in the MRA. The calibration performance was quantified
 349 by calculating the root mean square error (RMSE), the correlation coefficient (r) for each
 350 calibration zone and the scaled RMSE (Table 3). The combination of a small RMSE and high r
 351 indicates a satisfying calibration. Furthermore, the scaled RMSE should be below the threshold
 352 value of 5%, as recommended by Anderson and Woessner (1992) and Giambastiani et al.
 353 (2012).

354 In Montana, good calibration is achieved in the south-west where the lowest RMSE (30.8 m) and
 355 the highest correlation coefficient (0.89) are found. In southern Alberta, south of the Milk River,
 356 the best calibration is achieved following the Toth and Corbet (1986) interpretation of the
 357 potentiometric surface, with a RMSE of 28.4 m and a satisfying r of 0.89. As the Toth and Corbet
 358 (1986) interpretation of the potentiometric surface mimics the topography, this suggests that the
 359 topography exerts a major influence on the pattern of the potentiometric surface. Figure 8b
 360 shows the effect of both Toth and Corbet's (1986) and Meyboom's (1960) interpretations on the

361 calibration performance. This result illustrates the uncertainty of the reference potentiometric
362 map as several interpretations lead to different calibration performance. However, the model
363 indicates that the Toth and Corbet (1986) interpretation of the potentiometric map appears to be
364 the most hydraulically plausible.

365 In southern Alberta and the south-western part in Montana, the scaled RMSE values are below
366 or close to the 5% limit (4.9 and 5.3%). The other two calibration zones in the south-eastern part
367 of the study area in Montana and in the Sweet Grass Hills area show lower correlation
368 coefficients (0.85 and 0.83) and have higher RMSE (44.4 and 35.9), which implies poorer
369 calibration. The scaled RMSE value are higher than the 5% limit in these calibration zones (7.7
370 and 6.2%).

371 The calibration is poorer in the south-east area in Montana because this area combines the
372 highest uncertainty in the geological model (due to a lack of geological data) and the highest
373 uncertainty in the reference water levels. Indeed, the water levels are derived from a regional
374 potentiometric map at the Montana-state level (Levings, 1982). The steep topography in the
375 Sweetgrass Hills area explains the difficulty to obtain a good calibration in this calibration zone.

376 A detailed analysis of the levels of calibration for each calibration zone is presented in the
377 electronic supplemental material ESM2. The comparison between contours maps of observed
378 and simulated heads shows a good representation of the observed radial groundwater flow
379 pattern as indicated in the electronic supplemental material ESM3. The groundwater divide north
380 of Cut Bank is also satisfactorily reproduced by the model. The scatter plot of simulated versus
381 measured heads is shown in Fig. 7a. The model tends to overestimate the simulated heads in
382 the south-eastern part of the study area in Montana and in southern Alberta.

383

384 **Table 3 Hydraulic head calibration performance of the groundwater flow model in the calibration**
385 **zones**

<i>Calibration Zone</i>	<i>RMSE (m)</i>	<i>Correlation coefficient (r)</i>	<i>Scaled RMSE (%)</i>
Montana, south-west	30.8	0.89	5.3
Montana, Sweet Grass Hills	35.9	0.83	6.2
Montana, south-east	44.4	0.85	7.7
Alberta, south of the Milk River (Toth and Corbet 1986)	28.4	0.89	4.9
Alberta, south of the Milk River (Meyboom 1960)	41.3	0.85	7.1

386

387

388 Simulated flowing artesian conditions in the MRA are mostly located in the northern part of the
389 study area, along the Chin and Etzikom Coulees, in the vicinity of Lake Pakowki and following
390 the Milk River reach in the north-eastern part of the study area in Montana (ESM4). In Alberta,
391 these locations are consistent with the delineation of the artesian zone from Dowling (1917) and
392 the observation of strongly flowing wells in the coulees (Meyboom, 1960). Approximately 62%
393 of the MRA nodes located in the flowing artesian area defined by Dowling (1917) are indeed
394 simulated as flowing nodes. At the four locations where recent pressure measurements were
395 obtained (ESM4), the magnitudes of the simulated artesian conditions are higher than the
396 measurements. This result was expected as pre-development artesian conditions must have had
397 a greater intensity than the present-day conditions. The numerical model thus successfully
398 represents the location and magnitude of the flowing artesian area in Alberta. In south-east
399 Montana, the simulated artesian conditions are located along the southern reach of the Milk
400 River and Sage Creek. The presence of artesian conditions at this location has not been
401 documented in previous studies. It is however plausible due to the combination of confined
402 conditions in the MRA and lower elevations, where the river is more incised, that could result in
403 strong artesian conditions, as observed in Alberta. Nevertheless, high magnitude in artesian
404 conditions locally is most probably due to the low-permeability Pakowki Formation which
405 outcrops along the Milk River due to uncertainties in the geological model in this area. Besides,
406 the high uncertainty in K of the MRA in this area can also explain these large values.

407
408 In order to assess the performance and uncertainty of the model relative to changes in
409 calibration parameters, a parameter sensitivity analysis was carried out (electronic supplemental
410 material ESM5). This analysis shows that the calibrated parameters provide a balance between
411 hydraulic heads calibration and the representation of artesian conditions. This analysis also
412 showed that the model is very sensitive to the recharge rate and that the most plausible
413 recharge value could only be between the calibrated value and a value 1.5 times greater. The
414 determination of the actual recharge rate is critical in addressing the sustainability of the MRA
415 exploitation.

416

417 **4.2 Simulation of groundwater extraction scenarios**

418 The steady-state calibrated model was used to simulate natural conditions without groundwater
419 exploitation as well as three groundwater extraction scenarios (Table 1), as presented in section
420 3.3 "Groundwater use evaluation in southern Alberta". One indicator of the long-term effect of
421 groundwater extraction is the simulated steady-state regional drawdown compared to natural
422 conditions that is caused by the three exploitation scenarios over Alberta (Fig. 8): a) mean
423 groundwater extraction, b) maximum groundwater extraction, and c) ultimate groundwater
424 extraction corresponding to the maximum historical rate in each exploitation zone (Fig. 5).

425
426 For scenario 1 with mean groundwater extraction rate, drawdowns between 2 and 10 m are
427 simulated in the central part of the domain where a large proportion of the groundwater pumping
428 is found. The drawdown is also higher in the vicinity of the main municipal wells who were
429 represented individually in the numerical model, notably in Foremost. These drawdowns have
430 led to a loss of 9% of artesian conditions compared to natural conditions.

431 In the case of scenario 2 representing the maximum groundwater extraction rate, higher
 432 drawdown is simulated (5 to 20 m) in the central and north-eastern parts of the domain. In that
 433 case, artesian conditions decrease by 15% compared to natural conditions. Drawdowns of about
 434 122 m in the Foremost area approach the top of the MRA.

435 In scenario 3 representing the ultimate MRA exploitation rate in each zone, simulated
 436 drawdowns increase dramatically in the central and north-eastern part of the domain where their
 437 range is from 10 to 30 m in the central part. The loss of artesian conditions is 21% compared to
 438 the natural conditions. The simulated local drawdown of the Foremost well reaches the top of the
 439 MRA in scenario 3.

440 Since groundwater extraction is spatially distributed in the numerical model, the simulated local
 441 drawdown is minimized compared to the actual drawdown that would be observed for a specific
 442 water well. To assess the impact of groundwater extraction on the economic exploitability of the
 443 MRA, it is necessary to take into account the additional local drawdown at extraction wells,
 444 except for the main municipal wells that are already represented individually in the model. Table
 445 4 shows the local expected drawdown for the range of transmissivity values of the MRA. The
 446 drawdown was calculated using 50% of a typical pumping rate of 7500 m³/y for domestic and
 447 stock use (Alberta Water Act, 2000), a well casing radius of 0.06 m and the transmissivity data of
 448 the MRA from Meyboom (1960) and Persram (1992). This result indicates that an additional few
 449 meters of drawdown could be expected locally, depending on the transmissivity and the actual
 450 pumping rate.

451
 452 **Table 4 Calculated local drawdown at exploitation wells for the range of MRA transmissivity values**
 453 **in Alberta**

$T (m^2/s) \times 10^{-5}$	20	2	9	0.52
Local drawdown (m)	64	6	1	0

454

455 Fig. 8d presents the approximate drawdown derived from local water wells between the late
 456 1950s and the late 1990s (AGRA Earth and Environmental Limited, 1998). Drawdowns may be
 457 higher than those simulated in the model due to local effects. However, the order of magnitude
 458 as well as the spatial distribution of the drawdown are similar. This suggests that the system
 459 rapidly equilibrates with respect to the period of exploitation. Therefore, it seems acceptable to
 460 assess the sustainability of the MRA exploitation under steady-state conditions.

461 The water level drawdown induces a loss of storage in the MRA. It is possible to estimate the
 462 loss of storage from the mean simulated drawdown as shown in Table 5. With increasing
 463 groundwater abstraction, the loss of storage increases from 4.1 Mm³ to 8.3 Mm³ and from 12.4
 464 to 24.8 Mm³, using a storage coefficient of 1×10^{-4} and 3×10^{-4} respectively. This range of value in
 465 the storage coefficient of the MRA is derived from local estimates from pumping tests in the
 466 Foremost and Sweet Grass Hills area (Meyboom 1960, DNRC, unpublished report).

467 **Table 5 Loss of groundwater storage in the MRA in Alberta under the three simulated exploitation**
 468 **scenarios**

		Scenario 1	Scenario 2	Scenario 3
Mean drawdown (m) in the MRA		4.1	6.5	8.2
Loss of storage (Mm ³)	S= 1×10^{-4}	4.1	6.5	8.3
	S= 3×10^{-4}	12.4	19.6	24.8

469

470 **4.3 Water balance**

471 The water flux into the model is controlled by the imposed recharge (Fig. 4c), thus the water
 472 balance of the full model is the same for simulated natural conditions, as well as for the three
 473 exploitation scenarios The entire model domain has a simulated annual groundwater flow of 243
 474 Mm³/y (1 Mm³ = 1×10^6 m³) and all models converged with a water imbalance of <0.01%.

475 Water balances were obtained from the numerical model under natural conditions as well as for
 476 the three exploitation scenarios. These water balances considered a subdomain of the numerical

477 model in the area north of the Milk River and east of the town of Warner, where the majority of
 478 groundwater exploitation takes place. The use of this subdomain also allows the comparison of
 479 the numerical model water balance with the one related to the conceptual model estimated by
 480 Pétré et al. (2016). In order to assess the impact of the MRA exploitation on the entire regional
 481 flow system, the water balance was first established for the regional flow system (Table 6) and
 482 then specifically for the MRA (Table 7).

483 **Table 6 Water balance of the entire flow system in Alberta under steady state conditions**

Aquifer system water balance (north of the Milk River) m³/d	Natural conditions	Mean extraction rate	Maximum extraction rate	Ultimate extraction rate
Recharge	21739	21739	21739	21739
Discharge to seepage locations	-21305	-19526	-17262	-16261
Discharge to Bow Island perimeter nodes	-1109	-1082	-1064	-1045
Net flux to(-)/from(+) the MRA	0	-2086	-3695	-4593
Flux from the western limit of the subdomain	98	112	124	124
Flux beyond the Milk River in other units	189.5	309.81	392.53	424.83
Water balance error	-388	-534	234	389
Water balance error (% of total recharge)	-1.8	-2.5	1.1	1.8
Aquifer system water balance (north of the Milk River) - Relative to the total recharge (%)				
Recharge	100.00	100.00	100.00	100.00
Discharge to seepage locations	-98.00	-89.82	-79.41	-74.80
Discharge to Bow Island perimeter nodes	-5.10	-4.98	-4.89	-4.81
Net flux to(-)/from(+) the MRA	0.00	-9.60	-17.00	-21.13
Flux from the western limit of the subdomain	0.45	0.52	0.57	0.57
Flux beyond the Milk River in other units	0.87	1.43	1.81	1.95

484 For the entire flow system, the water balance under exploitation sees a reduction in discharge to
 485 seepage locations that mostly compensates the groundwater abstracted from the MRA. In
 486 scenario 1 (mean exploitation), this reduction in seepage is in the order of 8%, but it could reach
 487 about 24% in scenario 3. Besides, the net flux to/from the MRA increases up to 4593 m³/d in

488 scenario 3, which correspond to 21% of the total recharge. The other components of the budget
489 do not change significantly under exploitation.

490 At the MRA scale, the water balance components are depicted in Table 7 and Fig. 9 and further
491 described below:

492 **Transboundary fluxes and effective MRA recharge.** The simulated transboundary flux is
493 16,320 m³/d (5.96 Mm³/y), while in the conceptual model it is 24,657 m³/d (9.0 Mm³/y) (Pétre et
494 al., 2016). The groundwater flux transmitted through the international border also corresponds to
495 the effective recharge rate of the MRA by assuming that this flux is solely due to the portion of
496 the potential recharge that actually reaches the aquifer. In the numerical model, an effective
497 recharge rate of 10 mm/y was applied on the outcrop area of the MRA to produce the simulated
498 transboundary flux. Besides, the steady-state hypothesis in the area south of the Milk River is
499 confirmed since the recharge flux is greater than the groundwater extraction and the simulated
500 drawdowns are small in this area. Under exploitation, the transboundary flux does not change
501 significantly.

502 **Table 7 Water balance of the MRA in Alberta under steady state conditions**

	Natural conditions	Scenario 1	Scenario 2	Scenario 3
Transboundary flux	16,320	16,390	16,448	16,459
Flux beyond the Milk River	1,414	1,497	1,555	1,576
Flux from overlying units	177	2,383	3,761	4,674
Flux to the bedrock valleys	-684	-297	-66	-81
Flux to the underling aquitard	-896	-861	-838	-816
Groundwater use	0	-2,639	-4,307	-5,357
Budget error	11	82.65	105.74	-3.87
Budget error (%)	0.70	2.13	1.99	-0.06

503
504 **Ground water flow interception by the Milk River.** The simulated flux intercepted by the Milk
505 River and its tributaries is about 14,906 m³/d (5.44 Mm³/y). This value represents 91.3% of the

506 incoming groundwater flux flowing from the south of the Milk River in the MRA. The numerical
507 model thus shows that the Milk River is effectively the main discharge feature of the MRA in
508 terms of magnitude. The groundwater flow interception slightly decreases with increasing
509 groundwater use. Indeed, the simulated flux transmitted beyond the Milk River increases from
510 1,414 m³/d (0.52 Mm³/y) in natural conditions to 1,576 m³/d (0.58 Mm³/y) in scenario 3.

511 **Cross-formational flow (vertical leakage).** In southern Alberta, north of the Milk River, the
512 numerical model simulates a downward vertical flow of 896 m³/d (0.33 Mm³/y) (directed from the
513 MRA to the Colorado Group) and an upward flow of 684 m³/d (0.25 Mm³/y) along the bedrock
514 valleys (directed from the MRA to the Pakowki/Claggett Formation). This discharge mechanism
515 of the MRA was proposed by Borneuf (1976) and Toth and Corbet (1986) but this flux had not
516 previously been quantified. Both cross-formational flow values are in the range of estimates from
517 the conceptual model formulated by (Pétre et al., 2016). Under exploitation of the MRA, the
518 cross-formational flow decreases. For instance, the flow to the bedrock valley decrease of about
519 89% in scenario 3 in comparison with natural conditions whereas there is an 8% decrease in the
520 downward flow to the underlying aquitard.

521 **Groundwater inflow from the overlying units.** The MRA receives a groundwater inflow of
522 about 177 m³/d (0.07 Mm³/y) from overlying units, especially in the vicinity of the topographic
523 highs (Cypress Hills). This result supports the statement from Toth and Corbet (1986), according
524 to which the MRA receives groundwater inflow from topographic highs in the study area. Under
525 exploitation of the MRA, the inflow from the overlying units compensate for most of the
526 groundwater extraction and reaches up to 4,674 m³/d (1.7 Mm³/y) in scenario 3.

527

528 **Groundwater storage in the MRA**

529 The estimated volume of water stored in the MRA ranges from 100 to 300 Mm³, using the range
530 of values for the storage coefficient of the MRA (1×10^{-4} to 3×10^{-4}) and knowing the total volume
531 of the MRA in the budget subdomain (1×10^{12} m³). The previous estimate from the conceptual
532 model was about 380 Mm³ of water stored in the MRA, using Meyboom's value for storage
533 coefficient (3×10^{-4}).

534
535 As mentioned in section 3.4 (Numerical model calibration criteria), the numerical groundwater
536 flow model was calibrated in a way that the direction and magnitude of the groundwater budget
537 components are in agreement with the conceptual model. Therefore, the comparison between
538 the simulated and estimated values under natural conditions shows that the conceptual model is
539 hydraulically plausible.

540 Particle tracking gives a representation of the groundwater flow pattern within the system, ESM6
541 shows both forward and backward particle tracking starting from the central nodes in the MRA
542 unit .

543
544 Under natural conditions, the MRA discharges to the surface especially along the coulees and
545 bedrock valleys as well as through the underlying Colorado Aquitard (ESM6a). Under
546 exploitation of the MRA (ESM6 b) c) and d)), the groundwater flow patterns significantly change
547 in the entire flow system. The discharge to the surface considerably decreases and the travel
548 time to the Colorado Aquitard increases, which illustrates the concept of capture to compensate
549 for the groundwater withdrawals in the MRA.

550 The quantification of the sources of groundwater abstracted over the exploitation period is a
551 crucial question in understanding how the MRA and the entire flow system adapt to groundwater
552 extraction. Table 8 provides an estimation of the sources of the groundwater abstracted over

553 the 108-year exploitation period. The first two columns correspond to a transient period of 108
 554 years during which an inflow from the storage depletion (estimated in Table 5) is considered.
 555 The groundwater flow coming from the change in flow patterns in and out of the MRA is obtained
 556 by subtracting the decrease in storage and the flux transmitted beyond the Milk River to the total
 557 exploitation. The range of plausible storage coefficient values is considered in this calculation.
 558 The last two columns of Table 8 describe the conditions simulated in the steady-state model and
 559 correspond to the new equilibrium reached by the aquifer system.

560 Under scenario 1 and using a storage coefficient of 1×10^{-4} , the loss of groundwater storage in
 561 the MRA (4.1 Mm^3) has provided 4 % of the water abstracted from the MRA over 108 years.
 562 This loss in storage represents about 4% of the total water stored in the MRA without
 563 exploitation (100 Mm^3). Direct inflow beyond the Milk River has contributed 59 Mm^3 (57%) of
 564 groundwater abstraction, and changes in flow patterns in and out of the MRA have contributed
 565 40.9 Mm^3 (39%). However, under the new steady-state conditions of mean MRA exploitation this
 566 change in flow patterns represent 46% of water abstraction from the MRA. Under scenario 2 and
 567 3 the loss of groundwater storage remains in the same proportions relatively to the total water
 568 abstracted whereas the contribution of the changes in the flow patterns in and out of the MRA
 569 increases. The volume of direct inflow beyond the Milk River does not change significantly with
 570 increased groundwater abstraction but its contribution increase as the flow system adjust to the
 571 pumping.

572 **Table 8 Sources of the abstracted groundwater in Alberta over 108 years under the three simulated**
 573 **exploitation scenarios. In the transient budget, the two figures in the decrease in storage and the**
 574 **proportion in changes in and out from the MRA corresponds to calculations using a storage**
 575 **coefficient value of 1×10^{-4} and 3×10^{-4} .**

Mean extraction rate (scenario 1)	Transient conditions over 108 years		New equilibrium conditions (steady-state)	
	Volumes (Mm^3)	Proportions (%)	Rate (m^3/d)	Proportions (%)
Total extraction	104.0	100	2,639.4	100.0
Decrease in storage	4.1/12.4	4.0/11.9	0	0.0
Flux beyond the Milk River	59.0	56.7	1,497.0	-56.7

Changes in out/inflow of MRA	40.9/32.6	39.3/31.4	1,225.1	-46.4
Maximum extraction rate (scenario 2)	Transient conditions over 108 years		New equilibrium conditions	
Total extraction	169.8	100.0	-4,307	100.0
Decrease in storage	6.5/19.6	3.9/11.6	0.0	0.0
Flux beyond the Milk River	61.3	36.1	1,555.4	-36.1
Changes in out/inflow of MRA	101.9/88.8	60.0/52.3	2,857.0	-66.3
Ultimate extraction rate (scenario 3)	Transient conditions over 108 years		New equilibrium conditions	
Total extraction	211.2	100.0	-5,357	100.0
Decrease in storage	8.3/24.8	3.9/11.7	0.0	0.0
Flux beyond the Milk River	62.1	29.4	1,576.0	-29.4
Changes in out/inflow of MRA	140.8/124.3	66.7/58.8	3,777.0	-70.5

576

577 Using a storage coefficient of 3×10^{-4} , the contribution from the loss of storage over 108 years
 578 increases to 12% in all three scenarios. As previously stated, the total volume of water stored in
 579 the MRA would be about 300 Mm³ in that case which means that the percentage of storage loss
 580 relative to the total volume of water stored in the MRA is the same (4%) as in the previous case
 581 ($S_s = 1 \times 10^{-4}$).

582 5. Discussion

583 5.1 Model limitations

584 The numerical groundwater flow model developed in the present study is a simplification of the
 585 real aquifer system. The following limitations and sources of uncertainties should be taken into
 586 consideration when examining model results.

587 A first source of uncertainty lies in the underlying geological model, which constitutes the basis
 588 of the groundwater flow model. The error associated with the geometry, thickness and structure
 589 of the hydrostratigraphic units are reflected in the groundwater flow model. This is especially true
 590 in the south-east part of the model in Montana, where the geological data were sparse (Pétre et
 591 al., 2015), thus increasing the degree of uncertainty in the geometry of the geological units.

592 The limits of the geological model further constrain the groundwater flow model, as it does not
593 extend in the Big Sandy Creek and Bear Paws Mountains areas in Montana due to a lack of
594 geological data. Therefore, the hydraulic head contours do not completely reflect the south-
595 west/north-east orientation of the potentiometric low along the Big Sandy Creek, which was
596 highlighted in the conceptual model. Collection of new data in this area would allow a better
597 representation of the groundwater flow pattern in the south-east corner of the study area.

598 The assumption that the potentiometric maps in Montana and south of the Milk River in Alberta
599 are representative of a steady-state situation might not be correct everywhere and could explain
600 the tendency towards overestimation of simulated heads south of the Milk River in Alberta and in
601 the south-east Montana. The south-east area in Montana combines the highest error associated
602 with the hydraulic conductivity of the MRA and a higher uncertainty concerning the geometry of
603 the geological unit. Thus, this is the less reliable area in the model domain.

604 The non-uniqueness problem occurs when different sets of parameters give equally good
605 models, in terms of matching the observations. The manual trial-and-error calibration done in the
606 present study does not rule out this issue. However, to address the non-uniqueness problem,
607 the spatial distribution of hydraulic conductivity of the MRA was obtained from measured data
608 and the hydraulic conductivity of the other hydrostratigraphic units was limited to a range
609 considered close to reasonable values. Besides, the sensitivity analysis (ESM5) shows that the
610 values used in the model lead to minimal errors and correspond to an optimal model calibration.

611 Another limitation of the model is that the steady-state simulations do not allow the determination
612 of the storage dynamics over time as well as the time required to reach an equilibrium. Finally,
613 the determination of the loss of storage from the confining units cannot be addressed in this
614 steady-state model.

615

616 **5.2 Implications for groundwater management**

617 The results of the steady-state model of the regional groundwater flow system encompassing
618 the MRA have some implications on the management of the groundwater resource. As seen
619 before, the numerical model shows that the conceptual model of the MRA is hydraulically
620 plausible by successfully representing the location of the groundwater divide in Montana as well
621 as the transboundary fluxes and the main components of the groundwater budget. Thus, the
622 numerical model confirms that the MRA is an internationally shared groundwater resource, with
623 two transboundary fluxes flowing from the recharge area in Montana to the north in Alberta.

624 It seems therefore appropriate to consider the implementation of a joint management of this
625 shared resource between Canada and the USA. Both the numerical and conceptual models of
626 the MRA allow the delineation of the proper management unit for such a transboundary
627 management. The appropriate management unit would be comprised between the north of the
628 groundwater divide in Montana and the south of the Canadian reach of the Milk River and
629 Verdigris Coulee (also defined as “zone 1a” in Pétré et al. (2016)).

630 As described in the “water balance” section, due to the major interception of the incoming
631 groundwater flow by the Milk River, the flux north of the Milk River is quite low (1,096 m³/d). This
632 groundwater flux represents the main external groundwater renewal mechanism of the MRA.
633 Thus, the area north of the Milk River only receives a small portion of the main recharge flux
634 from the MRA outcrop area whereas the area south of the Milk River benefits from the totality of
635 the transboundary flux coming from the south.

636 The numerical simulation under natural conditions showed that the MRA is part of a large
637 regional flow system involving important flow through the confining units. The numerical
638 simulation of this flow system supports conceptual work from Toth (1963) and Freeze and
639 Witherspoon (1966a, 1966b). Groundwater withdrawals in the MRA has a considerable impact

640 on the entire flow system. It is then necessary to consider the flow system as a whole to assess
641 the sustainability of the MRA exploitation. The flow system can adapt to large exploitation levels
642 in the MRA and reach a new equilibrium. However, this new equilibrium implies significant
643 impacts such as major drawdowns locally and a decrease of the water flow emerging to the
644 surface. The exploitation of the MRA leads to a decrease in groundwater storage. This loss of
645 storage is estimated by comparing conditions with and without exploitation but its transient
646 evolution cannot be followed due to the steady-state regime. It is possible that a large proportion
647 of this loss of storage will not be recoverable (Konikow and Neuzil, 2007). Groundwater use
648 scenario 2 corresponding to the maximum exploitation rate thus possibly better represents the
649 permanent loss of storage in the MRA.

650 The average historical level of exploitation of the MRA (scenario 1) appears to be sustainable
651 whereas scenario 3 leads to major drawdowns that would prevent the MRA exploitation, or make
652 it very expensive. The loss of artesian conditions would either require installing water pumps
653 where they were not required before or installing pumps deeper in wells, which could be
654 challenging economically and technically. Quality issues may also arise since the water coming
655 from the overlying units are more mineralized than water from the MRA.

656 Scenario 2 corresponds to intermediate conditions and would represent a limiting case of the
657 MRA exploitation as it still results in important regional drawdowns and captures a large
658 proportion of the water which originally flows back to the land surface.

659 **6. Conclusion**

660 A steady-state numerical model of the regional groundwater flow system comprising the
661 transboundary Milk River Aquifer (MRA) was developed with the objective of understanding the
662 impact of major and long-term exploitation on the entire aquifer system and to identify the
663 processes controlling the sustainable exploitation of the MRA. Results from this work has

664 implications on our general understanding of large flow systems and provides specific results on
665 the sustainable exploitation conditions of the MRA. Steady-state simulations of three
666 groundwater extraction scenarios demonstrate that groundwater withdrawals cause changes in
667 the entire flow system. Indeed, results indicate that loss of storage, less outflow and more inflow
668 supplied the groundwater that was extracted, illustrating the important role of capture in a
669 regional groundwater flow system. The quantification of cross-formational flow through aquitards
670 showed that the MRA is not an isolated hydrogeological unit but is rather part of a large
671 groundwater flow system. Sustainability of regional aquifers should therefore be defined by
672 taking into account the entire flow system. Overall, this work makes an important contribution to
673 major hydrogeological questions related to regional flow, especially through the aquitards, as
674 well as the impact of the exploitation on the flow system and the long-term reduction of
675 groundwater storage in a regional aquifer. Although this study provides interesting and relevant
676 results, the steady state numerical modelling did not allow the representation of storage
677 dynamics over time as well as the time required to reach an equilibrium.

678 Concerning the specific findings relative to the MRA sustainable exploitation, results show that
679 the numerical model is in agreement with the previously formulated conceptual model and thus
680 supports its hydraulic plausibility. Then, the mean historical extraction rate (scenario 1) appears
681 sustainable because the regional groundwater flow system can adapt and is able to reach a new
682 equilibrium. In contrast, the highest level of exploitation (scenario 3) does not seem sustainable
683 as it leads to dramatic drawdowns in the MRA and an important capture of cross-formational
684 flow. This level of exploitation would endanger the technical and economical exploitation of the
685 MRA and could possibly threaten the beneficial use of surface water from the coulees. The
686 intermediate scenario 2 corresponds to a borderline condition which should not be exceeded.
687 The numerical model also illustrates that the MRA is a transboundary groundwater resource.
688 Thus, an internationally shared management strategy of the MRA would be warranted,

689 especially in the area comprised between the groundwater divide in Montana and the southern
690 reach of the Milk River in Alberta. Finally, future modelling would have to accompany
691 management decisions to assess future exploitation scenarios.

692

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862 8. Electronic supplement material

863 **ESM1. Hydraulic conductivity map of the MRA**

864 The K map of the MRA (Figure 10) shows a wide range of K values from 8×10^{-9} to 9×10^{-4} m/s
865 with higher values concentrated in the middle of the domain. This distribution is associated with
866 faulting and fracturing around the igneous intrusion of the Sweet Grass Hills as well as structural
867 deformation in the south-west corner of the study area (Tuck, 1993; Zimmerman, 1967).

868 **ESM2. Model calibration levels map**

869 To better visualize the spatial distribution of errors, the level of calibration was calculated for
870 each calibration target. The highest degree of calibration (level 1) indicates that the simulated
871 heads fall within the calibration target (± 30 m). Levels 2 and 3 correspond respectively to
872 simulated values that fall within two and three times the associated error of the calibration target
873 (i.e. ± 60 and 90 m) (Anderson and Woessner 1992). Table 9 shows the levels of calibration in
874 each calibration zone of the model and the spatial distribution of the levels of calibration for each
875 observation point is shown in Fig. 11. In southern Alberta, the calibration targets derived from
876 the potentiometric surface of Toth and Corbet (1986) show the highest calibration level. The
877 lower levels of calibration in Montana can be explained by the uncertainties of the underlying
878 geological model in this area. Indeed, the geological data are scarce in Montana and the
879 resulting error in the geometry of the hydrostratigraphic units can lead to a decrease in the
880 representative nature of the model and thus of its calibration performance. Besides, the south-
881 east area in Montana has the largest uncertainties associated with water level observation
882 points. These observation points are derived from a potentiometric map drawn at the regional
883 scale that includes only a few isolines in south-east Montana (Levings 1982). The uncertainty is
884 then higher in south-east Montana, whereas observation points derived from small scale
885 potentiometric maps such as Tuck (1993) in the Sweetgrass Hills area or Zimmerman (1967) in

886 the south-west Montana are more accurate. The lowest level of calibration (level 3) is found on
 887 the flanks of the eastern butte of the Sweetgrass Hills. The steep topography in this area
 888 explains the difficulty in obtaining a satisfying calibration.

889

890 **Table 9 Levels of calibration for the simulated heads in the Milk River Aquifer**

Calibration zones	Total number of observation points	Levels of calibration					
		Level 1		Level 2		Level 3	
		Number of obs. points	%	Number of obs. points	%	Number of obs. points	%
Montana, south-west	38	33	86.8	5	13.2	0	0.0
Montana, south-east	72	46	63.9	23	31.9	3	4.2
Montana, Sweetgrass Hills area	22	19	86.4	3	13.6	0	0.0
Alberta, south of the Milk River (Toth and Corbet 1986)	80	73	91.2	7	8.8	0	0.0
Alberta, south of the Milk River (Meyboom 1960)	80	56	70	24	30	0	0.0

891

892

893 **ESM3. Simulated heads contour map**

894 The calibration performance was also qualitatively assessed with the comparison of contour
 895 maps between observed and simulated heads, as shown in Fig. 12. This map shows that the
 896 regional groundwater flow patterns simulated and observed are similar in northern Montana and
 897 south of the Milk River in southern Alberta. More specifically, the groundwater divide north of Cut
 898 Bank, Montana, is well reproduced by the model. The radial pattern from the Sweetgrass Hills
 899 area to the north, east and south-east is satisfactorily replicated. In Alberta, the observed
 900 contour map is from Toth and Corbet (1986), as their interpretation of the heads distribution is in
 901 closer agreement with simulated heads.

902 **ESM4. Simulated artesian conditions**

903 Insert Figure 13 here

904

905

906

907 **ESM5. Parameter sensitivity analysis**

908 A sensitivity analysis was carried out to test the effect of change in parameters in achieving the
909 calibration result on the RMSE and on the number of calibration targets with a level 1. The
910 sensitivity of the calibration is tested by changing several parameters (one at a time) from the
911 reference case (the calibrated model without pumping). The RMSE and the percentage of level 1
912 calibration targets were calculated for each simulation. The magnitude of change in RMSE and
913 level 1 calibration target is a measure of the sensitivity of the solution to that particular
914 parameter.

915 Fig. 14 shows the effect of change in the recharge rate (a), the vertical anisotropy of the Belly
916 River (b) and Pakowki formations (c), the hydraulic conductivity of the Pakowki Formation (d),
917 the bedrock valleys (e) and the Milk River Formation in the south-east corner of the study area
918 (f) on the RMSE and the number of level 1 calibration targets.

919 Recharge rates were modified in both the outcrop area of the MRA and the remaining of the
920 domain. The recharge rate in the aquitard outcrop area remains unchanged (0 mm/y). Results
921 are shown for the different calibration zones to illustrate areas of the model domain that are most
922 sensitive to variations of a specific parameter.

923 The calibration zones in south-west Montana and in the Sweet Grass Hills are the least
924 sensitive to tested parameters, whereas the south-east Montana and the area south of the Milk
925 River in Alberta are the most sensitive.

926 In the calibration process, a compromise was reached between the minimization of the RMSE
927 and the maximisation of the number of flowing artesian nodes north of the study area. Therefore,
928 the calibrated value does not always minimize the RMSE in the model, rather it reflects the
929 adjustment between a satisfying representation of the artesian conditions and the minimisation
930 of the RMSE. For example, a recharge rate increased by a factor of 2 or 4 will increase both the
931 artesianism and the RMSE. Therefore, the number of level 1 calibration targets is lowered in all
932 calibration zones except in south-west Montana. When the recharge rate is lowered by a factor 2
933 the RMSE slightly decreases whereas the decrease in artesianism is significant (more than
934 10%).

935 The same objective was followed when testing the vertical anisotropy of the Pakowki/Claggett
936 and Belly/Judith River formations. The area south of the Milk River in Alberta and the area in
937 south-east Montana are the most sensitive to the change in the vertical anisotropy.

938 The change in artesianism is especially significant (between 8 and 10%) when varying the
939 hydraulic conductivity of the bedrock valleys and the vertical anisotropy of the Belly/Judith River
940 Formation. When the hydraulic conductivity of the bedrock valleys is higher than the calibrated
941 value, the flowing artesianism tends to disappear and the buried valleys do not function as
942 drains anymore. The vertical flux between the buried valleys and the MRA is even reversed
943 (directed downward from the buried valleys).

944 The performance of the model is sensitive to the change in the hydraulic conductivity of the
945 Pakowki Formation in south-east Montana and southern Alberta only. This result was expected
946 as the Pakowki Formation is little or not present at all in the other calibration areas of the model.

947 Changing the hydraulic conductivity of the Milk River Formation in the south-east corner of the
948 study area in Montana only affects the performance of the model in that specific area.

949

950 **ESM6. Particle tracking**

951 **Insert Figure 15 here**

952

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956

957 **Fig. 1. Study area and limits of the numerical groundwater flow model (continuous black line)**

958 **Fig. 2. Conceptual hydrogeological model of the MRA. Location of the cross-section AA' is shown**
959 **on Fig. 1. (modified from Pétré et al. 2016)**

960 **Fig. 3. MRA 3D numerical model with geological units. Vertical exaggeration is 30 times.**

961 **Fig. 4. Boundary conditions in the model domain (h is hydraulic head and z is elevation relative to**
962 **sea level): (a) specified head at the surface of the model (top of layer 1); (b) specified head at the**
963 **perimeter of the Bow Island sandstone (layer 14); (c) zones of specified recharge in the MRA**
964 **outcrop/subcrop area (10 mm/y) and elsewhere over the model surface (1 mm/y), except where the**
965 **Colorado Group aquitard outcrops (0 mm/y); (d) summary of the boundary conditions imposed in**
966 **the model.**

967 **Fig10. Hydraulic conductivity of the Milk River Aquifer (derived from Pétré et al. 2016)**

968 **Fig. 5. Delineation of the six MRA exploitation zones, along with the location of the water wells**
969 **completed within the MRA and the extent of the subdomain budget (blue line).**

970 **Fig. 6. Location of the MRA water-level observation points in the model domain. A distinction is**
971 **made between the observation points derived from historical potentiometric maps (squares) and**
972 **those corresponding to monitoring wells (triangles). Recent pressure measurements are indicated**
973 **(diamonds). Prior to MRA exploitation, flowing artesian conditions were found east and north of**
974 **the line defined by Dowling (1917).**

975 **Fig. 7. Simulated versus measured heads in the Milk River aquifer (a) for the calibration zones (Fig.**
976 **6). The calibration target interval (+/- 30 m) is also shown (grey dashed lines). The calibration**
977 **targets derived from monitoring wells are indicated as "data" in the legend. The measured heads**
978 **in southern Alberta are from Toth and Corbet's (1986) interpretation. b Simulated heads**
979 **comparison between the two interpretations of the 1960 potentiometric surface in Alberta (Toth**
980 **and Corbet, 1986 and Meyboom, 1960) in the calibration zone located south of the Milk River in**
981 **Alberta**

982 **Fig. 13. Artesian conditions in the Milk River Formation (simulated natural conditions without**
983 **exploitation)**

984 **Fig. 8. Simulated drawdown (m) in the MRA caused by the three exploitation scenarios (Table 1):**
985 **(a) Scenario 1 (mean extraction); (b) scenario 2 (maximum extraction); (c) scenario 3 (ultimate**
986 **extraction) and (d) approximate water level change (AGRA Earth and Environmental Limited, 1998)**
987 **derived from observed water levels from late 1950s to late 1980s-1990s.**

988 **Fig. 9. Comparison of groundwater budget components: (a) between the conceptual model (Pétre**
989 **et al. 2016) and the numerical model (simulated natural conditions); (b) between several**
990 **groundwater use scenarios in the area north of the Milk River. Positive figures are inflow to the**
991 **domain whereas negative values are outflow (loss from the budget domain).**

992 **Fig. 15. Particle tracking from and to the central nodes of the MRA and artesianism in the MRA (a)**
993 **natural conditions; (b) mean extraction (scenario 1); (c) maximum extraction (scenario 2) and (d)**
994 **ultimate extraction (scenario 3). Location of the cross section B-B' is indicated on Fig. 1**

995 **Fig. 11. Spatial distribution of hydraulic head calibration levels for the available observation points**
996 **in the MRA**

997 **Fig. 12. Comparison of observed and simulated potentiometric surfaces.**

998 **Fig. 14. Parameter sensitivity analysis of the RMSE and calibration level 1 (for each calibration**
999 **zone) to the following parameters: (a) Recharge rate; (b) Anisotropy ratio of the Belly River**
1000 **Formation; (c) Anisotropy ratio of the Pakowki Formation; (d) Hydraulic conductivity of the**
1001 **Pakowki Formation; (e) Hydraulic conductivity of the bedrock valleys; (f) Hydraulic conductivity of**
1002 **the MRA in the south-east corner of the study area**

1003

1004 HIGHLIGHTS

- 1005 • Goal to understand mechanisms controlling sustainability through numerical modeling
- 1006 • Groundwater withdrawals lead to flow pattern changes over the entire flow system
- 1007 • Loss of storage, capture and cross-formational flow compensate withdrawals
- 1008 • The mean historical level of exploitation of the MRA appears sustainable

1009

1010

1011 Declaration of interests

1012

1013 The authors declare that they have no known competing financial interests or personal
1014 relationships that could have appeared to influence the work reported in this paper.

1015

1016 The authors declare the following financial interests/personal relationships which may be considered
1017 as potential competing interests:

1018



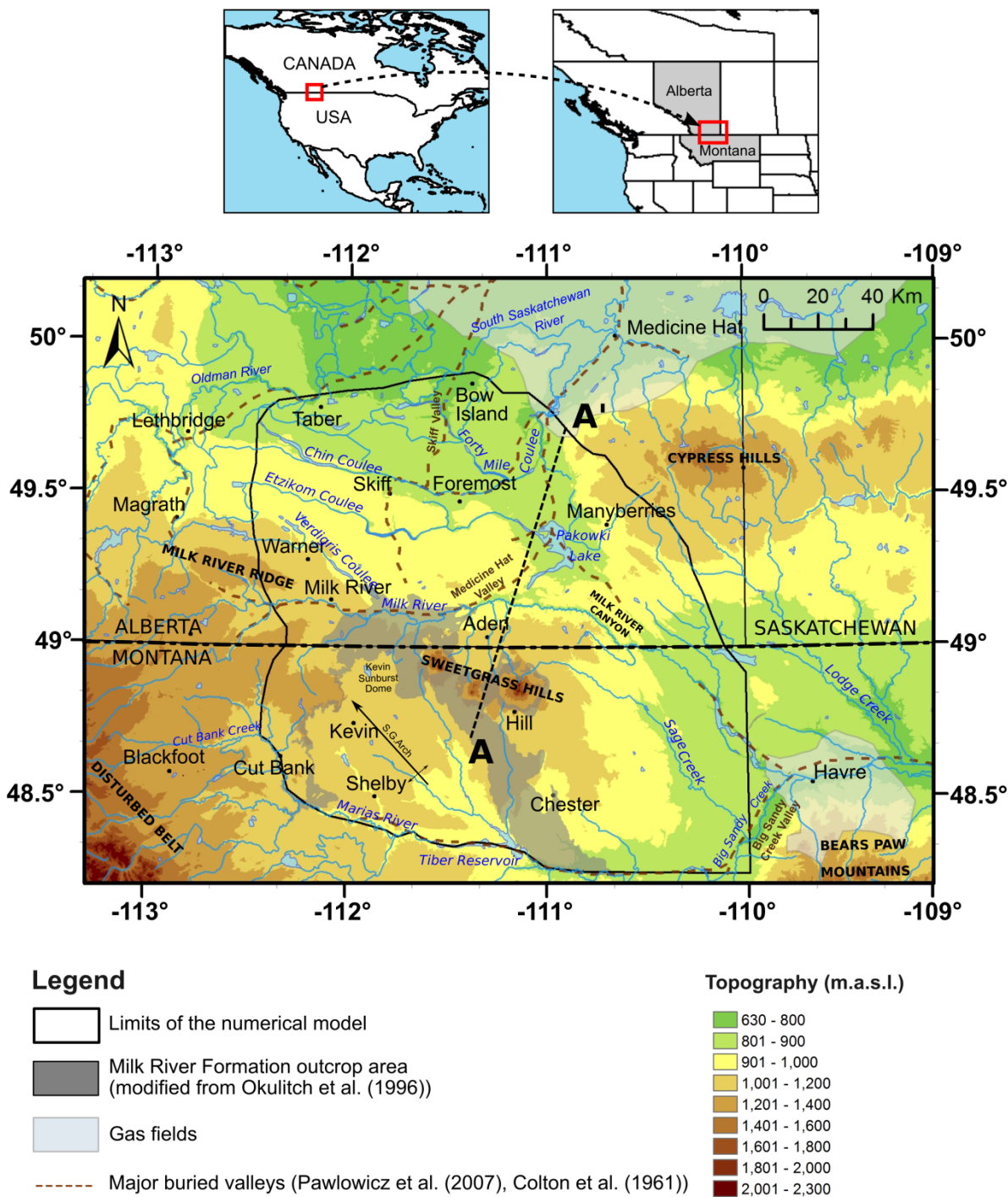
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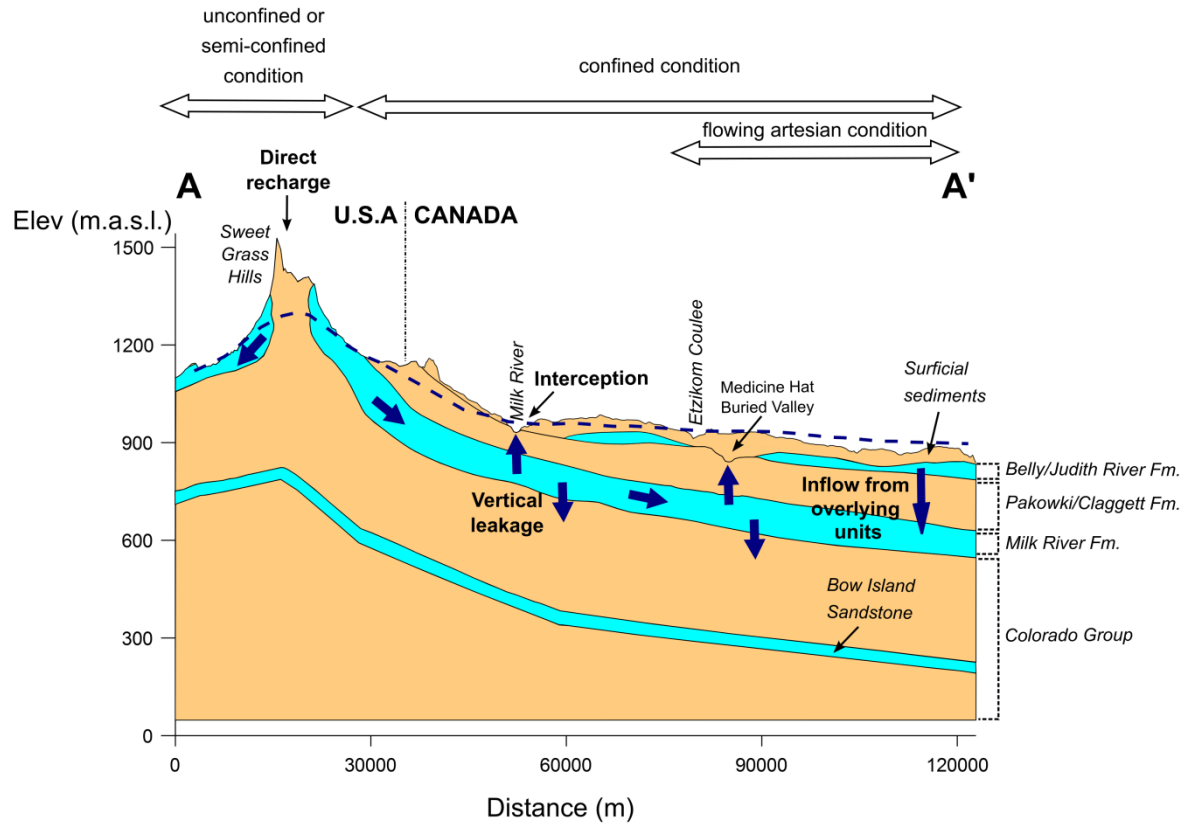
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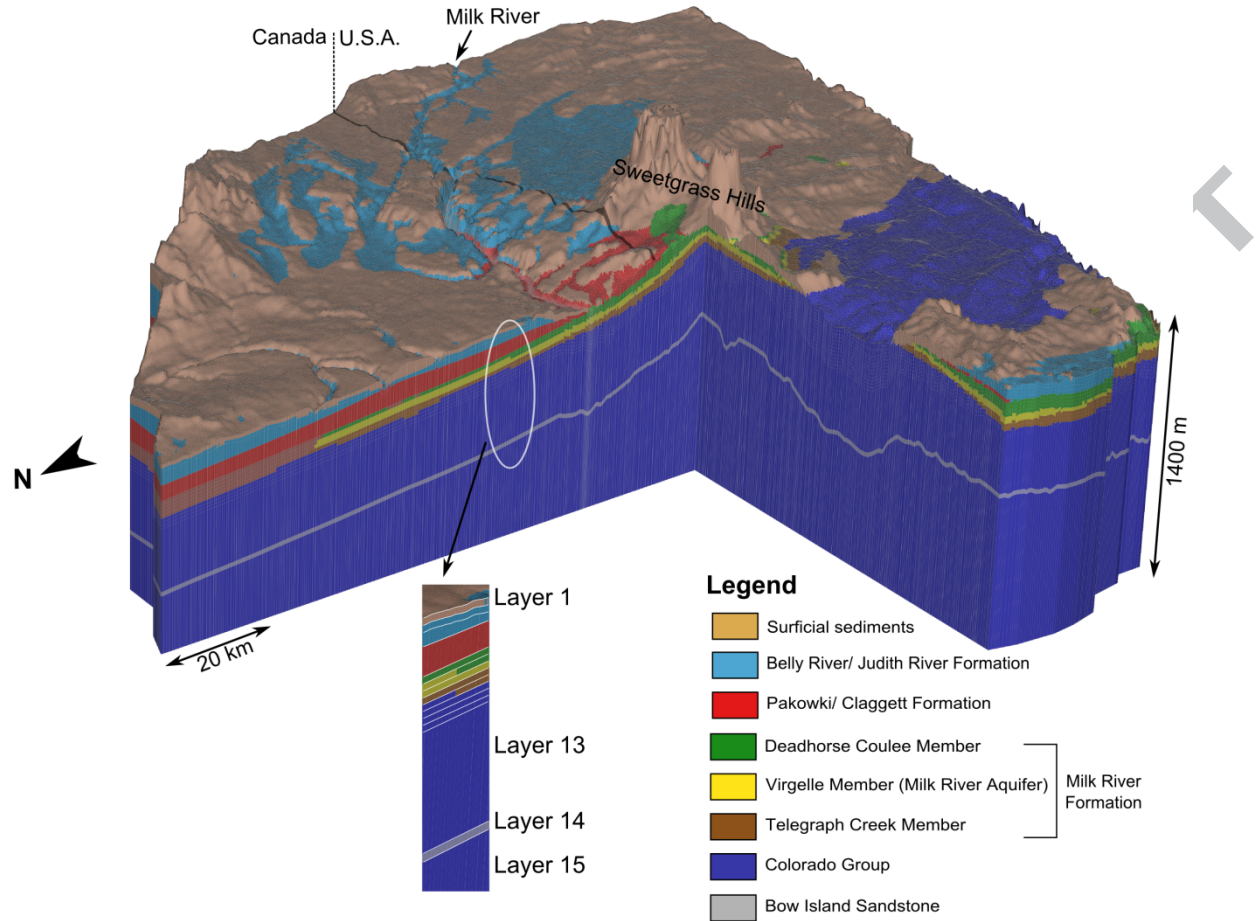
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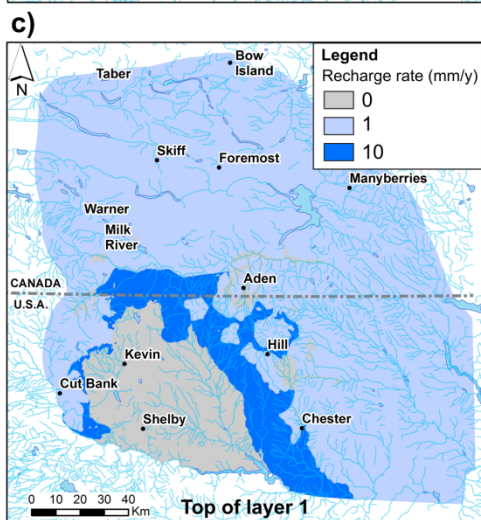
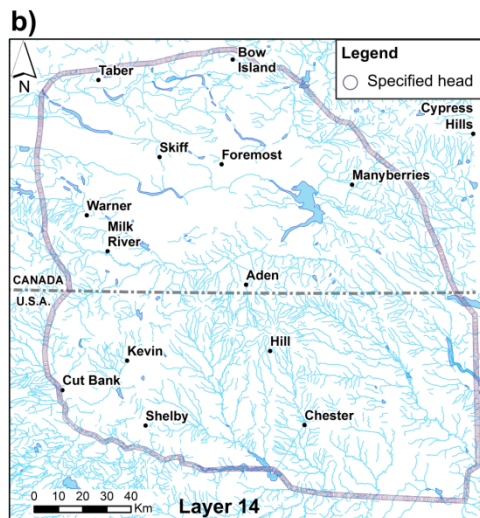
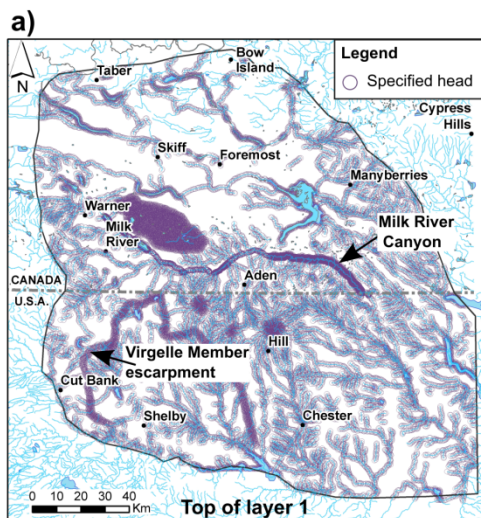


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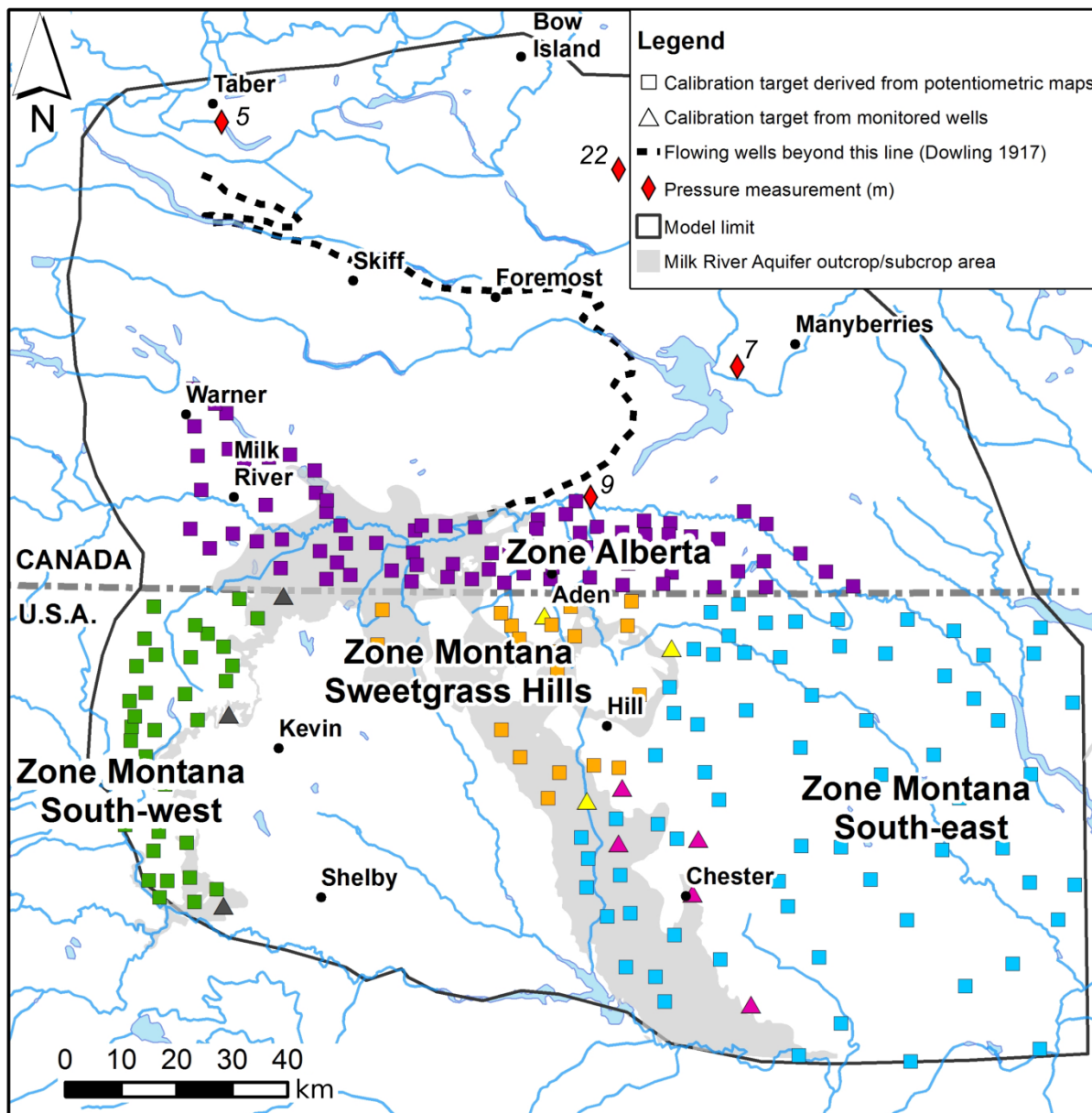
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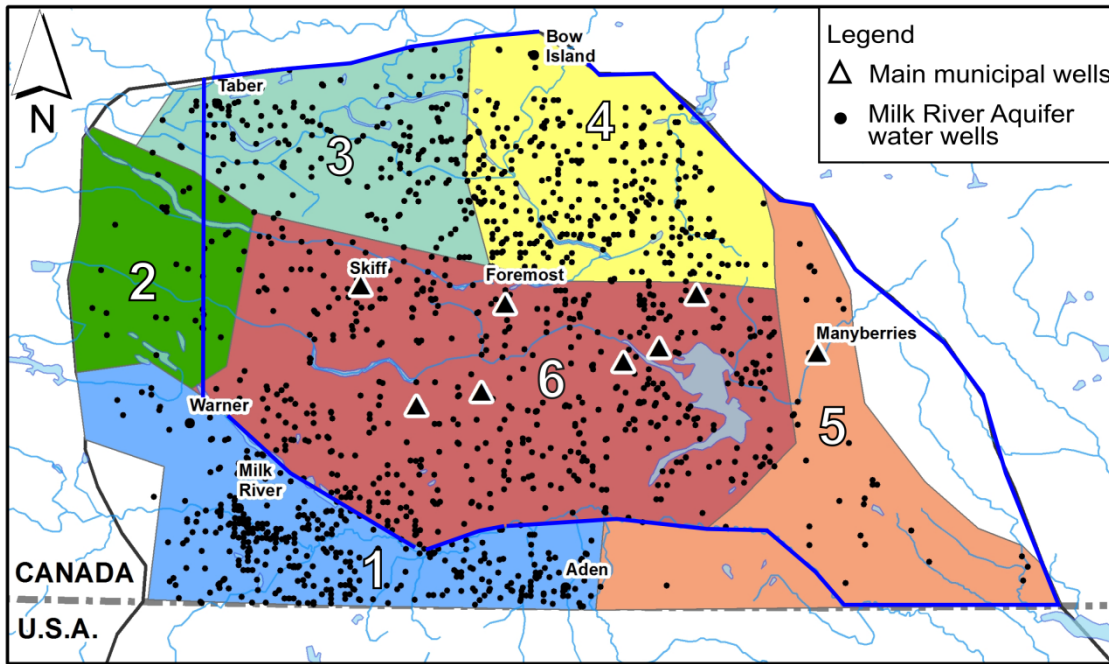
Slice	Boundary type	Boundary condition
Top of layer 1 (Surface)	Specified head with seepage constraint	- Surface drainage: $h=z$ - Virgelle member escarpment: $h= z-3$ m - Milk River Canyon: $h= z-30$ m
	Recharge	- MRA outcrop area: $R= 10$ mm/y - Aquitard outcrop area: $R= 0$ mm/y - Remaining of study area: $R= 1$ mm/y
Layer 14 (Bow Island Sandstone)	Specified head	- Perimeter of layer 14: $h= 750$ m

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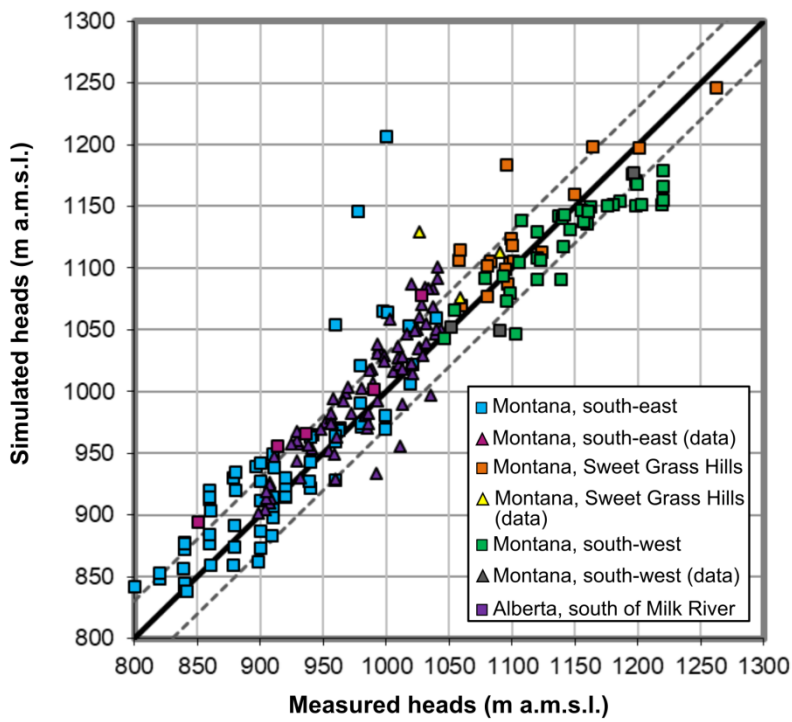


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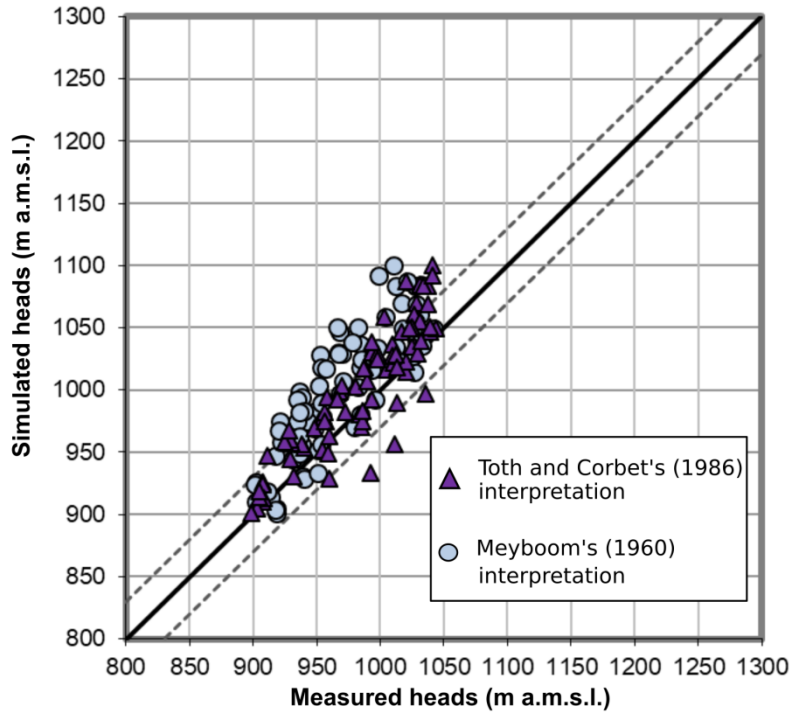
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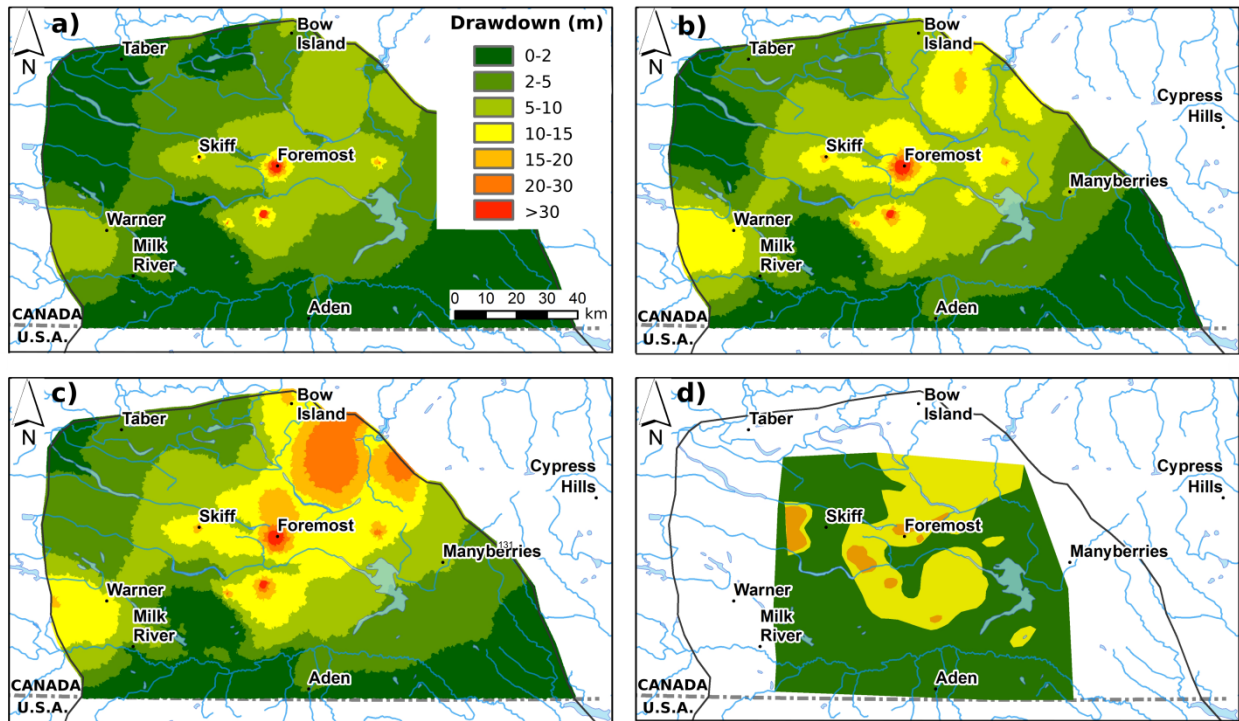
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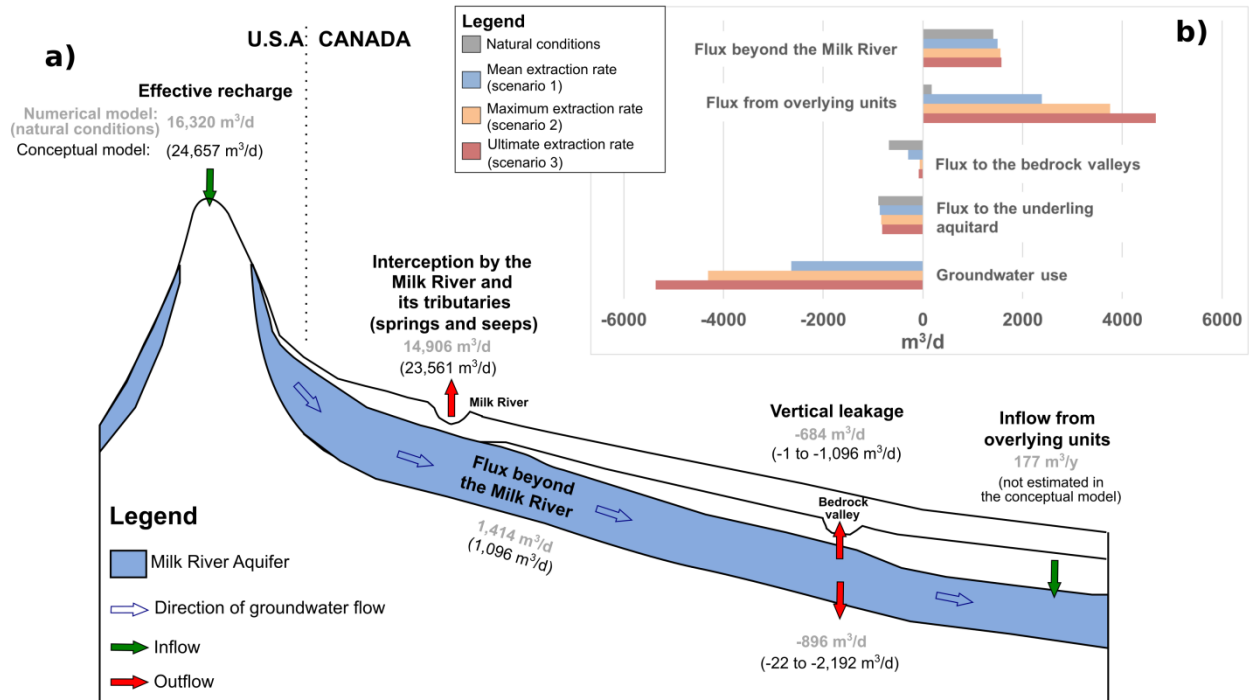
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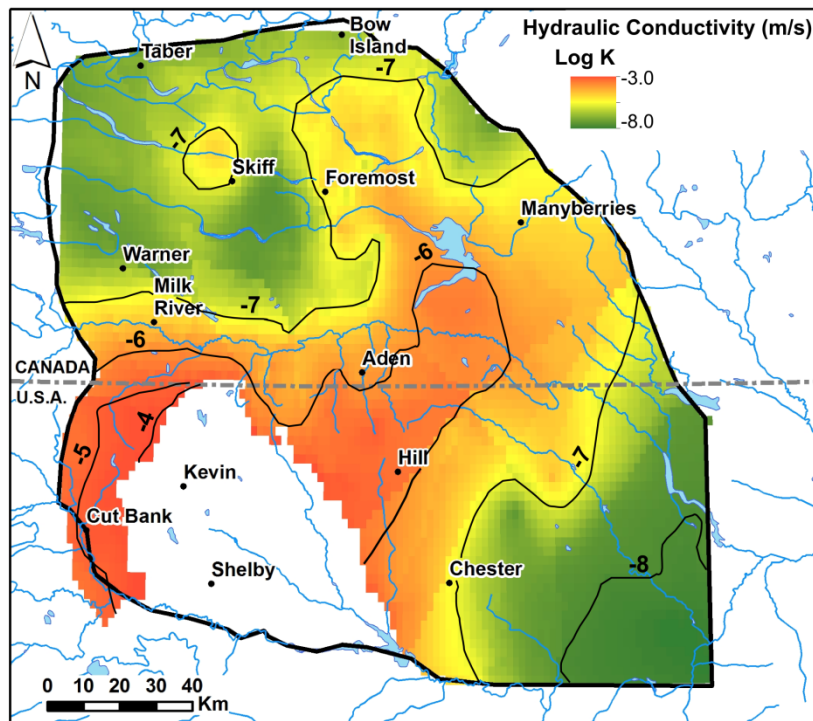
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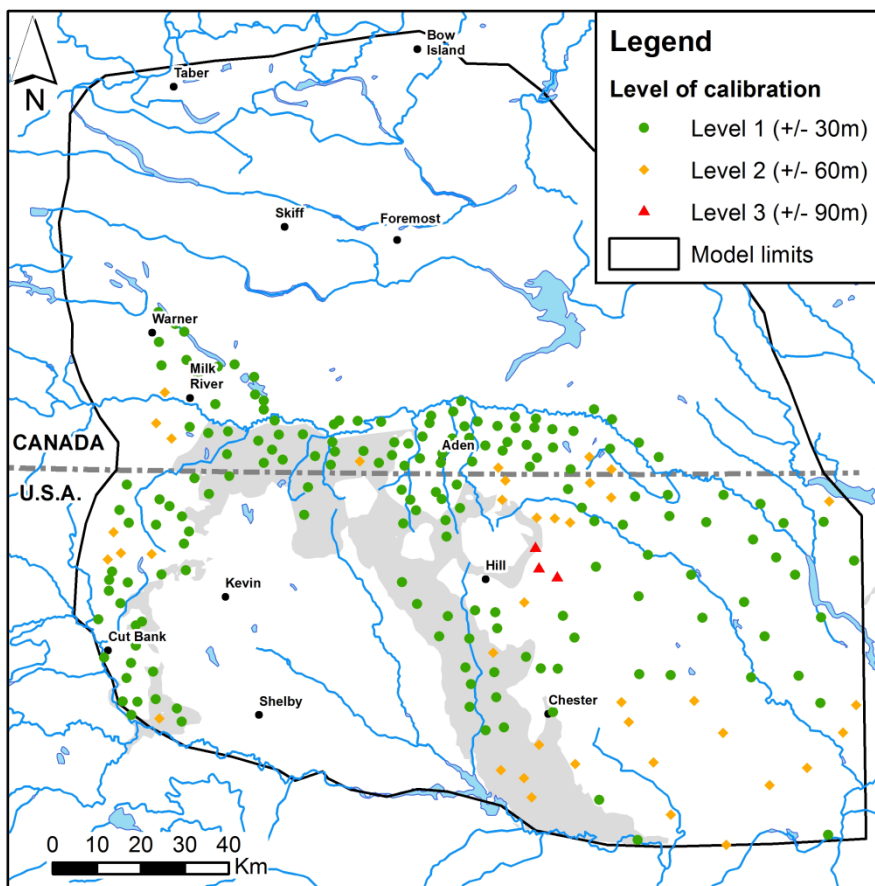
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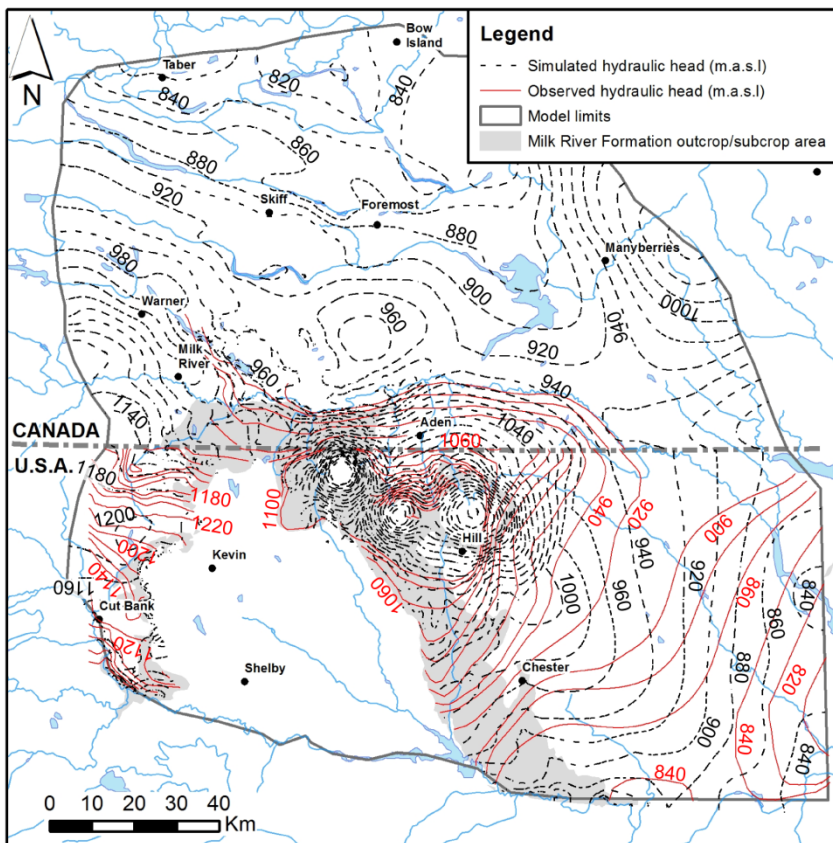
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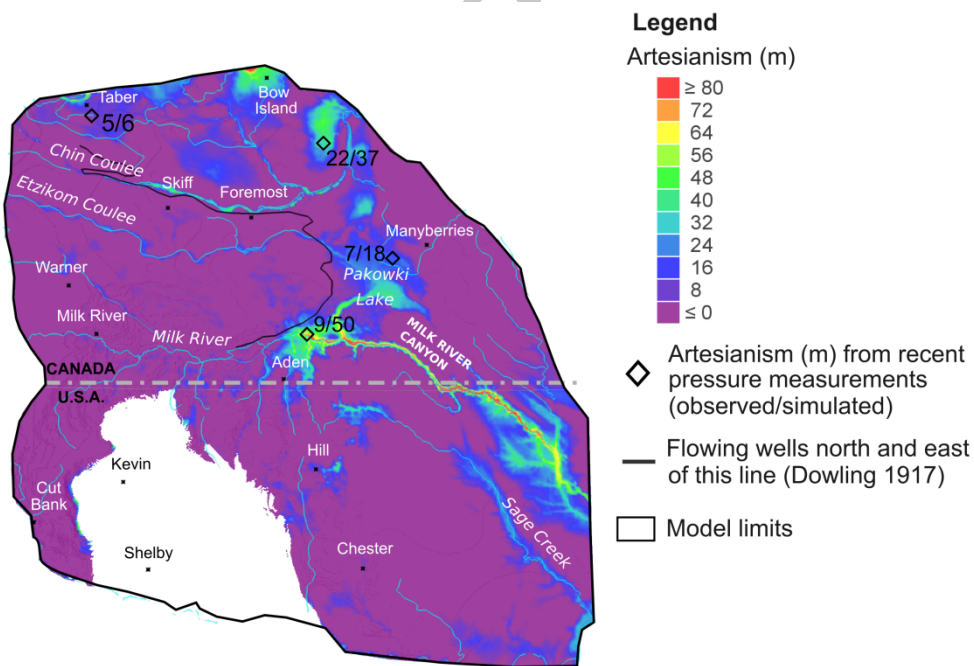
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1035

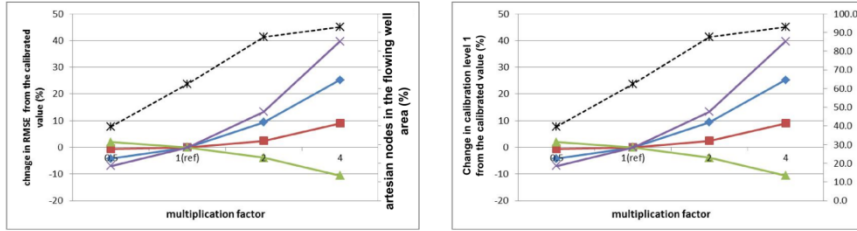


1036

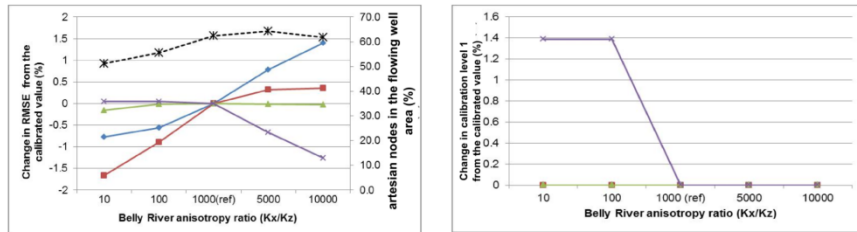


1037

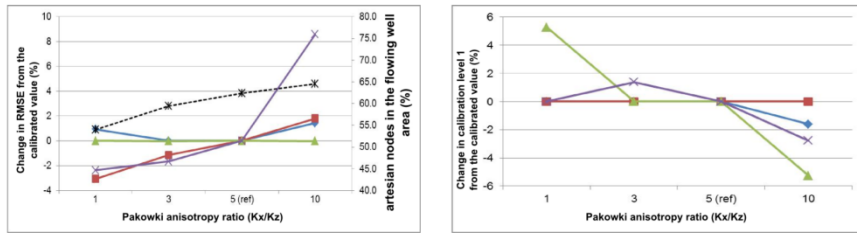
a) Recharge rate



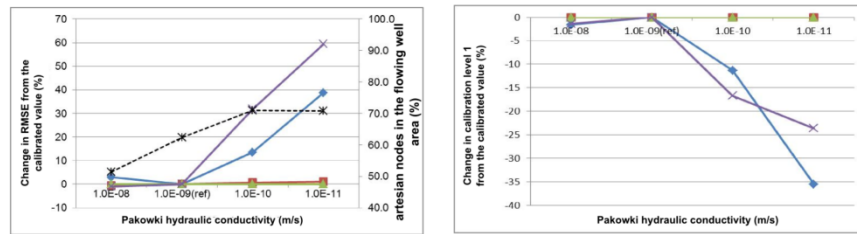
b) Vertical anisotropy of the Belly River Formation



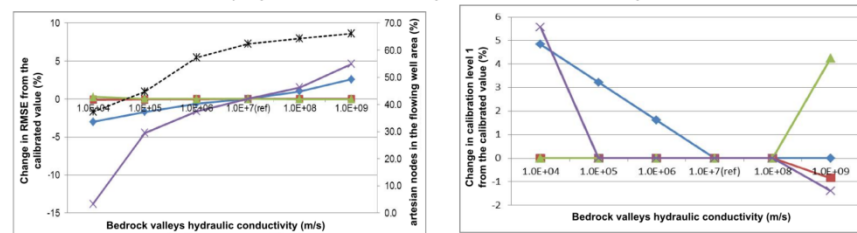
c) Vertical anisotropy of the Pakowki Formation



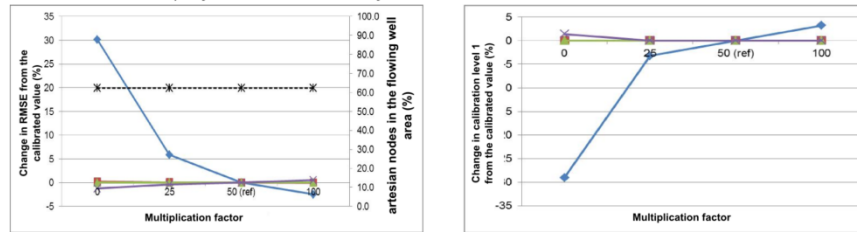
d) Hydraulic conductivity of the Pakowki Formation



e) Hydraulic conductivity of the Bedrock Valleys



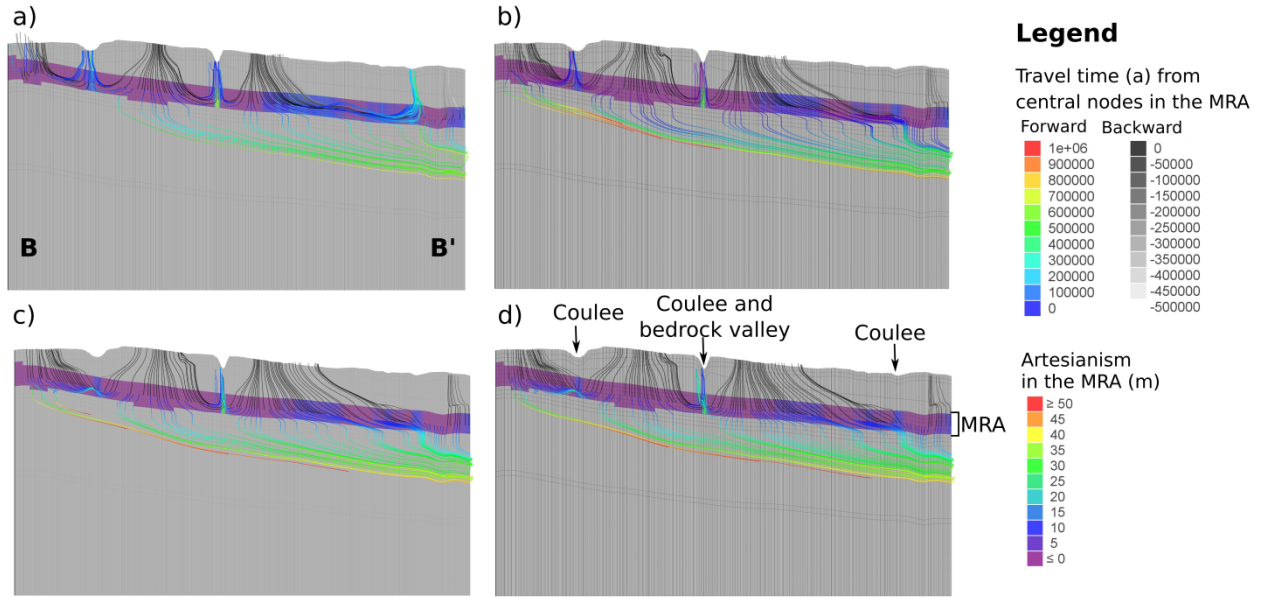
f) Hydraulic conductivity of the MRA in South-east Montana



Legend

- South-east Montana
- Sweetgrass Hills Montana
- ▲— South-west Montana
- ◆— Alberta (south of the Milk River)
- *--- Artesianism

SCRIPT



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ACCEPTED MANUSCRIPT