

# Accepted Manuscript

Subcellular distributions of trace elements (Cd, Pb, As, Hg, Se) in the livers of Alaskan yelloweye rockfish (*Sebastes ruberrimus*)

Benjamin D. Barst, Maikel Rosabal, Paul E. Drevnick, Peter G.C. Campbell, Niladri Basu



PII: S0269-7491(18)30204-5

DOI: [10.1016/j.envpol.2018.06.077](https://doi.org/10.1016/j.envpol.2018.06.077)

Reference: ENPO 11274

To appear in: *Environmental Pollution*

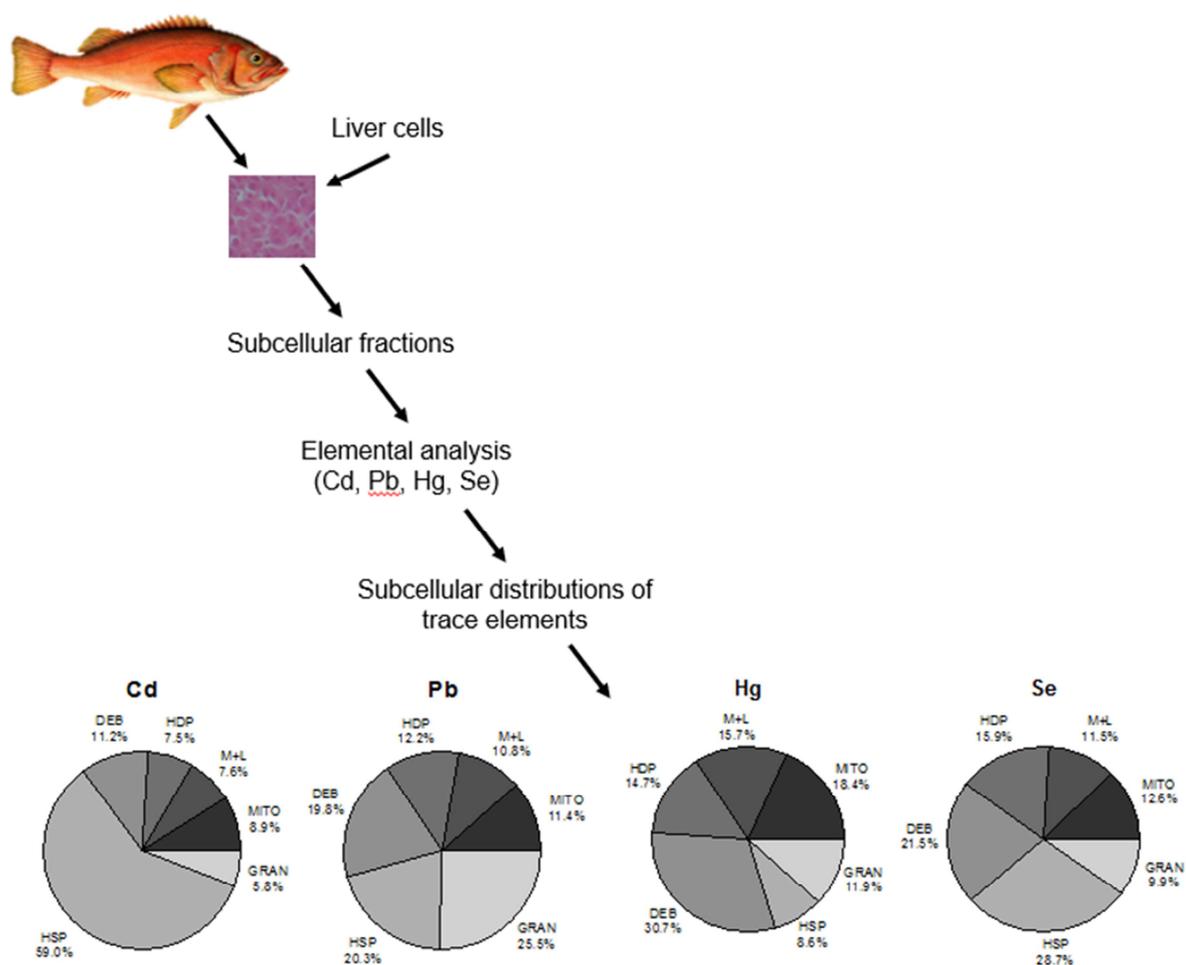
Received Date: 15 January 2018

Revised Date: 29 May 2018

Accepted Date: 22 June 2018

Please cite this article as: Barst, B.D., Rosabal, M., Drevnick, P.E., Campbell, P.G.C., Basu, N., Subcellular distributions of trace elements (Cd, Pb, As, Hg, Se) in the livers of Alaskan yelloweye rockfish (*Sebastes ruberrimus*), *Environmental Pollution* (2018), doi: 10.1016/j.envpol.2018.06.077.

This is a PDF file of an unedited manuscript that has been accepted for publication. As a service to our customers we are providing this early version of the manuscript. The manuscript will undergo copyediting, typesetting, and review of the resulting proof before it is published in its final form. Please note that during the production process errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.



1 Subcellular distributions of trace elements (Cd, Pb, As, Hg, Se) in the livers of  
2 Alaskan yelloweye rockfish (*Sebastes ruberrimus*)  
3

4 Benjamin D. Barst<sup>a,b,\*</sup>, Maikel Rosabal<sup>c</sup>, Paul E. Drevnick<sup>a,d</sup>, Peter G.C. Campbell<sup>a</sup>, and Niladri Basu<sup>b</sup>  
5  
6

7 <sup>a</sup> Institut national de la recherche scientifique, Centre Eau Terre et Environnement (INRS-ETE), 490 de  
8 la Couronne, Québec, QC, Canada G1K 9A9

9 <sup>b</sup> Faculty of Agricultural and Environmental Sciences, McGill University, Montreal, QC, Canada, H9X  
10 3V9

11 <sup>c</sup> Département des sciences biologiques, Université du Québec à Montréal (UQÀM), Montréal, QC,  
12 Canada, H2X 1Y4

13 <sup>d</sup> Environmental Monitoring and Science Division, Alberta Environment and Parks, Calgary, AB Canada,  
14 T2E 7L7

15  
16  
17  
18  
19  
20  
21 \*Corresponding author: Tel.: +1 514 216 6019; benjamin.barst@mcgill.ca  
22  
23  
24  
25

26 **Abstract**

27 Yelloweye rockfish (*Sebastes ruberrimus*) is an extremely long-lived species (up to ~120 years) of fish,  
28 which inhabits the coastal waters of Alaska. Due to their long lifespans, yelloweye are known to  
29 accumulate high levels of mercury, and potentially other trace elements, in their tissues. Relatively little is  
30 known about the subcellular distribution of trace elements in the tissues of yelloweye rockfish; such  
31 information can provide important insights into detoxification/toxicity mechanisms at the subcellular  
32 level. To address this, we collected yelloweye rockfish (n=8) from the eastern coast of Prince of Wales  
33 Island, Alaska in 2014. We determined the subcellular partitioning of trace elements (cadmium (Cd), lead  
34 (Pb), arsenic (As), total mercury (Hg), and selenium (Se)) in yelloweye livers with a partitioning  
35 procedure designed to separate liver cells into putative metal-sensitive fractions (cytosolic enzymes,  
36 organelles) and detoxified metal fractions (metallothionein or metallothionein-like proteins and peptides,  
37 granule-like structures) using differential centrifugation, NaOH digestion, and heat denaturation steps.  
38 The resulting fractions were then analyzed for total Hg with a direct Hg analyzer and for trace element  
39 concentrations by inductively coupled plasma-mass spectrometry (ICP-MS). For Cd, Pb, and As, the  
40 greatest contributions were found in the detoxified fractions, whereas the majority of total Hg was found  
41 in sensitive fractions. Selenium, an essential trace element, was distributed to a similar degree between  
42 the sensitive and detoxified compartments. Results indicate that although yelloweye sequestered and  
43 immobilized potentially toxic metals in detoxified fractions, the extent of binding differed among metals  
44 and followed the order: Cd > As > Pb > Hg. In yelloweye rockfish livers, the accumulation of non-  
45 essential elements at sensitive sites could lead to deleterious effects at the subcellular level, which should  
46 be evaluated in future studies.

47 **Keywords: subcellular partitioning; trace elements; mercury; detoxification; yelloweye rockfish;**

48 **Alaska**

49 **Capsule: Subcellular partitioning of yelloweye rockfish livers aids in understanding**  
50 **detoxification of trace elements**

## 51 1. Introduction

52 Yelloweye rockfish (YR; *Sebastes ruberrimus*) are one of the largest species of rockfish inhabiting the  
53 marine waters of western North America, where their range extends from the Aleutian Islands to the Baja  
54 Peninsula (Love et al., 2002). As their name would suggest, they often pass most their considerable  
55 lifespans (up to ~120 years) near steep rock piles on the ocean floor. Their long lifespans, large body size,  
56 and late age at sexual maturity render them particularly susceptible to both recreational and commercial  
57 fishing pressures. Currently, YR are listed as threatened in the Puget Sound-Georgia Basin of the United  
58 States (NMFS, 2010) and as a species of special concern in Canada (COSEWIC, 2009). In addition to  
59 overfishing, YR may be at risk from exposure to contaminants, as their tissues are known to contain  
60 elevated concentrations of mercury (Hg) (Barst et al., 2015). For example, total Hg concentrations in the  
61 edible muscle tissue of YR often exceed  $0.5 \mu\text{g g}^{-1}$  wet weight (ww), the level at which sublethal effects  
62 in fish are likely to occur (Sandheinrich and Wiener, 2011).

63 Despite reports of contamination of YR tissues, the associated health effects have remained  
64 largely unexplored, as is the case for many wild species of fish. An exception, by Barst et al. (2015),  
65 compared concentrations of essential (selenium (Se), copper (Cu), zinc (Zn)) and non-essential trace  
66 elements (nickel (Ni), cadmium (Cd), mercury (Hg)) with the relative areas of melano-macrophage  
67 aggregates (MA) in YR livers. Melano-macrophage aggregates are collections of immune cells that serve  
68 to store and process the products of cell breakdown, and are considered a general biomarker of  
69 contaminant exposure in fish (Wolke, 1992). An increase in MA area is often interpreted as an indication  
70 of tissue damage. In YR livers, the relative areas of MA increased with increasing hepatic concentrations  
71 of Hg, Se, Cd, and Cu, and these elements tended to be more concentrated in MA than the surrounding  
72 tissues. The accumulation of non-essential metals in the MA of fish may indicate increased cell turnover  
73 due to metals exposure. Interestingly, Hg and Se accumulated in MA to a similar extent, suggesting that  
74 the two were present as a mercury selenide complex (Barst et al., 2015). The biological interaction of Hg  
75 and Se has been well documented (Khan and Wang, 2009; Wang et al., 2012), and HgSe is widely  
76 regarded as a non-bioavailable end-product of Hg detoxification in the organs of different species (Korbas

77 et al., 2010; Palmisano et al., 1995). Furthermore, Se may also protect against Cd toxicity in wild fishes  
78 (Ponton et al., 2016).

79         With this in mind, the determination of the subcellular distributions of trace elements may be  
80 useful in determining the likelihood for toxic effects in the livers of YR, and may provide an overall  
81 greater understanding of trace element partitioning in wild fish. Subcellular partitioning allows for the  
82 distinction between metal-binding to potentially sensitive target molecules (e.g., cytosolic enzymes) and  
83 organelles (e.g., mitochondria), where the binding of non-essential metals may lead to negative effects,  
84 and metal accumulation in detoxified metal fractions (e.g., heat-stable proteins and metal-rich granules),  
85 which may minimize toxic effects (Campbell and Hare, 2009; Wallace et al., 2003). In this context, we  
86 determined the subcellular partitioning of Cd, Pb, As, total Hg, and Se in order to further our  
87 understanding of the internal handling of these elements in the livers of YR collected in southeast Alaska,  
88 USA. A subcellular partitioning procedure using differential centrifugation, NaOH digestion, and heat  
89 denaturation steps was used to separate liver cells into operationally-defined metal-sensitive fractions  
90 (mitochondria, microsomes and lysosomes, and heat-denatured proteins) and detoxified-metal fractions  
91 (heat-stable proteins and metal-rich granules). Following separation, trace elements were measured in  
92 each fraction to determine the degree to which YR are able to detoxify non-essential elements effectively,  
93 and to identify non-essential elements of concern for risk assessment.

## 94 **2. Material and Methods**

### 95 *2.1. Sampling site and fish collection*

96 In July 2014, adult YR (n=8; Table 1) were collected from Ernest Sound (55°51'59"N, 132°12'46"W)  
97 located near Prince of Wales Island, Alaska. The sampling location was selected based on metal  
98 concentrations presented in Barst et al. (2015). Fish were caught using rod and reel and were euthanized  
99 immediately after capture. Fish lengths (cm) and weights (g) were recorded. Livers were removed,  
100 divided for subcellular and bulk tissue analyses, and immediately frozen and maintained at -25 °C in the

101 field (2 weeks). After returning from the field, frozen samples were kept at -80 °C (at INRS-ETE) until  
102 processing.

### 103 2.2. *Subcellular partitioning procedure*

104 Yelloweye rockfish liver samples were separated into subcellular fractions (Figure S1, Supplementary  
105 Information): nuclei and debris; granule-like; mitochondria; microsomes and lysosomes; heat-denatured  
106 proteins (HDP), which include cytosolic enzymes; and heat-stable proteins and peptides (HSP), such as  
107 metallothionein (MT) and glutathione (GSH). The subcellular partitioning procedure was adapted from  
108 previous protocols described by Wallace et al. (2003) and Giguère et al. (2006). The effectiveness of the  
109 procedure at isolating subcellular fractions has been assessed previously by using enzymes as molecular  
110 markers for specific fractions or organelles (Rosabal et al., 2015). We stress that these fractions are  
111 operationally-defined in nature. Furthermore, the designation “microsomes” refers to structures which  
112 form a pellet at a given centrifugation speed, rather than structures found within cells. The details of the  
113 partitioning procedure were based on previously published methods (Rosabal et al. 2012; Rosabal et al.  
114 2014) and can be found in the Supplementary Information.

### 115 2.3. *Trace element measurements and quality control*

116 The preparation of tissue homogenates and subcellular fractions is provided in the Supplementary  
117 Information. Total Cd, Pb, As, and Se concentrations in all subcellular fractions were measured using an  
118 inductively coupled plasma-mass spectrometer (ICP-MS; Thermo Elemental X Series, Winsford,  
119 England, United Kingdom). Samples of similar weight of a certified reference material (TORT-2, lobster  
120 hepatopancreas, National Research Council of Canada, NRCC, Halifax, Nova Scotia, Canada) were  
121 subjected to the same digestion procedure and analyzed concurrently with YR fractions. The recovery of  
122 elements from TORT-2 (n = 2) was  $91 \pm 0.11\%$  for Cd,  $79 \pm 4.6\%$  for Pb,  $106 \pm 0.1\%$  for As, and  $90 \pm$   
123  $5.6\%$  for Se. The relative percent difference (RPD) between duplicate samples for Cd, Pb, As, and Se  
124 were 0.1%, 8.3%, 0.07%, and 8.8%, respectively.

125 Total Hg measurements were carried out using a direct mercury analyzer (DMA-80, Milestone Inc.,  
126 Monroe, CT), which uses thermal decomposition, amalgamation, and atomic absorption  
127 spectrophotometry according to the U.S. Environmental Protection Agency (US EPA) Method 7473 (US  
128 EPA, 2007). Quality assurance consisted of analysis of certified reference materials MESS-3 (marine  
129 sediments; n=7) and DOLT-4 (dogfish liver; n=6), National Research Council of Canada, NRCC, Halifax,  
130 Nova Scotia, Canada). Mean percent recovery of total Hg from MESS-3 was  $97 \pm 1.4$  % and the relative  
131 standard deviation (RSD) was 1.5 %. Mean percent recovery of total Hg from DOLT-4 was  $98 \pm 3.8$  %  
132 and the RSD was 3.9 %. Mass balances for Cd, Pb, As, Hg, and Se are reported in the Supplementary  
133 Information.

#### 134 *2.4. Total mercury measurements in bulk muscle tissue*

135 Freeze-dried samples of bulk muscle tissues were analyzed for total Hg using a direct Hg analyzer (DMA-  
136 80, Milestone Inc., Monroe, CT). Quality assurance consisted of the analysis of certified reference  
137 materials (DORM-4: fish protein, National Research Council Canada, Ottawa, Canada) and duplicate  
138 samples. The recovery of total Hg from DORM-4 was 96 % (n=2) and the RPD between duplicate  
139 samples was 0.80 % (n=2). In order to compare Hg concentrations to published values, wet weight  
140 concentrations in YR muscle samples were estimated by assuming a moisture content of 80%, which is  
141 consistent with a previous study with YR (Barst et al. 2015).

#### 142 *2.5. Data analyses*

143 The contribution of each subcellular fraction relative to the total element burden was estimated as a ratio  
144 defined by the element burden in a given fraction divided by the sum of element burdens in all fractions,  
145 multiplied by 100 to give results as percentages (%). Element concentrations in all subcellular fractions  
146 are expressed as total element burden (nmol) divided by the liver dry weight (g, dw). Liver dry weights  
147 were determined by weighing subsamples of liver tissue before and after freeze-drying. All numerical  
148 data are represented by means  $\pm$  standard deviations (SD), unless otherwise noted. Relationships among  
149 variables (trace element concentrations and relative contributions) were initially examined in bivariate

150 scatterplots and tested by simple correlation (Pearson  $r$ ) after checking the assumption of normality  
151 (Shapiro–Wilk test) and testing for outliers (Grubb’s test). Percentage data (relative contribution of each  
152 subcellular fraction to the total metal burden) were arcsine transformed. If non-normality persisted, a non-  
153 parametric correlation was reported (Spearman  $r$ ). When bivariate plots indicated a possible linear  
154 relationship, simple regression models were tested using the ordinary least-squares equation when the  
155 necessary assumptions (normality and homoscedasticity of residuals) were satisfied. The Shapiro–Wilk  
156 test was used to verify the normality of distributions of the regression residuals. The Breusch-Pagan test  
157 was used to test the homoscedasticity of the regression residuals. For the trace elements (Pb, Hg, and Se)  
158 that showed significant increases in more than one subcellular fraction within either the sensitive or  
159 detoxified compartments, the slopes of the linear regressions were compared using analysis of covariance  
160 (ANCOVA) in order to compare the responses along the bioaccumulation gradient. Note that a parametric  
161 assessment of covariance was preferred given that the residuals of the linear regressions were normally  
162 distributed. To explore similarities (or differences) in how hepatic trace metals were partitioned between  
163 both subcellular compartments (detoxified and metal-sensitive), we performed two separate PCA analyses  
164 that combined data into two-component models and explained 80 - 86% of the variation. The first PCA  
165 used trace element concentrations in combined sensitive fractions, and the second used trace element  
166 concentrations in combined detoxified fractions. An  $\alpha < 0.05$  was used as the threshold of significance for  
167 all statistical tests. All statistical analyses were performed with JMP Pro 13 Statistical Analysis Software  
168 (SAS Institute, Cary, NC, USA).

### 169 3. Results

#### 170 3.1. Trace element accumulation in yelloweye rockfish liver

171 Trace element concentrations in YR livers, as well as YR lengths and weights, are reported in Table 1.  
172 Liver trace element concentrations did not vary significantly with the length or weight of the fish. The  
173 ratio of maximum to minimum trace element concentrations ( $[M]_{\max}/[M]_{\min}$ ) in YR livers was greatest for  
174 Cd (9.0), followed by Pb (4.7), Hg (4.1), Se (3.0), and As (2.9). Total Hg concentrations ( $\mu\text{g g}^{-1}$  ww) are  
175 also reported for YR muscle in Table 1. Note that the simplest explanation for the variations in  
176  $[M]_{\max}/[M]_{\min}$  ratios is that the uptake : elimination ratio along the sampling gradient differs for the  
177 measured trace elements (Luoma and Rainbow, 2005). Muscle total Hg concentrations ranged from 0.3 to  
178  $1.4 \mu\text{g g}^{-1}$  ww. Muscle total Hg concentrations are reported in  $\mu\text{g g}^{-1}$  ww so that concentrations are easily  
179 comparable to established toxicity thresholds (Dillon et al., 2010; Sandheinrich and Wiener, 2011). Note  
180 that we did not determine the ages of the rockfish in the present study, however Barst et al. (2015)  
181 reported ages of YR, collected from the same sampling location and of comparable size, ranging from 16  
182 to 119 years.

#### 183 3.2. Trace element subcellular partitioning

184 We plotted trace element concentrations in whole liver against concentrations ( $\text{nmol g}^{-1}$  dw) in potentially  
185 sensitive (mitochondria, microsomes and lysosomes, and heat-denatured protein fractions) and detoxified  
186 subcellular fractions (heat-stable proteins and granule-like fractions), in order to explore potential changes  
187 in partitioning with increasing element concentrations in whole liver (Figures 1 and 3). We also  
188 investigated possible relationships between the percentage of each trace element found in the various  
189 subcellular fractions relative to the total trace element concentrations ( $\text{nmol g}^{-1}$  dw) (Figures S2 and S3,  
190 Supplementary Information). If the percentages of trace elements in each fraction did not change  
191 significantly along the bioaccumulation gradient, then data for all fish were combined to produce mean  
192 percent contributions for the various subcellular fractions per element (Figures 2 and 4). As the  
193 toxicological significance of trace element accumulation in the nuclei and debris fraction is ambiguous,

194 this fraction has been generally ignored in the ecotoxicological literature (Campbell and Hare, 2009).  
195 However, as this fraction tends to accumulate unbroken cells, it may indicate the efficacy of the  
196 homogenization step; a low and constant proportion of the trace-element in question found in the nuclei  
197 and debris suggests an efficient and precise homogenization. In the following sections, we present the  
198 results of the subcellular partitioning procedure for each of the studied trace elements.

### 199 3.3. *Cd (cadmium)*

200 There were no significant relationships between total hepatic Cd and concentrations in any of the  
201 subcellular fractions in the potentially-sensitive compartment (Figure 1A). However, the concentration of  
202 Cd in the HSP fraction increased significantly as the total hepatic Cd concentration increased ( $r^2 = 0.71$ ;  
203 slope = 0.75;  $P = 0.02$ ; Figure 1B). There were no significant relationships between the relative  
204 contributions of the fractions and total hepatic Cd (Figures S2A and S2B, Supplementary Information).  
205 When data from all fish were combined, the majority of Cd was associated with the detoxified  
206 compartment (65%), with the sensitive compartment contributing only 25%. The HSP fraction  
207 contributed the majority of the Cd in the detoxified compartment (59%), with only a minor contribution  
208 attributed to the granule-like fraction (6%). In the potentially-sensitive compartment, Cd was more or less  
209 equally distributed among the mitochondria (9%), the microsomes and lysosomes (8%), and the HDP  
210 (8%) fractions (Figure 2).

### 211 3.4. *Pb (lead)*

212 Note that total Pb concentrations ( $< 1 \text{ nmol g}^{-1}$  dry wt) in the livers of YR were much lower than molar  
213 concentrations of Cd, As, Hg, and Se. Nevertheless, the Pb concentrations increased in both the  
214 mitochondria ( $r^2 = 0.75$ ; slope =  $0.11 \pm 0.03$ ;  $P = 0.006$ ) and HDP ( $r^2 = 0.74$ ; slope =  $0.091 \pm 0.02$ ;  $P =$   
215  $0.006$ ) fractions with increasing Pb concentration in whole liver to a similar extent, as evidenced by the  
216 similar slopes of the two regressions ( $P = 0.82$ ) (Figure 1C). The concentration of Pb increased  
217 significantly in both the HSP ( $r^2 = 0.85$ ; slope =  $0.24 \pm 0.04$ ;  $P = 0.001$ ) and granule-like ( $r^2 = 0.87$ ; slope  
218 =  $0.31 \pm 0.06$ ;  $P = 0.0007$ ) fractions along the bioaccumulation gradient (Figure 1D). The rates of

219 increase in Pb concentration within the two detoxified fractions were not significantly different from one  
220 another ( $P = 0.98$ ). When comparing sensitive and detoxified fractions, the slope of the line representing  
221 the HDP fraction was significantly lower than the slopes for both the HSP ( $P = 0.04$ ) and granule-like  
222 fractions ( $P = 0.04$ ) (Figure 1C and 1D). The total hepatic Pb concentration was not significantly  
223 correlated with the relative contribution of Pb in any of the subcellular fractions (Figures S2C and S2D,  
224 Supplementary Information). When combining data from all fish, the mean relative contribution of the  
225 detoxified compartment was 46%, of which the majority was contributed by the granule-like fraction  
226 (26%). The potentially-sensitive compartment contributed only 35% of the total Pb. In the potentially-  
227 sensitive compartment, Pb was distributed similarly among the HDP (12%), mitochondria (11%), and the  
228 microsomes and lysosomes (11%) fractions (Figure 2).

### 229 3.5. As (*arsenic*)

230 Though the concentration of As increased in the mitochondria fraction along the bioaccumulation  
231 gradient ( $r^2 = 0.57$ ;  $P = 0.03$ ), the rate of increase began to plateau at higher total hepatic As  
232 concentrations (Figure 1E). Conversely, there were no significant relationships between total hepatic As  
233 and concentrations in either the HDP or microsomes and lysosomes fractions. The concentration of As in  
234 the HSP fraction increased significantly with increasing As in whole liver ( $r^2 = 0.84$ ; slope = 0.86;  $P =$   
235 0.01) (Figure 1F). The relative contributions of As increased significantly in the HSP ( $r^2 = 0.60$ ;  $P = 0.02$ ),  
236 while decreasing in the microsomes and lysosomes ( $r^2 = 0.51$ ;  $P = 0.046$ ) along the bioaccumulation  
237 gradient (Figures S2E and S2F, Supplementary Information). Among all fish, the relative contributions of  
238 the HSP and granule-like fractions to the total hepatic As burden varied from 50% to 75% and 2% to 9%,  
239 respectively. The relative contributions of the mitochondria, microsomes and lysosomes, and HDP  
240 fractions varied from 7% to 14%, 4% and 10%, and 5% and 10%, respectively.

### 241 3.6. Hg (*mercury*)

242 As the concentration of total Hg increased in whole liver, the concentrations of total Hg increased in the  
243 mitochondria ( $r^2 = 0.86$ ; slope = 0.25;  $P = 0.001$ ), microsomes and lysosomes ( $r^2 = 0.91$ ; slope = 0.17;  $P =$

244 0.0002), HSP ( $r^2 = 0.58$ ; slope = 0.04;  $P = 0.03$ ), and granule-like fractions ( $r^2 = 0.64$ ; slope = 0.07;  $P =$   
245 0.03) (Figures 3A and 3B). Within the sensitive compartment, the increase in total Hg in the  
246 mitochondria fraction was greater than in the microsomes and lysosomes fraction ( $P = 0.03$ ). Conversely,  
247 within the detoxified compartment, total Hg increased in the HSP and granule-like fractions to a similar  
248 extent along the bioaccumulation gradient ( $P = 0.45$ ). Total Hg increased in both the mitochondria and  
249 microsomes and lysosomes fractions to a greater extent than in the HSP or granule-like fractions  
250 ( $P < 0.05$ ). The relative contributions of the various subcellular fractions to total hepatic Hg did not vary  
251 significantly along the bioaccumulation gradient (Figures S3A and S3B, Supplementary Information). For  
252 all fish, the mean proportion of total Hg in the detoxified compartment was 21%, with about half being  
253 contributed by the granule-like fraction (12%). Discounting the nuclei and debris fraction, the majority of  
254 the total hepatic Hg burden was associated with the potentially-sensitive compartment (49%). Within this  
255 compartment, the contributions of the individual fractions decreased in the order mitochondria (18%)  $\geq$   
256 microsomes and lysosomes (16%)  $\geq$  HDP (15%) (Figure 4).

### 257 3.7. Se (selenium)

258 Concentrations of Se increased significantly in all of the subcellular fractions, except for the granule-like  
259 fraction, as Se increased in whole liver. Within the potentially-sensitive compartment, the relation with  
260 total hepatic Se was tightest for HDP ( $r^2 = 0.95$ ; slope =  $0.16 \pm 0.02$ ;  $P = 0.0002$ ) followed by  
261 mitochondria ( $r^2 = 0.90$ ; slope =  $0.16 \pm 0.02$ ;  $P = 0.0003$ ), and microsomes and lysosomes ( $r^2 = 0.67$ ;  
262 slope =  $0.08 \pm 0.02$ ;  $P = 0.01$ ) fractions, though there were no significant differences among the slopes of  
263 the regressions (Figure 3C). Within the detoxified compartment, Se increased significantly only in HSP  
264 ( $r^2 = 0.76$ ; slope = 0.27;  $P = 0.0051$ ) (Figure 3D). The increase in Se concentration in HSP was  
265 significantly greater than in the mitochondria ( $P = 0.04$ ) or microsomes and lysosomes ( $P = 0.002$ )  
266 fractions. The relative contributions of the various subcellular fractions to the total hepatic Se  
267 concentration did not vary significantly as a function of the total hepatic Se concentration (Figures S3C  
268 and S3D, Supplementary Information). When data for all fish were combined, Se was associated with

269 potentially-sensitive and detoxified compartments similarly (40% and 39%, respectively). A larger  
270 percentage of Se was found in the HSP fraction (29%), than the granule-like fraction (10%). In the  
271 potentially-sensitive compartment, the relative contributions of Se were similar, decreasing in the order  
272 of the HDP fraction (16%) followed by mitochondria (13%), and microsomes and lysosomes (12%)  
273 (Figure 4).

## 274 **4. Discussion**

### 275 *4.1. General considerations*

276 The limited sample size in the present study was dictated by difficulties in sampling the Alaskan YR and  
277 the conservative catch limits enforced by the State of Alaska. Despite the small sample size, our results  
278 increase knowledge related to the internal handling of trace elements in wild fish. Our study is unique in  
279 that YR have an unusually long life span compared to other fish for which the subcellular partitioning of  
280 trace elements has been reported. Note too that most other studies in this area have been limited to  
281 ‘traditional’ metals such as Cd, Hg and Pb, whereas in the present study we have also included two  
282 metalloids, As and Se.

283 Numerous studies have applied subcellular partitioning procedures to determine the distribution of  
284 non-essential elements in the tissues of aquatic organisms (Giguère et al., 2006; Rosabal et al., 2015;  
285 Wang et al., 2016). These procedures provide insight into how aquatic organisms cope with non-essential  
286 metals and indicate whether toxicological effects are likely to occur. However, partitioning procedures are  
287 subject to potential problems which have been described in depth elsewhere (Campbell and Hare, 2009;  
288 Hinton et al., 1997). Subcellular fractions are operationally-defined, and accordingly the interpretation of  
289 partitioning results should be carried out with circumspection.

290 In this context, the terms “MT-like” or “granule-like” should be considered carefully. A previous  
291 study on *Chaoborus* larvae showed that not all metals measured in the HSP fraction are necessarily  
292 associated with MT or MTLP (Rosabal et al., 2016; Caron et al. 2018). This caveat is likely particularly  
293 important for As and Se, which tend to form covalent bonds with oxygen or reduced sulphur and may

294 exist as oxyanions or esters in the intracellular environment, rather than as chelated cations. Furthermore,  
295 the definitions of the “metal detoxified pool” and the “metal-sensitive pool” were designed for (soft)  
296 cationic metals, not for As (a metalloid that is reduced to and is largely present as As(III) in living cells)  
297 or Se. Moreover, the lumping of fractions into potentially-sensitive (mitochondria, HDP, microsomes and  
298 lysosomes fractions) and detoxified-metal compartments (HSP and granule-like), is likely an  
299 oversimplification (Campbell and Hare, 2009; Wallace et al., 2003). For example, microsomes cannot be  
300 separated effectively from lysosomes using the procedure in the present study, and this renders the  
301 interpretation of results more difficult. If a non-essential metal is primarily associated with lysosomes,  
302 then the metal has likely been detoxified. Conversely, the endoplasmic reticulum is a potential target for  
303 metal toxicity, and therefore, metals associated with these vesicles may cause deleterious effects. Previous  
304 studies have often grouped the microsome and lysosome fraction in the metal-sensitive compartment due  
305 to the important functions carried out by the endoplasmic reticulum, Golgi apparatus, and ribosomes in  
306 the liver. However, strict inclusion of this fraction in the sensitive compartment may not be appropriate  
307 for all metals. For instance, a recent study demonstrated, with electron energy loss spectrometry (EELS),  
308 that Hg accumulates in the hepatic lysosomes of wild yellow perch (*Perca flavescens*) (Müller et al.,  
309 2015). With this in mind, Hg (and potentially other trace elements) within the microsomes and lysosomes  
310 fraction may be associated with lysosomes to a greater extent than microsomes. As previously mentioned,  
311 the accumulation of non-essential metals in the nuclei and debris fraction is also difficult to interpret.

#### 312 4.2. Cd (cadmium)

313 The high proportion of Cd associated with the HSP fraction suggests that Cd is largely detoxified by MT  
314 in YR livers, particularly for concentrations up to  $\sim 150 \text{ nmol g}^{-1} \text{ dw}$ . The lack of increasing trends  
315 between total hepatic Cd and the Cd concentration in sensitive fractions suggests this nonessential metal  
316 is largely kept under control in the livers of YR. Furthermore, the much lower Cd concentrations in  
317 sensitive fractions relative to that in HSP suggest effective detoxification. An interaction between HSP  
318 and Cd is consistent with the classification of Cd as a class B metal, which exhibits preferences for

319 reduced sulphur within cells (Mason and Jenkins, 1995). Elevated proportions of Cd were also found in  
320 the HSP fractions isolated from livers of wild American and European eels (Rosabal et al., 2015), as well  
321 as wild yellow perch (Giguère et al., 2006). Laboratory studies also indicate the importance of the HSP  
322 fraction in detoxifying Cd in fish. For example, Olsson and Hogstrand (1987) showed that Cd was  
323 associated with MT in rainbow trout livers following a 1-week aqueous exposure to  $^{109}\text{Cd}$  (3-60 ng/L). Ng  
324 and Wood (2008) noted elevated proportions of Cd in fractions containing MT-like proteins isolated from  
325 the gut tissue of rainbow trout fed contaminated oligochaetes. Similarly, Zhang and Wang (2006) reported  
326 that the HSP fraction was the major storage compartment for Cd in the viscera of juvenile marine grunt  
327 fed brine shrimp previously exposed to aqueous  $^{109}\text{Cd}$ . Together, these studies indicate that maintaining  
328 Cd in the HSP fraction is an important metal-handling strategy for fish species.

329         Though YR in the present study maintained the majority of Cd in the detoxified compartment,  
330 detoxification was not complete given the presence of some Cd at sensitive sites. In the sensitive  
331 compartment, isolated from whole zebrafish (*Danio rerio*) fed contaminated chironomids (153 – 288  $\mu\text{g}$   
332  $\text{g}^{-1}$  dw), Cd was found mainly in the “organelles” fraction (including mitochondria and microsomes and  
333 lysosomes) followed by the HDP fraction (Bécharde et al., 2008), which was similar to results for YR in  
334 the present study. Interestingly, the concentration of Cd did not increase significantly in sensitive  
335 fractions of YR livers along the bioaccumulation gradient, contrary to what was shown for wild yellow  
336 perch (Giguère et al., 2006) and wild American and European yellow eels (Rosabal et al., 2015). This  
337 may be due to the more modest Cd gradient in the present study ( $[\text{Cd}]_{\text{max}} : [\text{Cd}]_{\text{min}} = 9.0$ ) relative to the  
338 gradients reported by (Giguère et al., 2006) ( $[\text{Cd}]_{\text{max}} : [\text{Cd}]_{\text{min}} = 14$ ) and Rosabal et al. (2015) (American  
339 eels  $[\text{Cd}]_{\text{max}} : [\text{Cd}]_{\text{min}} = 103$  and European eels  $[\text{Cd}]_{\text{max}} : [\text{Cd}]_{\text{min}} = 200$ ). Moreover, the lack of significant  
340 relationships between Cd in sensitive fractions and Cd in whole liver may be a result of the variability in  
341 Cd concentrations. For example, two YR in the present study had total hepatic Cd concentrations of  $\sim 90$   
342  $\text{nmol g}^{-1}$  dw, yet the Cd concentrations within the sensitive fractions of these individuals were quite  
343 different. This may indicate that individual YR can have less-(or more-)effective metal detoxification

344 systems than other fish in their population, a phenomenon that has been demonstrated previously for  
345 yellow perch sampled from metal-impacted lakes (Couture and Pyle, 2008).

346 Nevertheless, Cd was associated with sensitive subcellular sites within YR livers, and this could lead  
347 to negative health effects. Accumulation of Cd in mitochondria may affect functioning of this organelle,  
348 including an inhibition of citrate synthase, which has been documented for yellow perch livers collected  
349 from a highly contaminated lake in Canada (Couture and Rajotte, 2003). In a separate study with yellow  
350 perch, Ponton et al. (2016) noted that individuals suffering from oxidative stress had higher percentages  
351 of Cd, Cu, and Zn in potentially sensitive subcellular fractions, which may highlight the importance of  
352 maintaining these metals in detoxified fractions.

#### 353 4.3. *Pb (lead)*

354 Within the detoxified compartment, Pb associated with the granule-like fraction to a slightly greater  
355 extent than with the HSP fraction. The preference for the granule-like fraction over the HSP fraction  
356 containing MT may be due to the fact that Pb, as a borderline metal, associates less readily with thiol  
357 groups than other “softer” metals, such as Cd and Ag (Mason and Jenkins, 1995). Nonetheless, the  
358 concentration of Pb increased in both the metal-rich granule and HSP fractions along the bioaccumulation  
359 gradient. Given the low Pb concentrations, the association of Pb with the HSP fraction may be an indirect  
360 response to Cd, which has consistently been shown to induce synthesis of MT. Following the induction of  
361 MT by Cd, MT is available to bind both Cd and other metals such as Pb. Alternatively, the association of  
362 Pb with the HSP fraction could be an indication of a detoxification response, although this seems unlikely  
363 given the very low concentrations of hepatic Pb. Much of the previous research involving the subcellular  
364 distribution of Pb has focused on invertebrates, with only a few studies reporting Pb partitioning in the  
365 tissues of fish. The distribution of Pb within the detoxified compartment of YR livers agrees well with the  
366 results of studies on invertebrates (Marigómez et al., 2002; Mason and Jenkins, 1995; Sánchez-Marín and  
367 Beiras, 2017; Wang et al., 2016) and other fish (Goto and Wallace, 2010; Rosabal et al., 2015), which  
368 show collectively that metal-rich granules are the primary binding pool within the detoxified

369 compartment. Contrary to these results, Dang et al. (2012) reported that the majority of Pb was found in  
370 the HSP fraction of intestinal cells of the marine grunt (*Terapon jarbua*).

371         Though Pb increased in detoxified fractions of YR liver, detoxification was incomplete given that  
372 Pb was also present in metal-sensitive fractions and increased along the bioaccumulation gradient. In  
373 terms of percentage, Pb was distributed similarly among the metal-sensitive fractions, only showing a  
374 slight preference for HDP. Consistent with our results, Rosabal et al. (2015) also reported increases of Pb  
375 in metal-sensitive fractions of yellow eels with increasing concentrations in whole liver. However, Pb did  
376 not increase in the metal-sensitive fractions of mummichogs (*Fundulus heteroclitus*) collected from  
377 metal-polluted salt marshes, though the two-fold bioaccumulation gradient of Pb was somewhat limited  
378 (Goto and Wallace, 2010). In YR, the association of Pb with metal-sensitive fractions may result in  
379 deleterious effects, in view of the ability Pb to replace other essential metals, such as Ca, within biological  
380 systems (Rogers et al., 2003). Furthermore, associations between Pb and mitochondria, enzymes, and  
381 microsomes would be expected to disrupt cellular processes. However, given the very low total hepatic  
382 Pb concentrations in the livers of YR in the present study, it is unlikely that this metal was of great  
383 toxicological concern to these fish.

#### 384 4.4. As (arsenic)

385 Within the detoxified compartment of YR livers, As was primarily associated with the HSP fraction,  
386 containing MT. A steady increase of As in HSP suggests that this fraction may be involved in As  
387 detoxification within YR livers. To our knowledge, there are no previous studies on subcellular  
388 partitioning of As in the livers of rockfish species, though our results compare well with As partitioning  
389 in American and European yellow eels (*Anguilla rostrata* and *Anguilla anguilla*); like YR, both eel  
390 species maintained As in the HSP fraction, and to a lesser extent, in the metal-rich granule fraction  
391 (Rosabal et al., 2015). Similarly, in seabass (*Lateolabrax japonicas*) and seabream (*Pagrosomus major*)  
392 muscle, As was largely associated with the HSP fraction (He et al., 2010). These results are also  
393 consistent with a study on marine grunt (*Terapon jarbua*) exposed to dietary or aqueous As(III) and  
394 As(V) at environmentally-relevant concentrations for 10 d. In grunt muscle tissues, As accumulated

395 mainly in the HSP fraction, whereas less was associated with metal-rich granules (Zhang et al., 2012).  
396 Although As was associated with the HSP fraction there may not be a significant interaction between As  
397 and MT given the lower affinity this metalloid has for thiol functional groups.

398 In the metal-sensitive compartment, As was predominately associated with the mitochondria  
399 fraction, which is consistent with results from a previous study on American and European yellow eels  
400 (Rosabal et al., 2015). Our results differ slightly from the study by Rosabal et al. (2015) in that in eel  
401 livers, the microsomes and lysosomes fraction was a more important binding pool for As than the HDP  
402 fraction. In the present study, the microsomes and lysosomes and HDP fractions contributed roughly the  
403 same amount of As to the total As burden in YR livers. A study by Dang et al. (2012) noted binding of As  
404 in the “organelles” fraction (mitochondria and microsomes and lysosomes) of a polychaete worm (*Nereis*  
405 *diversicolor*) sampled from a contaminated estuary. In the same study, the contaminated worms were fed  
406 to fish (*Terapon jarbua*), which subsequently accumulated As in the hepatic “organelle” fraction. The  
407 combined results from these studies suggest that across species, organelles may be an important target for  
408 As at the subcellular level.

409 Along the As contamination gradient, we noted an increase in the relative contribution of As in  
410 the HSP fraction within YR livers, coupled with a decrease in the relative contribution of the microsomes  
411 and lysosomes fraction, suggesting an activation of detoxification mechanisms. Interestingly, the relative  
412 contributions of As in the mitochondria and HDP fractions increased with increasing As in whole liver  
413 until approximately 80 nmol g<sup>-1</sup> dw, at which point the relative contribution of As in the fractions  
414 decreased along the bioaccumulation gradient (this was also the case for the microsomes and lysosomes  
415 fraction, though the increase and subsequent decrease in relative contribution of As was less pronounced,  
416 thus allowing for a significant linear trend to be fitted) (Figures S2E and S2F, Supplementary  
417 Information). The distributions of the relative contributions of As along the gradient suggest that  
418 detoxification of As became more effective above ~80 nmol g<sup>-1</sup> dw in the livers of YR. This is similar to  
419 what Rosabal et al. (2012) noted for Cd in *Chaoborus*, i.e. a certain threshold of Cd was necessary to  
420 “turn on” detoxification mechanisms fully.

421 In the present study we did not determine As speciation in the livers of YR, though speciation of As  
422 is a major determinant of its toxicity (Watanabe and Hirano, 2013). Future work should focus on  
423 determining the oxidation state and methylation levels of As in wild fish at the subcellular level.  
424 Additionally, we highlight the need for subsequent work to identify the As-bearing molecules in the  
425 cytosol, specifically in the HSP fraction, which appears to be involved in As detoxification.

#### 426 4.5. Hg (mercury)

427 Relatively few studies have focused on the subcellular distribution of Hg in fish liver (Araújo et al., 2015;  
428 Barst et al., 2016; Peng et al., 2016). Araújo et al. (2015) measured total Hg in subcellular fractions of the  
429 livers of wild mullets (*Liza aurata*), and found low contributions of Hg in the HSP and granule fractions,  
430 which the authors attributed to Hg concentrations below a physiological threshold to activate  
431 detoxification mechanisms. In Arctic char (*Salvelinus alpinus*) liver cells, Barst et al. (2016) reported that  
432 the HSP fraction was the primary binding pool for total Hg within the detoxified compartment, and less  
433 than 1% of the total Hg burden was found in the metal-rich granule fraction. Similarly, metal-rich  
434 granules played a less important role in the detoxification of methylmercury (MeHg) than MT-like  
435 proteins in the livers of rabbitfish (*Siganus canaliculatus*) (Peng et al., 2016). Results of the present study  
436 contrast with those from these earlier studies suggesting that the HSP fraction is more important than  
437 metal-rich granules in the detoxification of Hg in fish liver cells. A plausible explanation for the greater  
438 importance of the metal-rich granule fraction than the HSP fraction in sequestering Hg in the livers of YR  
439 may be linked to Hg speciation. In a previous study on YR, inorganic mercury (InHg) comprised a major  
440 proportion of the total Hg (mean =  $58 \pm 14.2\%$ ) in liver tissue (Barst et al. 2015). In addition to elevated  
441 proportions of InHg in the livers of YR, Barst et al. (2015) demonstrated a co-localization of Hg and Se  
442 within hepatic MA. The authors hypothesized that InHg was bound to Se, forming HgSe granules that are  
443 thought to be the end-product of MeHg detoxification (Wang et al. 2012). The accumulation of total Hg  
444 in the metal-rich granule fractions of YR may therefore represent a long-term accumulation of HgSe  
445 granules in YR livers. In the same manner, the low accumulation of total Hg in the metal-rich granule

446 fraction of Arctic char (Barst et al. 2016) and rabbitfish (Peng et al. 2016) may be a result of the low  
447 proportions of InHg in their tissues.

448 Despite the accumulation of total Hg in HSP and metal-rich granules, detoxification in the livers  
449 of YR was incomplete, as evidenced by the total Hg present in metal-sensitive fractions, and the increase  
450 in total Hg concentration in these fractions along the bioaccumulation gradient. The total Hg found in the  
451 sensitive fractions of YR liver could negatively impact the health of these fish. For example, Cambier et  
452 al. (2009) noted an inhibition of both state 3 mitochondrial respiration and cytochrome c oxidase activity  
453 in the muscle fibers of zebrafish exposed to an environmentally-relevant dose of dietary MeHg for 49  
454 days. As the mitochondria fraction contributed the greatest percentage of total Hg among sensitive and  
455 detoxified fractions, YR could be suffering from inhibited respiration and thus, altered energy  
456 metabolism. The accumulation of total Hg in the HDP fraction could have consequences for the redox  
457 defense system; Se-dependent enzymes, such as glutathione peroxidase (GSH-Px) and thioredoxin  
458 reductase (TrxR), are likely molecular targets for intracellular Hg due to the high binding affinity Hg has  
459 for Se. In support of this, a laboratory feeding study documented decreased activity of GSH-Px in the  
460 brains of juvenile Atlantic salmon (*Salmo salar*) exposed to MeHg (Berntssen et al., 2003). Decreased  
461 activities of GSH-Px and TrxR have also been documented in the tissues of zebra-seabream (*Diplodus*  
462 *cervinus*) following aqueous exposure to either MeHg or InHg (Branco et al., 2012). As previously stated,  
463 the presence of total Hg in the microsomes and lysosomes fraction is more difficult to interpret, and our  
464 inability to classify this fraction in either the detoxified or sensitive compartment provides an opportunity  
465 for future studies.

#### 466 4.6. Se (selenium)

467 In contrast to the other elements measured in YR livers, Se is essential to normal cellular function.  
468 Despite this essential nature, above a threshold concentration Se may become toxic. Given that the range  
469 of concentrations in YR livers in the present study is within the range of those reported to afford  
470 protection from oxidative stress in the livers of yellow perch (Ponton et al. 2016), we discuss Se in YR in

471 the context of its potential ameliorative effects. Selenium is known to have an interaction with non-  
472 essential metals such as Cd, As, and Hg (Sasakura and Suzuki, 1998), and this may confer a protective  
473 action against toxicity (Banni et al., 2011; Wang et al., 2013). We chose to include Se in the present study  
474 because it is known to have a strong binding affinity for Hg and it is well understood that Hg handling in  
475 the subcellular environment involves interactions with selenols (Wang et al. 2012). Additionally, previous  
476 work has demonstrated colocalization of Hg and Se in immune cells, suggesting an interaction between  
477 the two elements within YR livers (Barst et al. 2015). The protective effects of Se on Hg toxicity have  
478 been the subject of a significant amount of research, some of which hypothesizes that Se is protective if  
479 molar ratios (Se:Hg) meet or exceed unity (Ralston et al. 2007). However this protective effect has been  
480 largely studied from a human health perspective and has not been well explored in terms of the health of  
481 wild fish nor focused on the liver with its major role in detoxification. Within the detoxified compartment,  
482 Se was found primarily in the HSP fraction, and to a lesser extent in the granule-like fraction. Within the  
483 HSP fraction, Se is most likely present as seleno-cysteine (Gladyshev, 2012) in thermostable metal-  
484 binding proteins, such as MT. As previously mentioned, Se within the granule-like fraction may be  
485 present as HgSe. In contrast to Arctic char livers, where less than 2% of the total Se burden was  
486 associated with the metal-rich granule fraction, the granule-like fraction isolated from YR livers  
487 comprised an average of 10% of the total Se burden. Interestingly, the percentages of Se in the metal-rich  
488 granules isolated from both species of fish were similar to the proportions of total Hg in the respective  
489 fractions. This is likely another indication of the interaction between the two elements. The proportion of  
490 Se within the metal-sensitive compartment roughly equaled that in the detoxified compartment. In metal-  
491 sensitive fractions, Se demonstrated a slight preference for the HDP fraction, which is not surprising  
492 given selenium's biochemical role in enzymes, such as GSH-Px and TrxR.

#### 493 *4.7. Overall subcellular element partitioning*

494 We noted clear differences in the partitioning of the trace elements As, Cd, Pb, Hg, and Se in the livers of  
495 YR. As Class B metals, Hg and Cd have high affinities for thiols in biological systems, and therefore

496 these non-essential metals would be expected to display similar subcellular partitioning. Both elements  
497 appear in the same quadrant in both PCA loading plots (bottom right) for the metal-sensitive and the  
498 detoxified metal compartments (Figure 5A-B), which presumably reflects the common affinity that these  
499 metals have for SH- functional groups. Conversely, both As and Pb are borderline elements, and their  
500 proximity within the PCA plots is likely a result of the lower affinities that these metals share for thiols  
501 relative to “soft” metals such as Cd and Hg (Figures 5A and 5B). Interestingly, for the detoxified-metal  
502 compartment, As and Hg vectors are directed in opposite directions (both vectors are projected at 180°) in  
503 the PCA figure, indicating a potential negative relationship between the partitioning of As and Hg in this  
504 compartment. A similar trend between As and Hg is also observed in the metal-sensitive compartment,  
505 where the metal vectors are projected perpendicularly. We speculate that both elements could be targeting  
506 similar biomolecules. Additionally, the location of the Se vector between the As and Hg vectors may  
507 indicate potential interactions between Se and these two elements.

508         Yelloweye rockfish were able to maintain some of these non-essential metals in detoxified  
509 fractions, suggesting an ability to cope with these metals to some extent. Both Cd and As were mainly  
510 found in the HSP fraction within the detoxified compartment, indicating a potential interaction with MT.  
511 In contrast, both Pb and Hg showed a greater preference for the granule-like fraction than the HSP  
512 fraction. For Pb, this trend is consistent with results from subcellular partitioning in eels (Rosabal et al.  
513 2015), but for Hg this apparent role of the granule-like fraction within the livers of YR is not consistent  
514 with results of a previous study with Arctic char, in which the HSP played a much more important role in  
515 detoxification. We hypothesize that the observed divergence may be related to differences in Hg  
516 speciation and/or age between the two species of fish. Future studies should focus on exploring the  
517 subcellular partitioning of InHg and MeHg to determine possible differences.

518         Although, non-essential metals were associated with detoxified fractions within the livers of YR,  
519 detoxification was incomplete as each of the non-essential metals was also associated with potentially-  
520 sensitive sites. However, only Pb and Hg increased significantly within potentially-sensitive fractions  
521 along the bioaccumulation gradient, suggesting that within the tissue concentration ranges reported here

522 YR are less efficient at detoxifying these metals than Cd and As. Within the potentially-sensitive  
523 compartment, the fraction containing mitochondria was consistently important for binding of Cd, Pb, As,  
524 and Hg. The accumulation of these non-essential metals in this fraction may lead to negative effects,  
525 given the key role that these organelles play in cellular metabolism. Subcellular partitioning procedures,  
526 such as the one employed in the present work, provide useful information on how trace elements are  
527 distributed within cells, thus moving beyond more simple measures in bulk tissues. This type of  
528 information can be useful when trying to understand risk associated with multiple non-essential elements.  
529 Collectively, our results suggest that Hg may be of greatest concern to the health of YR relative to the  
530 other non-essential metals studied; the majority of Hg was associated with the sensitive compartment,  
531 whereas the other non-essential metals were predominately associated with the detoxified compartment.  
532 Recent analyses of the available data for Hg toxicity in fish indicate that toxic effects are likely to occur at  
533 concentrations exceeding  $0.3 \mu\text{g g}^{-1}$  ww (Dillon et al. 2010; Sandheinrich and Wiener, 2011) (equivalent  
534 concentration in edible muscle  $0.5 \mu\text{g g}^{-1}$  ww). In the present study, 5 of the 8 YR exceed this toxicity  
535 threshold (Table 1), suggesting that they are indeed at risk for the toxic effects of Hg. Our work  
536 demonstrates that non-essential metals accumulate in potentially-sensitive sites (mitochondria,  
537 microsomes, and enzymes), which may have implications for the health of YR.

### 538 **Acknowledgements**

539 We thank A. Caron, J. Perreault and R. Rodrigue for laboratory assistance. We also thank members of the  
540 Griggers family at TreeTops Lodge for lodging and assistance in the field. P.G.C. Campbell and N. Basu  
541 are supported by the Canada Research Chairs programme. This project was supported by the NSERC  
542 Discovery Grant programme (Drevnick).

### 543 **Supplementary Information**

544 Consists of methods related to subcellular partitioning and trace element analyses, in addition to three  
545 figures (S2, S3, and S4).

**Table 1. Ranges in lengths and weights, mean hepatic trace element concentrations (nmol g<sup>-1</sup> dw), and total mercury concentrations in muscle (µg g<sup>-1</sup> ww) of yelloweye rockfish (*Sebastes ruberrimus*) collected in southeast Alaska.**

546

Fish	Length (cm)	Weight (g)	Liver concentrations (nmol g <sup>-1</sup> dw)					Muscle concentrations (µg g <sup>-1</sup> ww)
			As	Cd	Pb	Hg	Se	Hg
1	64.5	4320	59.5	147.3	1.2	14.2	249.6	0.63
2	57.5	2530	74.8	116.7	1.0	13.7	241.3	0.29
3	54.8	3020	51.6	89.6	0.3	25.0	148.3	0.38
4	51.5	2140	148.7	53.5	0.7	9.8	136.8	0.30
5	57.0	2750	110.1	131.9	1.2	18.3	226.9	0.59
6	57.0	2640	77.4	105.1	0.6	10.7	279.5	0.51
7	62.3	3740	85.8	478.8	1.5	39.6	417.0	1.44
8	79.0	8800	64.5	90.4	0.4	22.7	202.7	0.88
<i>n</i>	8	8	23 <sup>a</sup>	22 <sup>a</sup>	18 <sup>a</sup>	21 <sup>a</sup>	23 <sup>a</sup>	8
<i>min</i>	51.5	2140	51.6	53.5	0.3	9.8	136.8	0.3
<i>max</i>	79.0	8800	148.7	478.8	1.5	39.6	417.0	1.4
<i>max:min</i>	1.5	4.1	2.9	9.0	4.7	4.1	3.0	4.9

<sup>a</sup> n represents the total number of liver samples for which the mass balance recovery was between 61 and 150 %.

547

548

549

550

551 **5. References**

- 552 Araújo, O., Pereira, P., Cesário, R., Pacheco, M., Raimundo, J., 2015. The sub-cellular fate of  
553 mercury in the liver of wild mullets (*Liza aurata*)—Contribution to the understanding of metal-  
554 induced cellular toxicity. *Mar. Pollut. Bull.* 95, 412-418.
- 555 Banni, M., Chouchene, L., Said, K., Kerkeni, A., Messaoudi, I., 2011. Mechanisms underlying  
556 the protective effect of zinc and selenium against cadmium-induced oxidative stress in zebrafish  
557 *Danio rerio*. *BioMetals* 24, 981-992.
- 558 Barst, B.D., Bridges, K., Korbas, M., Roberts, A.P., Van Kirk, K., McNeel, K., Drevnick, P.E.,  
559 2015. The role of melano-macrophage aggregates in the storage of mercury and other metals: An  
560 example from yelloweye rockfish (*Sebastes ruberrimus*). *Environ. Toxicol. Chem.* 34, 1918-  
561 1925.
- 562 Barst, B.D., Rosabal, M., Campbell, P.G., Muir, D.G., Wang, X., Köck, G., Drevnick, P.E., 2016.  
563 Subcellular distribution of trace elements and liver histology of landlocked Arctic char  
564 (*Salvelinus alpinus*) sampled along a mercury contamination gradient. *Environ. Pollut.* 212, 574-  
565 583.
- 566 Béchar, K., Gillis, P., Wood, C., 2008. Trophic transfer of Cd from larval chironomids  
567 (*Chironomus riparius*) exposed via sediment or waterborne routes, to zebrafish (*Danio rerio*):  
568 Tissue-specific and subcellular comparisons. *Aquat. Toxicol.* 90, 310-321.
- 569 Berntssen, M., Aatland, A., Handy, R., 2003. Chronic dietary mercury exposure causes oxidative  
570 stress, brain lesions, and altered behaviour in Atlantic salmon (*Salmo salar*) parr. *Aquat. Toxicol.*  
571 65, 55-72.
- 572 Branco, V., Canário, J., Lu, J., Holmgren, A., Carvalho, C., 2012. Mercury and selenium  
573 interaction in vivo: effects on thioredoxin reductase and glutathione peroxidase. *Free Radical*  
574 *Biol. Med.* 52, 781-793.
- 575 Cambier, S., Benard, G., Mesmer-Dudons, N., Gonzalez, P., Rossignol, R., Brethes, D.,  
576 Bourdineaud, J.-P., 2009. At environmental doses, dietary methylmercury inhibits mitochondrial

- 577 energy metabolism in skeletal muscles of the zebra fish (*Danio rerio*). The international journal  
578 of biochemistry & cell biology 41, 791-799.
- 579 Campbell, P.G., Hare, L., 2009. Metal detoxification in freshwater animals. Roles of  
580 metallothioneins. Metallothioneins and related chelators, 239-277.
- 581 Caron, A., Rosabal, M., Drevet, O., Couture, P., Campbell, P.G., 2018. Binding of trace elements  
582 (Ag, Cd, Co, Cu, Ni, and Tl) to cytosolic biomolecules in livers of juvenile yellow perch (*Perca*  
583 *flavescens*) collected from lakes representing metal contamination gradients. Environmental  
584 Toxicology and Chemistry 37, 576-586.
- 585 COSEWIC, 2009. COSEWIC assessment and status report on the Yelloweye rockfish, *Sebastes*  
586 *ruberrimus*: Pacific Ocean inside waters population, Pacific Ocean outside waters population in  
587 Canada.
- 588 Couture, P., Pyle, G., 2008. Live fast and die young: metal effects on condition and physiology of  
589 wild yellow perch from along two metal contamination gradients. Hum. Ecol. Risk Assess. 14,  
590 73-96.
- 591 Couture, P., Rajotte, J.W., 2003. Morphometric and metabolic indicators of metal stress in wild  
592 yellow perch (*Perca flavescens*) from Sudbury, Ontario: a review. J. Environ. Monit. 5, 216-221.
- 593 Dang, F., Rainbow, P.S., Wang, W.-X., 2012. Dietary toxicity of field-contaminated invertebrates  
594 to marine fish: Effects of metal doses and subcellular metal distribution. Aquat. Toxicol. 120, 1-  
595 10.
- 596 Dillon, T., Beckvar, N., Kern, J., 2010. Residue-based mercury dose-response in fish: An  
597 analysis using lethality-equivalent test endpoints. Environ. Toxicol. Chem. 29, 2559-2565.
- 598 Giguère, A., Campbell, P.G., Hare, L., Couture, P., 2006. Sub-cellular partitioning of cadmium,  
599 copper, nickel and zinc in indigenous yellow perch (*Perca flavescens*) sampled along a  
600 polymetallic gradient. Aquat. Toxicol. 77, 178-189.
- 601 Gladyshev, V.N., 2012. Selenoproteins and selenoproteomes, Selenium. Springer, pp. 109-123.

- 602 Goto, D., Wallace, W.G., 2010. Metal intracellular partitioning as a detoxification mechanism for  
603 mummichogs (*Fundulus heteroclitus*) living in metal-polluted salt marshes. Mar. Environ. Res.  
604 69, 163-171.
- 605 He, M., Ke, C.-H., Wang, W.-X., 2010. Effects of cooking and subcellular distribution on the  
606 bioaccessibility of trace elements in two marine fish species. J. Agric. Food Chem. 58, 3517-  
607 3523.
- 608 Hinton, R.H., Mullock, B.M., Gilhuus-Moe, C., 1997. Isolation of subcellular fractions.  
609 Subcellular Fractionation: A Practical Approach, 31-69.
- 610 Luoma, S.N., Rainbow, P.S., 2005. Why is metal bioaccumulation so variable? Biodynamics as a  
611 unifying concept. Environmental Science & Technology 39, 1921-193
- 612 Love, M.S., Yoklavich, M., Thorsteinson, L.K., 2002. The rockfishes of the northeast Pacific.  
613 Univ of California Press.
- 614 Marigómez, I., Soto, M., Cajaraville, M.P., Angulo, E., Giamberini, L., 2002. Cellular and  
615 subcellular distribution of metals in molluscs. Microsc. Res. Tech. 56, 358-392.
- 616 Mason, A., Jenkins, K., 1995. Metal detoxification in aquatic organisms. Metal speciation and  
617 bioavailability in aquatic systems 3, 479-608.
- 618 Müller, A.-K., Brinkmann, M., Baumann, L., Stoffel, M.H., Segner, H., Kidd, K.A., Hollert, H.,  
619 2015. Morphological alterations in the liver of yellow perch (*Perca flavescens*) from a biological  
620 mercury hotspot. Environmental Science and Pollution Research, 1-13.
- 621 Khan, M.A.K., Wang, F., 2009. Mercury selenium compounds and their toxicological  
622 significance: Toward a molecular understanding of the mercury selenium antagonism. Environ.  
623 Toxicol. Chem. 28, 1567-1577.
- 624 Korbas, M., O'Donoghue, J.L., Watson, G.E., Pickering, I.J., Singh, S.P., Myers, G.J., Clarkson,  
625 T.W., George, G.N., 2010. The chemical nature of mercury in human brain following poisoning  
626 or environmental exposure. ACS chemical neuroscience 1, 810-818.

- 627 Ng, T.Y.-T., Wood, C.M., 2008. Trophic transfer and dietary toxicity of Cd from the oligochaete  
628 to the rainbow trout. *Aquat. Toxicol.* 87, 47-59.
- 629 NMFS, 2010. Endangered and threatened wildlife and plants: threatened status for the Puget  
630 Sound/Georgia Basin distinct population segments of yelloweye and canary rockfish and  
631 endangered status for the Puget Sound/Georgia Basin distinct population segment of bocaccio  
632 rockfish, Final Rule. *Federal Register*, pp. 22276–22290.
- 633 Olsson, P.E., Hogstrand, C., 1987. Subcellular distribution and binding of cadmium to  
634 metallothionein in tissues of rainbow trout after exposure to <sup>109</sup>Cd in water. *Environ. Toxicol.*  
635 *Chem.* 6, 867-874.
- 636 Palmisano, F., Cardellicchio, N., Zambonin, P., 1995. Speciation of mercury in dolphin liver: a  
637 two-stage mechanism for the demethylation accumulation process and role of selenium. *Mar.*  
638 *Environ. Res.* 40, 109-121.
- 639 Peng, X., Liu, F., Wang, W.X., 2016. Organ-specific accumulation, transportation and  
640 elimination of methylmercury and inorganic mercury in a low Hg accumulating fish. *Environ.*  
641 *Toxicol. Chem.*
- 642 Ponton, D.E., Caron, A., Hare, L., Campbell, P.G., 2016. Hepatic oxidative stress and metal  
643 subcellular partitioning are affected by selenium exposure in wild yellow perch (*Perca*  
644 *flavescens*). *Environ. Pollut.* 214, 608-617.
- 645 Ralston, N.V., Blackwell, J.L., Raymond, L.J., 2007. Importance of molar ratios in selenium-  
646 dependent protection against methylmercury toxicity. *Biological Trace Element Research* 119,  
647 255-268.
- 648 Rogers, J., Richards, J., Wood, C., 2003. Ionoregulatory disruption as the acute toxic mechanism  
649 for lead in the rainbow trout (*Oncorhynchus mykiss*). *Aquat. Toxicol.* 64, 215-234.
- 650 Rosabal, M., Hare, L., Campbell, P.G., 2012. Subcellular metal partitioning in larvae of the insect  
651 *Chaoborus* collected along an environmental metal exposure gradient (Cd, Cu, Ni and Zn).  
652 *Aquat. Toxicol.* 120, 67-78.

- 653 Rosabal, M., Hare, L., Campbell, P.G., 2014. Assessment of a subcellular metal partitioning  
654 protocol for aquatic invertebrates: preservation, homogenization, and subcellular fractionation.  
655 *Limnol. Oceanogr. Methods* 12, 507-518.
- 656 Rosabal, M., Mounicou, S., Hare, L., Campbell, P.G.C., 2016. Metal (Ag, Cd, Cu, Ni, Tl, Zn)  
657 binding to cytosolic biomolecules in field-collected larvae of the insect *Chaoborus*. *Environ. Sci.*  
658 *Technol.* 50, 3247-3255.
- 659 Rosabal, M., Pierron, F., Couture, P., Baudrimont, M., Hare, L., Campbell, P.G., 2015.  
660 Subcellular partitioning of non-essential trace metals (Ag, As, Cd, Ni, Pb, and Tl) in livers of  
661 American (*Anguilla rostrata*) and European (*Anguilla anguilla*) yellow eels. *Aquat. Toxicol.* 160,  
662 128-141.
- 663 Sánchez-Marín, P., Beiras, R., 2017. Subcellular distribution and trophic transfer of Pb from  
664 bivalves to the common prawn *Palaemon serratus*. *Ecotoxicol. Environ. Saf.* 138, 253-259.
- 665 Sandheinrich, M., Wiener, J., 2011. Methylmercury in freshwater fish: recent advances in  
666 assessing toxicity of environmentally relevant exposures. *Environmental Contaminants in Biota:*  
667 *Interpreting Tissue Concentrations* 2, 169-190.
- 668 Sasakura, C., Suzuki, K.T., 1998. Biological interaction between transition metals (Ag, Cd and  
669 Hg), selenide/sulfide and selenoprotein P. *Journal of inorganic biochemistry* 71, 159-162.
- 670 US Environmental Protection Agency, 2007. Method 7473: Mercury in solids and solutions by  
671 thermal decomposition, amalgamation, and atomic absorption spectrophotometry, Washington,  
672 D.C.
- 673 Wallace, W.G., Lee, B.-G., Luoma, S.N., 2003. Subcellular compartmentalization of Cd and Zn  
674 in two bivalves. I. Significance of metal-sensitive fractions (MSF) and biologically detoxified  
675 metal (BDM). *Mar. Ecol. Prog. Ser.* 249, 183-197.
- 676 Wang, Z., Feng, C., Ye, C., Wang, Y., Yan, C., Li, R., Yan, Y., Chi, Q., 2016. Subcellular  
677 partitioning profiles and metallothionein levels in indigenous clams *Moerella iridescens* from a  
678 metal-impacted coastal bay. *Aquat. Toxicol.* 176, 10-23.

- 679 Wang, F., Lemes, M., Khan, M., 2012. Metallomics of mercury: Role of thiol-and selenol-  
680 containing biomolecules. *Environmental Chemistry and Toxicology of Mercury*, 517.
- 681 Wang, Y., Wu, Y., Luo, K., Liu, Y., Zhou, M., Yan, S., Shi, H., Cai, Y., 2013. The protective  
682 effects of selenium on cadmium-induced oxidative stress and apoptosis via mitochondria pathway  
683 in mice kidney. *Food Chem. Toxicol.* 58, 61-67.
- 684 Watanabe, T., Hirano, S., 2013. Metabolism of arsenic and its toxicological relevance. *Arch.*  
685 *Toxicol.* 87, 969-979.
- 686 Wolke, R., 1992. Piscine macrophage aggregates: a review. *Annu. Rev. Fish Dis.* 2, 91-108.
- 687 Zhang, W., Huang, L., Wang, W.-X., 2012. Biotransformation and detoxification of inorganic  
688 arsenic in a marine juvenile fish *Terapon jarbua* after waterborne and dietborne exposure. *J.*  
689 *Hazard. Mater.* 221, 162-169.
- 690 Zhang, L., Wang, W.-X., 2006. Significance of subcellular metal distribution in prey in  
691 influencing the trophic transfer of metals in a marine fish. *Limnol. Oceanogr.* 51, 2008-2017.
- 692
- 693

### Figure Captions

- Figure 1.** Relationships between total hepatic trace element concentration (x-axes) and trace element concentrations in subcellular fractions (y-axes) isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Upper panels (A, C, E) represent sensitive fractions and the lower panels (B, D, F) represent detoxified fractions. The various subcellular fractions are: mitochondria (filled triangles), microsomes and lysosomes (white triangles), heat-denatured proteins (white stars), heat-stable proteins (white squares), and granule-like (black squares). Points represent means of replicate livers and error bars represent standard deviations. Lines represent statistically significant regressions ( $P < 0.05$ ). Points in boxes are outliers and were excluded from regressions.
- Figure 2.** Mean relative contributions of cadmium (Cd), and lead (Pb) in various subcellular fractions isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Subcellular fraction abbreviations are: MITO = mitochondria, M+L = microsomes and lysosomes, HDP = heat-denatured proteins, DEB = debris and nuclei, HSP = heat-stable proteins, and GRAN = granule-like. Note that arsenic (As) data are not presented as the relative proportions of As varied significantly in both the HSP and M+L fractions along the bioaccumulation gradient.
- Figure 3.** Relationships between total hepatic trace element concentration (x-axes) and trace element concentrations in subcellular fractions (y-axes) isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Upper panels (G and I) represent sensitive fractions and the lower panels (H and J) represent detoxified fractions. The various subcellular fractions are: mitochondria (filled triangles), microsomes and lysosomes (white triangles), heat-denatured proteins (white stars), heat-stable proteins (white squares), and granule-like (black squares). Points represent means of replicate livers and error bars represent standard deviations. Lines represent statistically significant

regressions ( $P < 0.05$ ). Points in boxes are outliers and were excluded from regressions.

**Figure 4.** Mean relative contributions of mercury (Hg) and selenium (Se) in various subcellular fractions isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Subcellular fraction abbreviations are: MITO = mitochondria, M+L = microsomes and lysosomes, HDP = heat-denatured proteins, DEB = debris and nuclei, HSP = heat-stable proteins, and GRAN = granule-like.

**Figure 5.** Principal Component Analysis (PCA) based on trace element concentrations in sensitive (A) and detoxified (B) fractions isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*).

**Figure S1.** Flow chart describing the partitioning procedure used to separate yelloweye rockfish livers into subcellular fractions.

**Figure S2.** Relationships between total hepatic trace element concentration (x-axes) and relative contributions of subcellular fractions (y-axes) isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Upper panels (A, C, E) represent sensitive fractions and the lower panels (B, D, F) represent detoxified fractions. The various subcellular fractions are: mitochondria (filled triangles), microsomes and lysosomes (white triangles), heat-denatured proteins (white stars), heat-stable proteins (white squares), and granule-like (black squares). Points represent means of replicate livers and error bars represent standard deviations. Lines represent statistically significant regressions ( $P < 0.05$ ). Points in boxes are outliers and were excluded from regressions.

**Figure S3.** Relationships between total hepatic trace element concentration (x-axes) and the relative contributions of subcellular fractions (y-axes) isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Upper panels (A, C) represent sensitive fractions and the lower panels (B, D) represent detoxified fractions. The various subcellular fractions are: mitochondria (filled triangles), microsomes and lysosomes

(white triangles), heat-denatured proteins (white stars), heat-stable proteins (white squares), and granule-like (black squares). Points represent means of replicate livers and error bars represent standard deviations. Points in boxes are outliers and were excluded from regressions.

694

ACCEPTED MANUSCRIPT

Figure 1. Relationships between total hepatic trace element concentration (x-axes) and trace element concentrations in subcellular fractions (y-axes) isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Upper panels (A, C, E) represent sensitive fractions and the lower panels (B, D, F) represent detoxified fractions. The various subcellular fractions are: mitochondria (filled triangles), microsomes and lysosomes (white triangles), heat-denatured proteins (white stars), heat-stable proteins (white squares), and granule-like (black squares). Points represent means of replicate livers and error bars represent standard deviations. Lines represent statistically significant regressions ( $P < 0.05$ ). Points in boxes are outliers and were excluded from regressions.

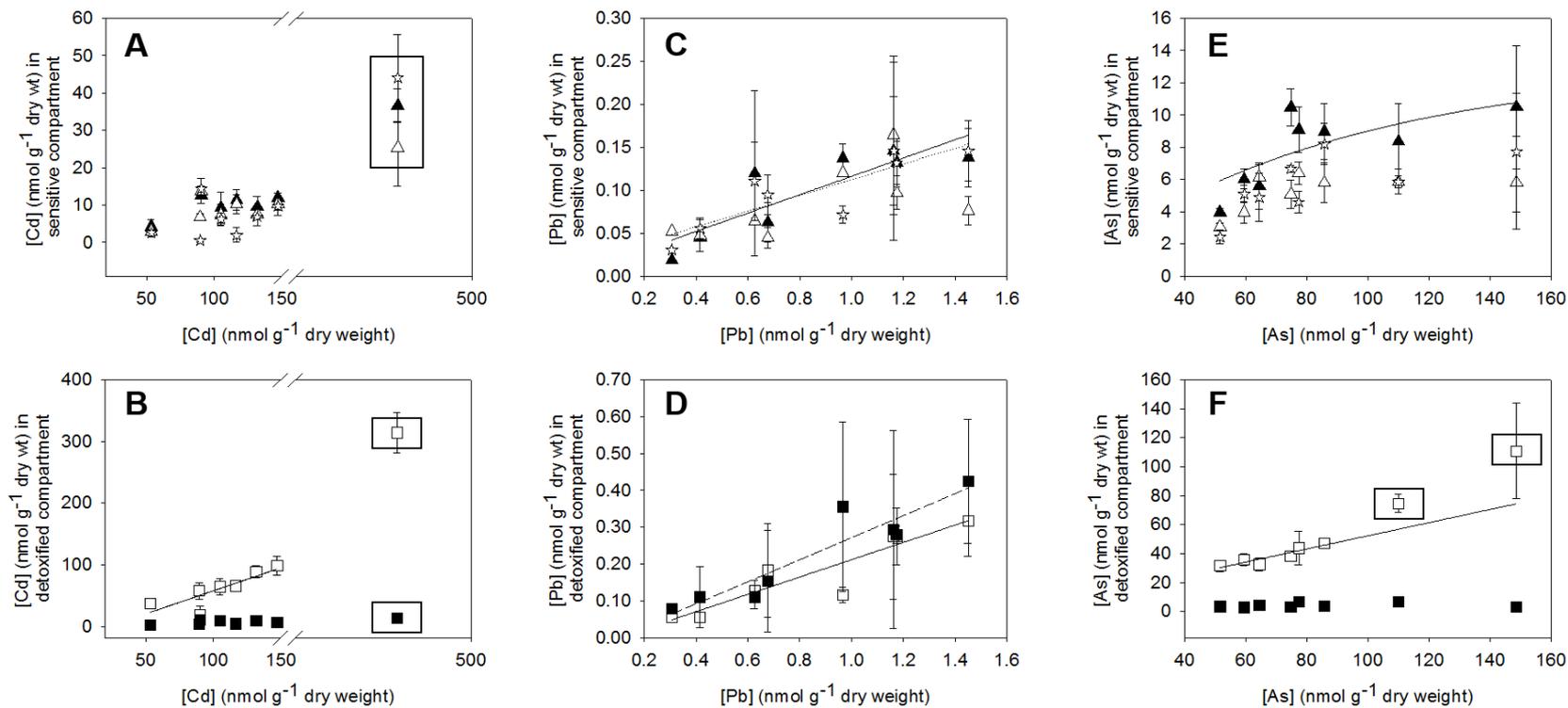


Figure 2. Mean relative contributions of cadmium (Cd), and lead (Pb) in various subcellular fractions isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Subcellular fraction abbreviations are: MITO = mitochondria, M+L = microsomes and lysosomes, HDP = heat-denatured proteins, DEB = debris and nuclei, HSP = heat-stable proteins, and GRAN = granule-like. Note that arsenic (As) data are not presented as the relative proportions of As varied significantly in both the HSP and M+L fractions along the bioaccumulation gradient.

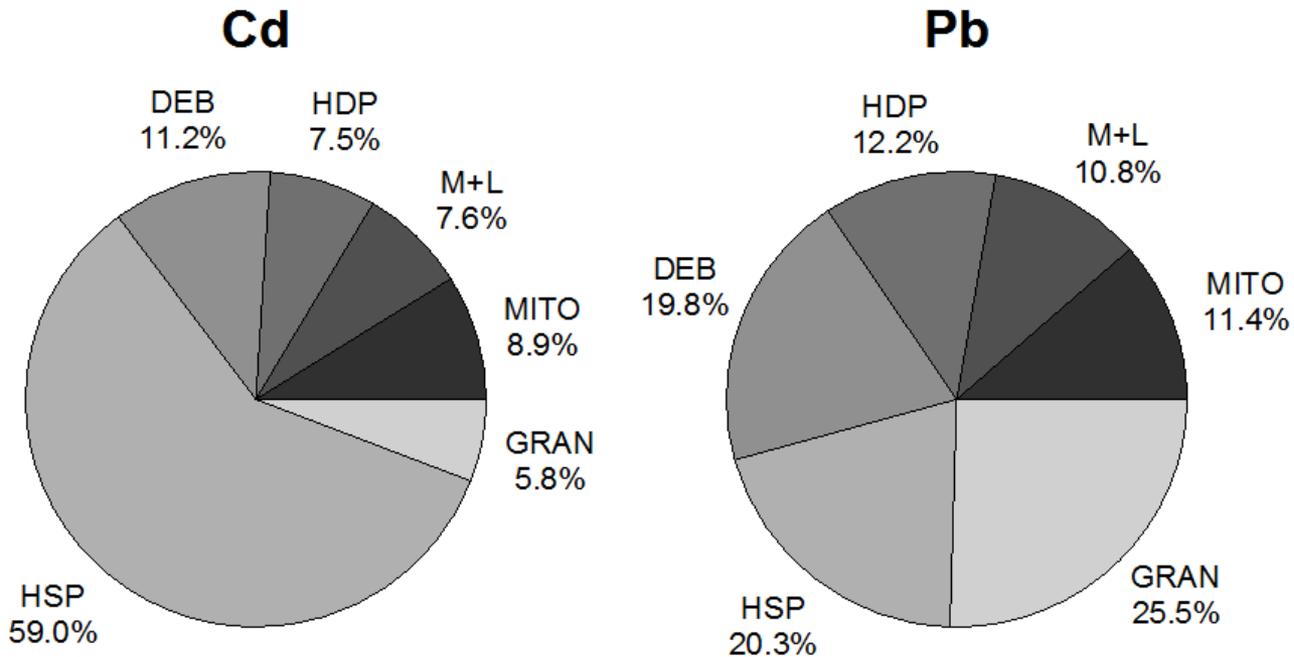


Figure 3. Relationships between total hepatic trace element concentration (x-axes) and trace element concentrations in subcellular fractions (y-axes) isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Upper panels (G and I) represent sensitive fractions and the lower panels (H and J) represent detoxified fractions. The various subcellular fractions are: mitochondria (filled triangles), microsomes and lysosomes (white triangles), heat-denatured proteins (white stars), heat-stable proteins (white squares), and granule-like (black squares). Points represent means of replicate livers and error bars represent standard deviations. Lines represent statistically significant regressions ( $P < 0.05$ ). Points in boxes are outliers and were excluded from regressions.

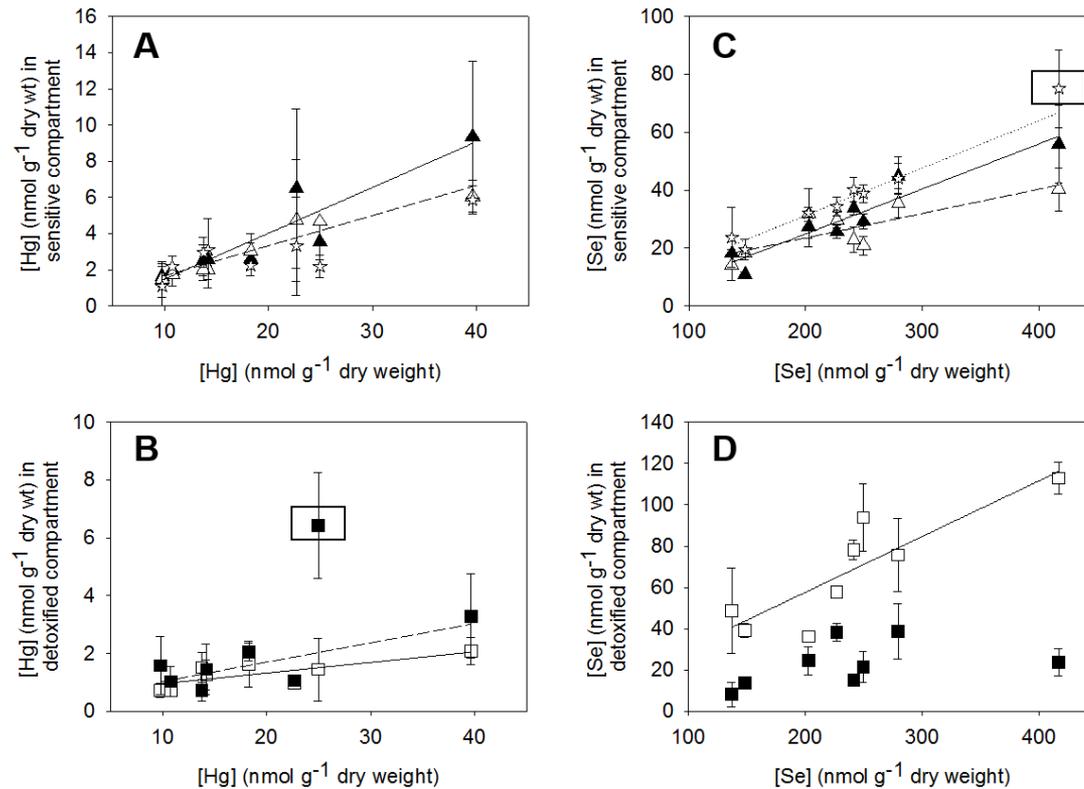


Figure 4. Mean relative contributions of mercury (Hg) and selenium (Se) in various subcellular fractions isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Subcellular fraction abbreviations are: MITO = mitochondria, M+L = microsomes and lysosomes, HDP = heat-denatured proteins, DEB = debris and nuclei, HSP = heat-stable proteins, and GRAN = granule-like.

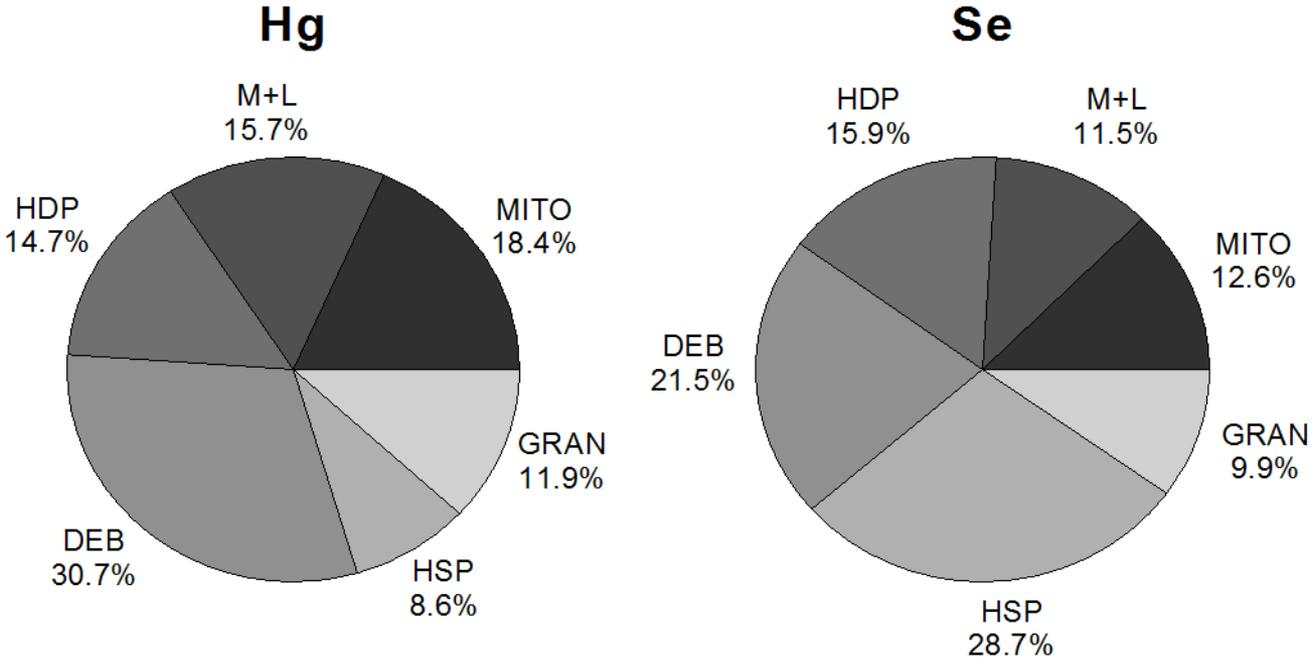


Figure 5. Principal Component Analysis (PCA) based on trace element concentrations in sensitive (A) and detoxified (B) fractions isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*).

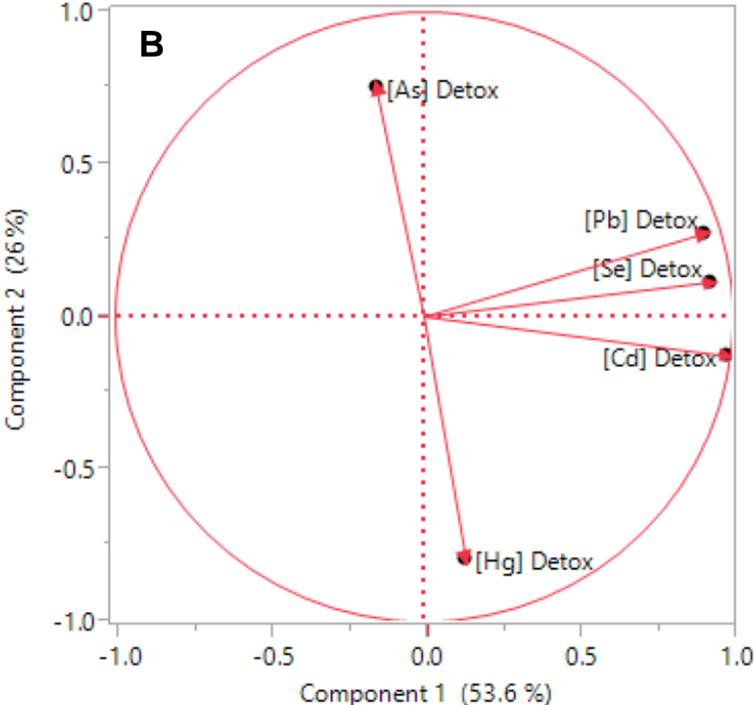
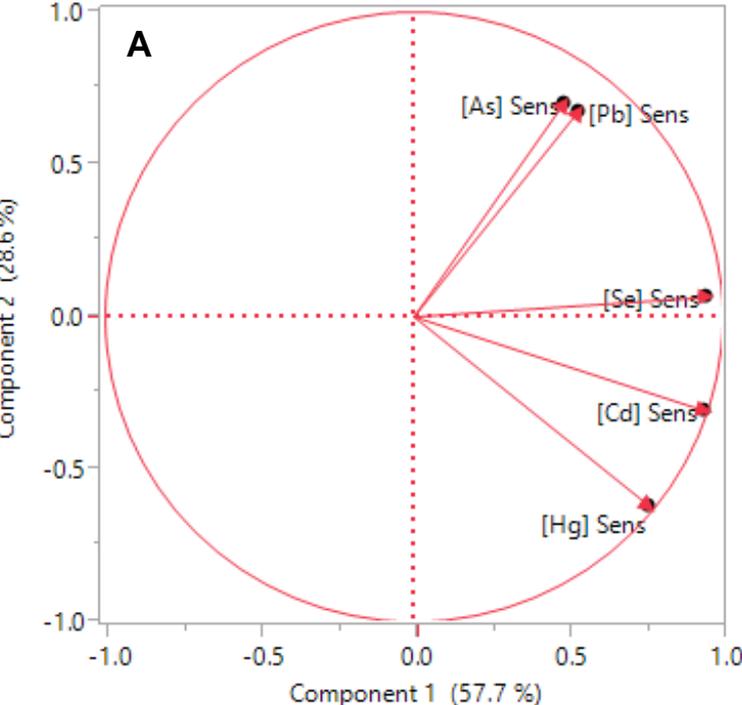


Figure S1. Flow chart describing the partitioning procedure used to separate yelloweye rockfish livers into subcellular fractions.

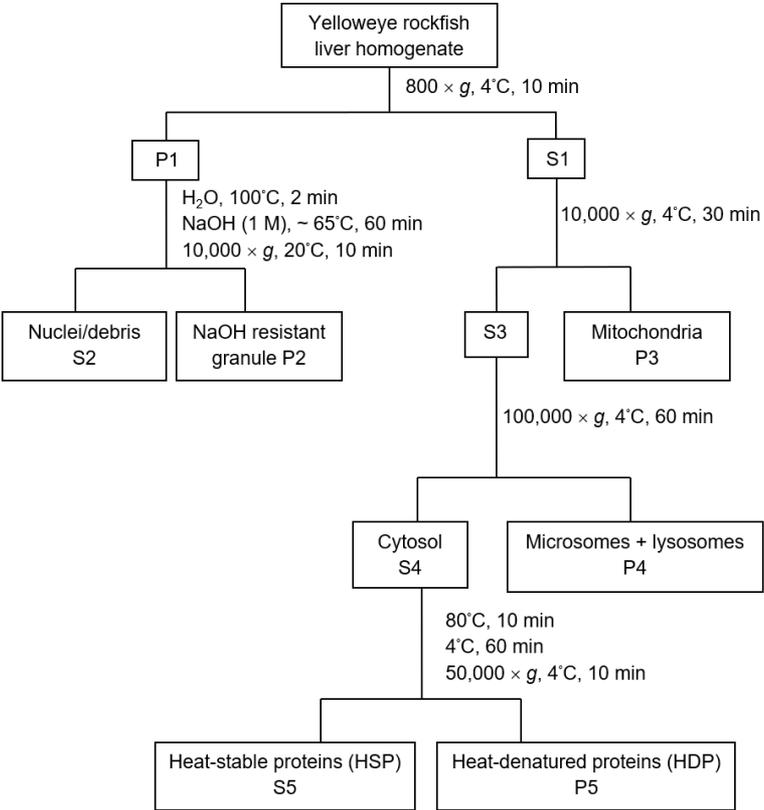


Figure S2. Relationships between total hepatic trace element concentration (x-axes) and relative contributions of subcellular fractions (y-axes) isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Upper panels (A, C, E) represent sensitive fractions and the lower panels (B, D, F) represent detoxified fractions. The various subcellular fractions are: mitochondria (filled triangles), microsomes and lysosomes (white triangles), heat-denatured proteins (white stars), heat-stable proteins (white squares), and granule-like (black squares). Points represent means of replicate livers and error bars represent standard deviations. Lines represent statistically significant regressions ( $P < 0.05$ ). Points in boxes are outliers and were excluded from regressions.

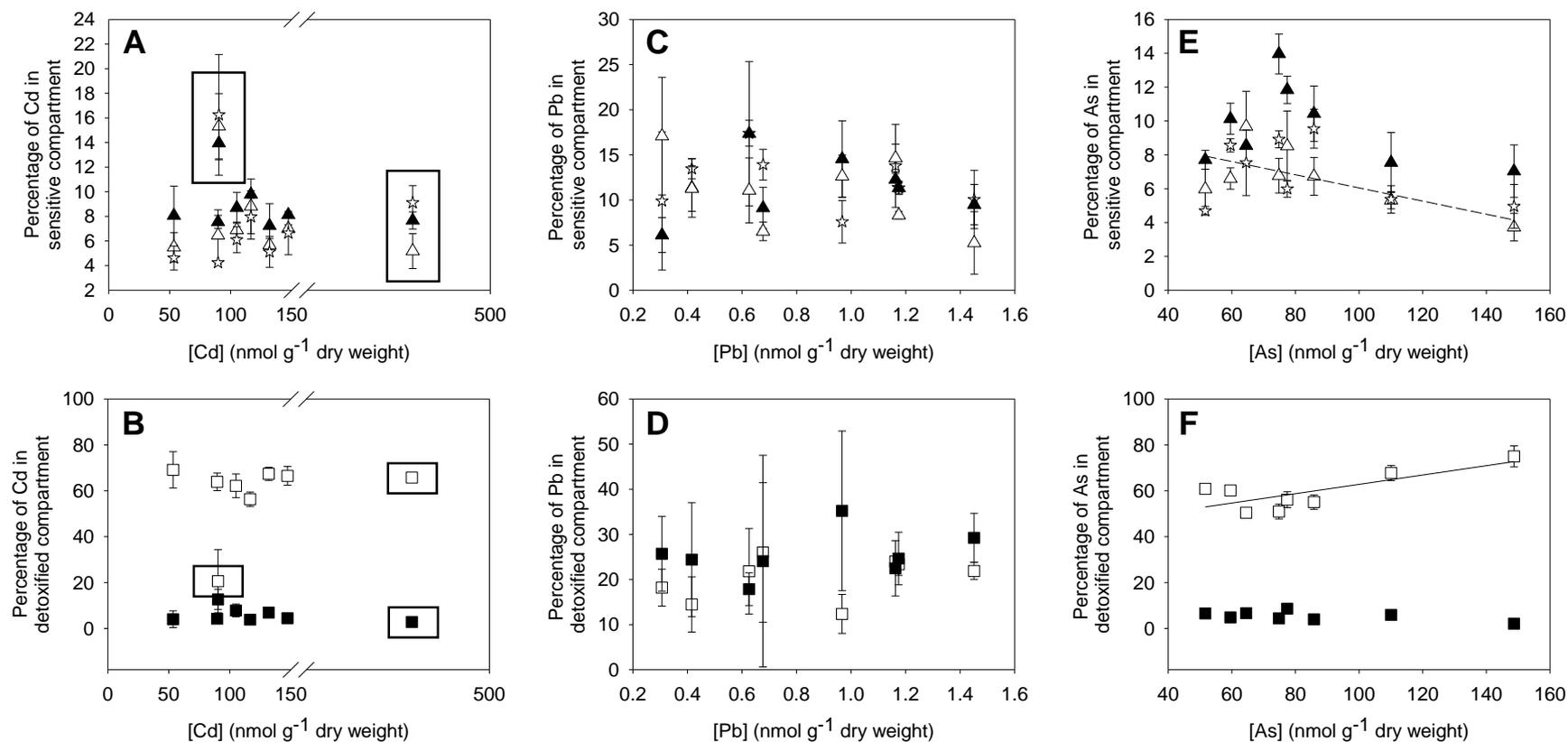
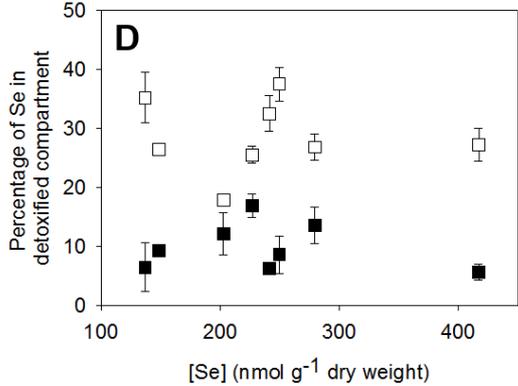
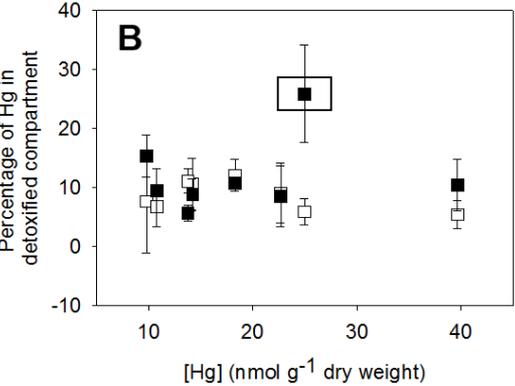
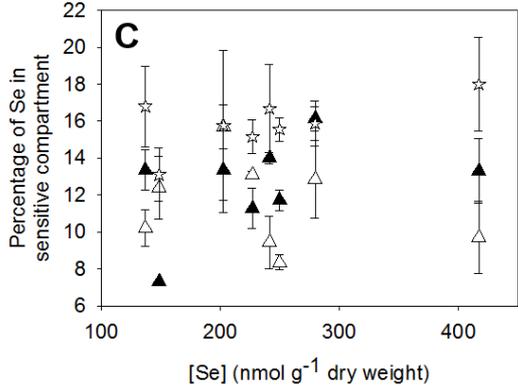
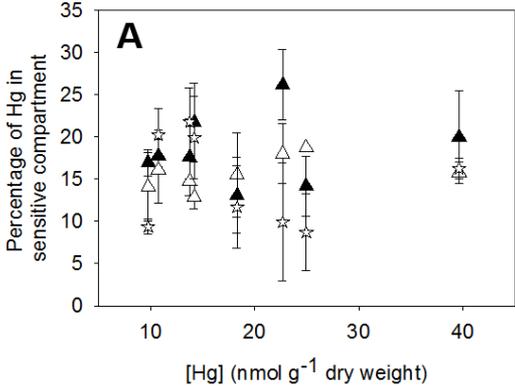


Figure S3. Relationships between total hepatic trace element concentration (x-axes) and the relative contributions of subcellular fractions (y-axes) isolated from the livers of yelloweye rockfish (*Sebastes ruberrimus*). Upper panels (A, C) represent sensitive fractions and the lower panels (B, D) represent detoxified fractions. The various subcellular fractions are: mitochondria (filled triangles), microsomes and lysosomes (white triangles), heat-denatured proteins (white stars), heat-stable proteins (white squares), and granule-like (black squares). Points represent means of replicate livers and error bars represent standard deviations. Points in boxes are outliers and were excluded from regressions.



**Highlights**

- Subcellular partitioning of Cd, Pb, As, Hg, and Se was determined in the livers of yelloweye rockfish.
- Though non-essential elements were found in detoxified fractions, the extent of binding differed among elements.
- Mercury may be of particular concern, as it was present mainly in sensitive sites.