

Big Data for a Big Country: The Second Generation Canadian Wetland Inventory Map at 10 Meters Resolution

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Abstract

Recently, there has been a significant increase in efforts to better inventory and manage important ecosystems across Canada using advanced remote sensing techniques. In this study, we improved the method used in creating the first generation Canadian wetland inventory map at 10-m resolution. The main contribution of this study, as it compares to the previous one, is training Random Forest (RF) models on the Google Earth Engine (GEE) platform within the boundaries of ecozones rather than provinces, in order to increase wetland classification accuracy. The ecozone boundaries divide the Canadian landscape based on similar biotic and abiotic factors, i.e., land cover, human activity, climate, wildlife, soil, vegetation, and geomorphology. Therefore, it should produce more accurate and meaningful wetland classification results. In the first generation of this product, there was a lack of training data in some ecozones, making it impossible to apply the classification method at the ecozone level, as training data is a significant bottleneck in the machine learning algorithms. In this study, a considerable effort has been devoted to data collection, preparation, standardization of datasets for each ecozone. The result of data cleaning reveals a data gap in several Northern ecozones. Accordingly, high-resolution optical data, from Worldview-2 and Pleiades, were acquired to delineate wetland training data based on visual interpretation in those regions. By using this well-distributed training data, the second generation of a Canadian wetland inventory map was improved by an overall accuracy approaching 86%. This wetland map represents an improvement of 7% compared to the first generation map. Accuracy varied from 76% to 91% in different ecozones, depending on available resources. Furthermore, the results of RF variable importance, which was carried out for each ecozone, demonstrate that $\frac{|S_{VV}|^2}{|S_{VH}|^2}$ and NDVI extracted from Sentinel-1 and Sentinel-2 data, respectively, were the most important features for wetland mapping.

Keywords: Wetland classification, remote sensing, Random Forest, Google Earth Engine, Sentinel, feature extraction, Canada

42

43 **1. Introduction**

44 Until very recently, land cover mapping at large scales has been a challenging, and in some cases,
45 an impossible task, given the required costs and resources for image analysis (Hu et al., 2017). In
46 particular, collecting, storing and processing the datasets required to cover large geographic areas,
47 and the hardware limitations associated with such data processing, were a significant barrier for
48 the production of large-scale land cover maps (Mahdianpari et al., 2020; Shelestov et al., 2017).
49 This issue is often referred to as the *geo big data* problem and is currently being addressed through
50 the application of newly available technologies and resources designed for best managing large
51 volumes of geospatial imagery (Gorelick et al., 2017a).

52 Fortunately, the ever-increasing availability of high-resolution open-access Earth Observation
53 (EO) data and powerful cloud computing resources provide unprecedented opportunities for
54 applications at spatial and temporal scales previously impossible in the geospatial sciences
55 (Mahdianpari et al., 2018; Zhou et al., 2020). For example, data collected from the Copernicus
56 programs by the European Space Agency (ESA) through the Sentinel missions have contributed
57 significantly to the global monitoring of the environment over the past few years (Aschbacher and
58 Milagro-Pérez, 2012). The accessibility and usability of these and other open-access EO data
59 across large geographic areas and at high temporal frequencies has been made possible via
60 advances in cloud computing resources, such as NASA Earth Exchange, Amazon's Web Services,
61 Microsoft's Azure, and Google cloud platforms (Liu, 2015). Among these cloud computing
62 resources, Google Earth Engine (GEE) has been recognized as a well-established, open-access tool
63 that hosts a vast pool of satellite imagery and offers tools for advanced web-based algorithm
64 developments and result visualization (Shelestov et al., 2017). These developments have now
65 made it possible for the Earth to be mapped at a large geographical scale, opening up research

66 possibilities in the ocean and ecological sciences, as well as in natural resource management
67 (Aschbacher and Milagro-Pérez, 2012; Chen et al., 2017; Mahdianpari et al., 2020, 2018; Sidhu et
68 al., 2018; Zhou et al., 2020), to name only a few.

69 Nation-wide wetland inventory development, and in turn wetland management, monitoring, and
70 conservation, is one of the numerous areas that are expected to benefit from the increasing
71 availability of big data technologies. This new technology is of particular importance for countries
72 with extensive wetland coverage, such as Canada (Mahdianpari et al., 2020). Prior to 2019, a
73 majority of Canada's wetland inventories were created at local, regional, and provincial scales, for
74 example (DeLancey et al., 2020; Dingle Robertson et al., 2015; Jahncke et al., 2018; Mahdianpari
75 et al., 2018; Millard and Richardson, 2015; Mohammadimanesh et al., 2018b; Rezaee et al., 2018;
76 White et al., 2017). Many of these inventories were derived using a variety of methods (e.g., visual
77 assessment, optical and/or RADAR, topographical, and field-work), wetland definitions (Chen et
78 al., 2010; van der Kamp et al., 2016) classification systems (Alberta Environment and Sustainable
79 Resource Development, 2015; Ducks Unlimited Canada, 2014; Gerbeaux et al., 2016; National
80 Wetlands Working Group, 1997), and under various contexts were constrained by budgets,
81 available resources, locations, and objectives. While useful under some circumstances, the
82 methods used and purposes of these inventories impact their applicability within national or global
83 contexts (Hu et al., 2017). These issues, along with spectral and structural similarities between
84 various types of wetlands, and the lack of clear-cut borders between successional wetland classes,
85 have limited the capability of the machine learning tools for large-scale wetland mapping and
86 resulted in insufficient classification accuracies in some cases (Hu et al., 2017). Other issues arise
87 when comparing and contrasting spatial wetland information across political, geographical, or

88 disciplinary boundaries which can in-turn impact the quality, development and assessment of
89 wetland-related management and policies (Fournier et al., 2007; Hu et al., 2017).

90 Another major issue related to wetland mapping at national and global scales is the collection of
91 sufficient high-quality reference data (Mahdianpari et al., 2020). Developing a quality nation-wide
92 wetland inventory using supervised remote sensing methods requires a large amount of training
93 and testing data distributed across the entire country, to best represent Canada's expansive and
94 diverse landscape (Statistics Canada, 2018). Like many of Canada's wetland inventories, most
95 available training and testing data have been collected under a variety of contexts, using different
96 local and regional wetland definitions, for a number of purposes (often not remote sensing
97 focused), and using a variety of different methods. Additionally, obtaining such data is not always
98 a simple task, requiring the willing contribution of numerous collaborators and/or the collection of
99 freely available data with variable metadata quality or sometimes limited explanatory information.
100 While these discrepancies are an issue, they are not unexpected and as a result, training and testing
101 data in a large-scale study will require collaboration, substantial editing, and standardization. Other
102 issues include gathering accurate non-wetland land cover information which often requires the use
103 of freely available datasets and visual interpretation of satellite imagery available via Google
104 Earth. Like the wetland datasets, the non-wetland land cover data requires standardization in terms
105 of naming conventions, definitions, and polygon boundaries. The development of the training and
106 testing dataset is of utmost-importance, as the quality and accuracy of these inputs are ultimately
107 reflected in the final inventory output (Millard and Richardson, 2015; Mui et al., 2015).

108 In the face of increasing globalization, continued wetland loss, increasing population, urban
109 sprawl, and human-induced climate change, the importance and availability of consistent and
110 reliable large-scale wetland inventories both in Canada and around the globe has never been

111 greater. Such large-scale inventories will contribute to the improvement of the nation- and global-
112 wide wetland management, protection initiatives, and policies, allow for consistent estimations of
113 yearly trends in wetland loss or gain, analysis of biodiversity, and help improve the outputs of
114 large-scale climate models and estimates (Erwin, 2009).

115 Therefore, the overarching goal of the current study was to leverage state-of-the-art remote sensing
116 tools for the production of large-scale wetland inventory maps for Canada. Specifically, the main
117 objectives are to: (1) prepare structured, cleaned, consistent, and well-distributed training and
118 testing data for each of Canada's ecozones; (2) produce the second generation Canada-wide
119 wetland inventory; (3) improve the wetland classification accuracy compared to the first
120 generation Canadian wetland inventory map by running classifications within ecozones rather than
121 provincial boundaries; and (4) determine the most important features for national wetland mapping
122 via RF algorithms using built-in capacities in GEE.

123 **2. Methodology**

124 2.1. Study Area

125 The Ecological Framework of Canada (Statistics Canada, 2018), which delineates ecologically
126 distinct areas across Canada, defines a total of 15 ecozones. Ecozones represent areas of Canada's
127 land surface characterized by interacting abiotic and biotic factors. These ecozones are displayed
128 in Figure 1. The size of these ecozones ranges from 117,240 km² (Mixed wood Plains) to 1,857,530
129 km² (Boreal Shield). Please refer to Table 1 for a summary of the general characteristics of each
130 ecozone. Note that the three northern ecozones (Southern Arctic, Northern Arctic, and the Arctic
131 Cordillera) are referred to as the Northern Ecozones throughout the remainder of this study. These
132 three ecozones are grouped together for purposes of reference data development, processing, and
133 classification as a result of the limited available wetland data for this area. Additionally, the Boreal

134 Shield was split into two areas (east and west), and the Boreal and Taiga Cordillera ecozones were
 135 merged into one (Boreal/Taiga Cordillera), for processing and training data development purposes.
 136 The reasoning for this is discussed in section 2.2.

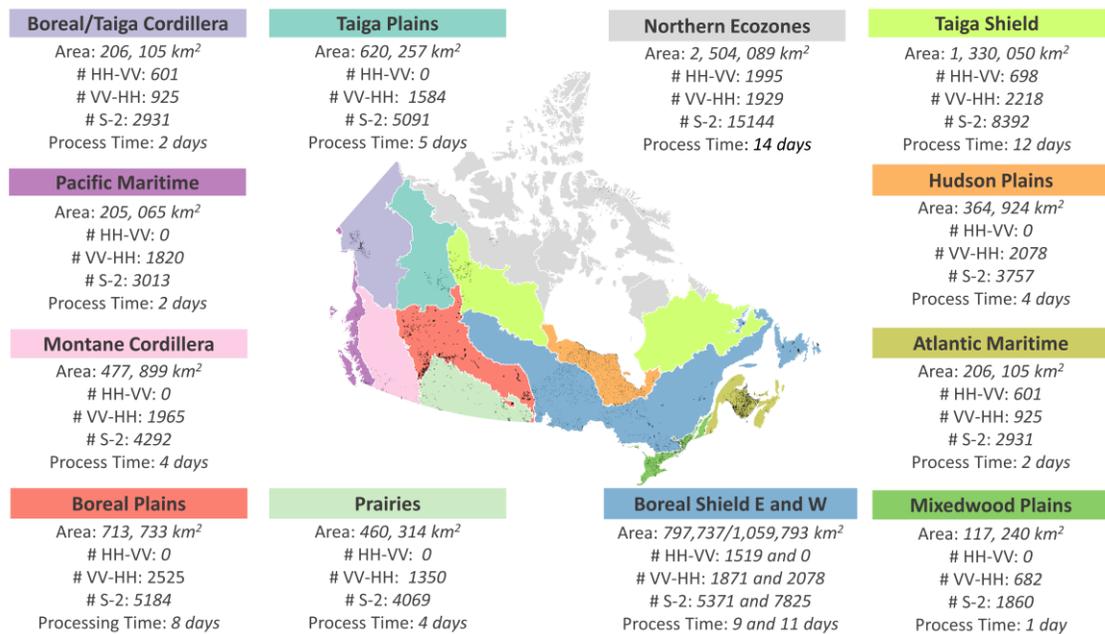


Figure 1. Canadian ecozones, ecozone sizes, and the processing time and the number of Sentinel-1 and -2 images required to produce classifications. Reference data distribution across Canada is displayed in black.

137 **Table 1.** A summary of the typical land cover characteristics of each ecozone (Ecosystem Classification Group, 2010;
 138 Environment and Climate Change Canada, 2016; Federal, Provincial, and Territorial Governments of Canada, 2010;
 139 Smith et al., 2004).

Ecozone	Spatial Location	Description
Atlantic Maritime (AM)		Has a typical maritime climate, generally cool and wet year round. The most common land cover type here is the forest. Agricultural activity is the most common human activity. The most common wetlands in this area are treed (swamp, bog, and fen).
Boreal and Taiga Cordillera (Boc / TC)		Summers are short and cool, and winters are long and cold. Dominating land cover includes extensive mountains and tundra to the north and forests to the south. Wetlands, and particularly peatlands, are less common here than the neighboring Taiga Plains. Forest and wetlands are most common in valleys and slopes.
Boreal Plains (BP)		Has a typical continental climate, with cold winters and cool summers. Forest is the most common natural land cover type and agriculture the most common anthropogenic land cover. Agricultural activity is largely present along the southern edge of the ecozone and to the north-east. The most common types of wetlands include conifer swamps, fens, and bogs.

Boreal Shield (BS)		Moderate summer and winter temperatures. The largest ecozone in Canada. Low elevation land dominated by forest and shrubland with relatively minimal anthropogenic land cover. Peatlands, including bog and fen, are the most common, particularly on the eastern side.
Hudson Plains (HP)		Has a maritime climate, and as a result, extensive wetlands are present, particularly peatlands. Marsh is more common along with the northern coast. This area is often referred to as Canada's largest wetland complex. There is relatively little forest cover present.
Mixedwood Plains (MP)		The most populated ecozone characterized by a climate of warm summers and cool winters. The landscape is generally flat and dominated by extensive agricultural land cover. Most wetland cover is located along the edge of the ecozone and the northeast. Swamp, bog, and fen are the most common.
Montane Cordillera (MC)		The most diverse topography and climate relative to other ecozones, with various mountain ranges present. Forest covers over half of the total land surface. There is relatively little wetland coverage and is mostly located along rivers and in valleys.
Northern Ecozones (NE)		Characterized by very low temperatures, permafrost, and limited vegetation. Mountains and glaciers dominate the furthest north, giving way to tundra barrens, hills, and plains to the south. There is relatively little human presence in these areas. Wetlands, particularly peatlands, are dispersed throughout the barrens and along waterways.
Pacific Maritime (PM)		Located along the coast of the Pacific Ocean with a mountainous maritime climate. The Coast mountains and extensive forests dominate most of this ecozone. Most anthropogenic land cover is located at the southern end of the ecozone. There are relatively few wetlands here.
Prairies (Pr)		More variable climate than other ecozones. Almost entirely covered in agriculture. The most common natural land cover is grassland. There are very few wetlands located here, having been lost to agricultural development. Wetlands that are present are very small "prairie potholes."
Taiga Plains (TP)		Largely flat area. The colder climate in the north part of the ecozone versus the warmer south. Most land cover is boreal forest and shrub, and there is a relatively small human presence. Wetlands of many types are widespread, including large deltas, swamps along rivers, peatlands, and marsh.
Taiga Shield (TS)		Open forest that transitions to shrub and tundra moving north. Temperatures are colder in the west versus the east. There is relatively little human activity here. Wetlands make up an estimated 13% of [SH1] in this area, though there are trends indicating wetland expansion due to changes in weather patterns and permafrost melting.

140

141 2.2. Reference Data

142 Broadly, the development of the reference data for this study required, for each ecozone, a dataset

143 comprised of accurately-delineated polygons representing bog, fen, swamp, and marsh wetland

144 classes, and polygons representing the most dominant non-wetland land-use. Generally, the

145 wetland data for this study was gathered from multiple collaborators across Canada, and the non-
146 wetland data was derived via visual polygon delineation with the aid of the Agriculture and Agri-
147 foods Canada 2018 Crop Inventory map (Agriculture and Agri-food Canada, 2018), with some
148 exceptions which are discussed below.

149 The wetland data for this study was acquired from a number of sources across Canada. Ultimately,
150 these wetland data were used to produce training and validation datasets for each ecozone. These
151 datasets were collected for a variety of purposes, over several years, at different scales, and using
152 different field, classification, and polygon delineation methods. As a result, the distribution and
153 amount of data available within each ecozone vary considerably (see Figure 1). For these reasons,
154 the datasets needed to go through several rounds of editing before being functionally incorporated
155 into an ecozone final reference dataset.

156 As a first step, the data were filtered to remove any polygons smaller than 1 hectare and greater
157 than 100 hectares, as small polygons would not contain any helpful spectral information for the
158 classifier according to the minimum mapping unit of this study, and the large polygons had a higher
159 chance of being highly spectrally heterogeneous. Next, some datasets were clipped to ensure that
160 each ecozone had its own specific dataset associated with it. This is because a number of these
161 datasets spanned the boundaries of multiple ecozones. Note that some ecozones did not have any
162 wetland training data located within their boundaries, and as a result, these ecozones were instead
163 classified using the reference data in an adjacent ecozone. These ecozones include the Taiga
164 Cordillera and the three northern-most ecozones. The three northern ecozones and the Taiga and
165 Boreal Cordillera ecozone boundaries were merged to create two broad multi-ecozone boundaries.
166 Additional data cleaning steps, including the standardization of naming conventions, removal of
167 some inaccurate polygons, re-classification of some polygons, and boundary modification of

168 others, were also performed. Additionally, in datasets where there were thousands of wetland
169 polygons (i.e., local wetland maps), a subset of these polygons was randomly selected for
170 incorporation into the final reference dataset.

171 Notably, there was no wetland data available to this study in the northern-most ecozones, and
172 because google earth has limited or inconsistent imagery in northern Canada, VHR imagery was
173 acquired for purposes of producing a northern ecozone wetland dataset. Wet areas along the
174 northern-coast were identified to collect coincident WorldView-2 and Pleaide's imagery for these
175 areas. An effort was made to select as the most recent summer imagery as possible, though the
176 selection was constrained by image availability, cloud cover, and cost. Because cloud-cover is a
177 significant issue in northern Canada, the most recent summer dates for which we could obtain
178 cloud-free imagery was during the summers of 2015 and 2016. Figure 2 shows some peatland and
179 swamp delineation via visual assessment in a WorldView-2 image taken near Kugluktuk, Nunavut
180 (top), and in a Pleaide's image near Bathurst Inlet (bottom). Because the assessor did not feel
181 confident differentiating between bog and fen wetlands in the imagery, all delineated peatlands
182 were referred to as fen. This imagery was essential for producing wetland data for the northern
183 ecozones; however, the dataset remained small due to the limited extents of the imagery. Non-
184 wetland classes were delineated using the VHR imagery as well.



Figure 2: Wetland delineation using VHR imagery. Top: Peatland delineation using June 29th 2016 Pleiades imagery. Bottom: Swamp delineation using June 29th, 2015 WorldView-2 imagery.

185 The Agriculture Agri-Food Canada 2018 Annual Crop Inventory map (Agriculture and Agri-food
186 Canada, 2018) guided the delineation of non-wetland polygons. As a first step, the most common
187 non-wetland land cover within each ecozone was calculated using the Crop Inventory map. Next,
188 polygons representing the most common land cover types were manually delineated, using both
189 Google Earth and the crop inventory map as a visual aid. Some ecozones did not have any coverage
190 by the crop inventory dataset, most commonly in ecozones located in the northern parts of Canada.
191 As such, visual identification of some common land cover types was conducted using the
192 interpretation of very high resolution (VHR) imagery or ancillary land cover datasets, including
193 multiple local land cover geospatial datasets and the 30 m resolution land cover map of Canada
194 provided by the Canada Centre for Mapping and Earth Observation (CCMEO). Table 2
195 summarizes the number of wetland reference polygons and their areal coverage for each ecozone.
196 To produce the final reference data, the wetland and non-wetland polygons for each ecozone were
197 randomly divided into two groups: 50% for training and 50% for testing. Specifics of the data
198 preparation for each ecozone are discussed below.

199 **Table 2.** Summary of the reference data employed for each ecozone.

Ecozones	# Wetland polygons	# Upland polygons	Discussion
AM	3000	802	Majority of the wetland reference data came from a large New Brunswick Wetlands dataset containing thousands of polygons. Removing polygons smaller than 1 hectare and greater than 100 hectares reduced the total size. Non-wetland polygons were produced using the crop inventory maps and Google Earth. The areal coverage of the polygons for each wetland class are similar.
Boc & TC	348	336	There is no Crop Inventory coverage in this area so non-wetland polygons were produced using visual assessment in Google Earth. The dataset is located in and around the Yukon communities of Haines Junction and Whitehorse. The dataset contains thousands of wetland polygons. The number of bog polygons is much smaller relative to the other classes such as a fen, marsh and swamp.
BP	200	480	Wetland data came from five datasets. The crop inventory maps guided all non-wetland land cover delineation. Because the reference data for this ecozone were derived from five different sources (unlike most of the other ecozones which had testing and training data derived from only one or two sources), it is likely that there is great variation in how bog, fen, swamp, and marsh wetlands were delineated.
BS East	612	550	Wetland data derived from multiple wetland-related datasets across various locations in Newfoundland and Labrador, originally for purposes of wetland classification using remote sensing data, using similar methods. Notably, there is a greater amount of bog wetlands, in terms of aerial coverage, versus some of the other wetlands. The crop inventory maps guided all non-wetland land cover delineation, for which this area had coverage.
BS West	2154	548	Wetland information derived from a very large wetland dataset in Ontario. Because the number of wetland polygons in this dataset was so great, after removing all wetlands less than 1 hectare and greater than 100 hectares in size, a further reduction was made by only keeping those wetlands that had been listed as being verified and evaluated. The areal coverage of the polygons in each wetland class are relatively similar. The crop inventory maps guided all non-wetland land cover delineation.
HP	2000	345	Because a large portion of the Hudson Plains ecozone fell within the province of Ontario, the Ontario wetland dataset was used to derive wetland polygons for this ecozone. Please refer to the section discussing the data for the Boreal Shield West ecozone for more information. However, because this area lacked crop inventory coverage, non-wetland polygons were delineated based on a visual assessment of Google Earth imagery.
MP	1165	600	The wetlands for this ecozone were derived from the Ontario wetland dataset. Please refer to the section discussing the data for the Boreal Shield West ecozone for more information. The crop inventory maps guided all non-wetland land cover delineation.
MC	26	209	No wetland data sourced for this ecosystem. As such, the Canadian Wetland Inventory by Ducks Unlimited (DCI), which is available online, was referred to. From the DCI, a small number of wetland polygons were gathered. Unfortunately, most of the data on the DCI map were very small and not useful for this study. Additionally, there were no bog polygons and very few fen polygons. As a result, the dataset for this ecozone is very small, relative to all other ecozones that have training datasets available.
NE	120	294	No available wetland data or crop inventory coverage of the three most northern ecozones. To address this problem, three 50 cm resolution summer images (i.e., one WorldView-2 and two Pleiades) were acquired covering some coastal low-land areas within the Southern Arctic ecozone. Using these images, visual interpretation was carried out to define wetland and non-wetlands. Because the interpreters did not feel confident in their ability to define bog wetlands in these images, only fen, swamp, and marsh polygons are present in the final dataset. While these images were certainly helpful, the amount of wetland reference data was limited by their extents.
PM	117	296	Wetland polygons for this ecozone was derived from a dataset collected in and around the Vancouver area. Relative to the marsh and swamp polygons, there was very little data for the bog class. Additionally, most of the bog polygons are derived from a single large bog, known as the Burns bog. As a result, these polygons may not be representative of other bog wetlands within the ecozone, particularly in the less-populated areas further north.
Pr	250	600	Datasets were all gathered around the Assinboine River Valley and Whitewater Lake in Manitoba. While these datasets contained a large number of wetland polygons, only a small number of them were of the appropriate size. There were also no bog polygons. The crop inventory map was used to delineate the non-wetland polygons.
TP	230	213	Datasets located within this ecozone were collected around the vicinity of Great Slave Lake in the Northwest Territories. Only half of the total polygons fell directly within the Taiga Plains ecozone (the other half fell within the Taiga Shield Ecozone). These datasets also provided training polygons for non-wetland land cover. This was welcome as this ecozone lacks any coverage by the Crop Inventory map. Using Google Earth, some additional non-wetland polygons were delineated.
TS	220	327	Wetland polygons obtained from the same dataset discussed in the Taiga Plains ecozone above. Only half of the training polygons provided by these datasets fell within the Taiga Shield. These datasets also provided training polygons for non-wetland land cover, which was welcome as this ecozone had no coverage by the Crop Inventory map. Using Google Earth, some additional non-wetland polygons were delineated via visual assessment. There was relatively little swamp data.

200 2.3. Remote Sensing Data and Image Processing

201 The Sentinel Earth Observation missions from the Copernicus program managed by the European
202 Commission in partnership with the ESA, consist of both radar and super-spectral imaging systems
203 for the land, ocean, and atmospheric monitoring. To improve the revisit time and coverage
204 capability, each mission benefits from a constellation of two satellites. In this study, the GEE data
205 catalog was used to obtain satellite imagery over our study area during 2017-2019 from Sentinel-
206 1 and Sentinel-2 data (Gorelick et al., 2017b). A total of 4,813 and 22,955 C-band Level-1 Ground
207 Range Detected (GRD) images were acquired in the HH-HV and VV-VH polarization modes of
208 Sentinel-1, respectively. Due to the mission of Sentinel-1, single-(HH) or dual-(HH-HV) polarized
209 data are collected over sea ice zones and single-(VV) or dual- (VV-VH) polarized data are
210 collected over all other observation zones (e.g., lands), we have the greater availability of VV-VH
211 compared to HH-HV polarization mode. Figure 3 demonstrates the spatial distribution of all
212 available Sentinel-1 observations.

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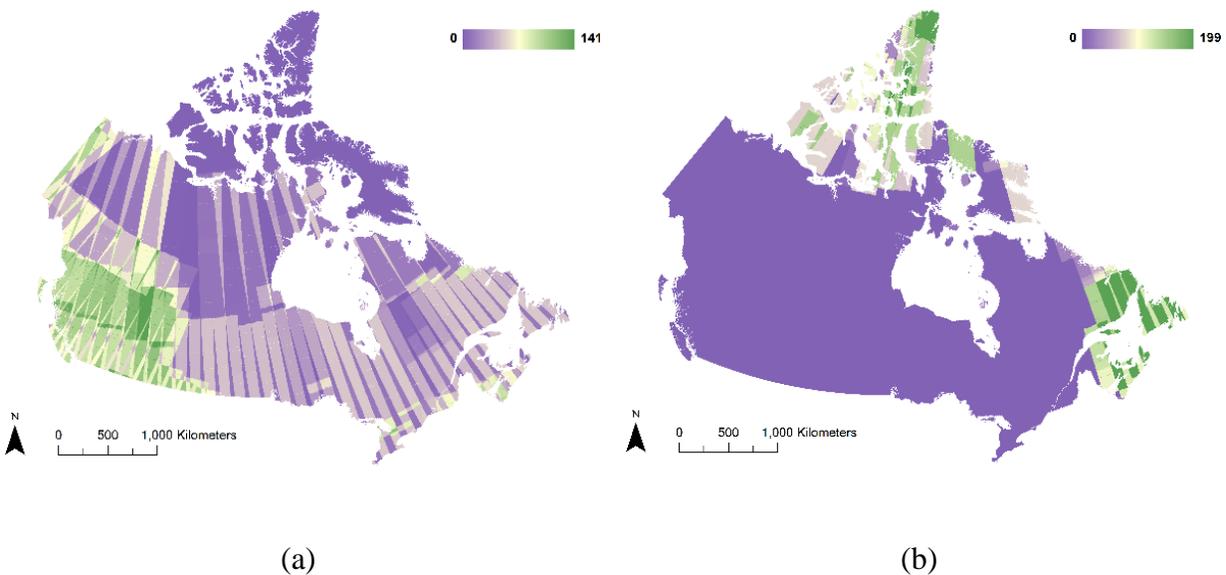


Figure 3. The total number of (a) Sentinel-1 in VV/VH mode and (b) Sentinel-1 HH/HV in mode observation during the summers of 2017-2019 in Canada. The color bar represents the number of collected images.

214

215 It should be noted that different pre-processing steps, including noise removal, radiometric
 216 calibration, and terrain correction, were already applied to the Sentinel-1 GRD data available in
 217 the GEE data catalog. To reduce the speckle noise from Sentinel-1 data, an adaptive sigma Lee
 218 filter with a pixel size of 7x7 was then applied. Next, SAR backscatter values and other derivatives
 219 of these values were extracted and incorporated into the classification scheme. Table 3 presents
 220 extracted features from Sentinel-1 and Sentinel-2 imagery for wetland classification.

221 **Table 3.** Features extracted from Sentinel-1 and Sentinel-2 imagery in this study.

Sentinel-1 (VV-VH)	Sentinel-1 (HH-HV)	Sentinel-2
σ_{VV}^0	σ_{HH}^0	Blue: B_2
σ_{VH}^0	σ_{HV}^0	Green: B_3
$ S_{VV} ^2$	$ S_{HH} ^2$	Red: B_4
$ S_{VH} ^2$	$ S_{HV} ^2$	
$ S_{VV} ^2 + S_{VH} ^2$	$ S_{HH} ^2 + S_{HV} ^2$	NIR: B_8
		$NDVI = \frac{B_8 - B_4}{B_8 + B_4}$
		$GCVI = \frac{B_8}{B_3} - 1$

222
 223 Among the extracted features from a dual-pol SAR data, σ_{HH}^0 is the most useful and frequently
 224 used for wetland mapping (Brisco et al., 2013; Mahdianpari et al., 2017; White et al., 2017;
 225 Mohammadimanesh et al. 2018c). This is because σ_{HH}^0 values are effective for characterizing the
 226 flooding status of wetland vegetation, and it is the most favorable SAR-based derivative for
 227 distinguishing flooded vegetation from herbaceous wetlands (Mohammadimanesh et al., 2018a).
 228 In cases of sparse canopy closure, σ_{VV}^0 values can also be appropriate for discriminating herbaceous
 229 wetland classes. The dominant backscattered signal from wetland' vegetation canopies is volume
 230 scattering, which is better represented by σ_{HV}^0 . Accordingly, all extracted SAR features in this
 231 study were stacked to generate a seasonal Sentinel-1 data composite using the GEE's array-based
 232 computational approach, and then, the images from multiple years (2017–2019) were combined.

233
 234 We obtained Sentinel-2A and Sentinel-2B Level-1C top of atmosphere images acquired on a tri-
 235 monthly period, from June to August. This is because generating a 10-m cloud-free Sentinel-2

236 composite for Canada over a shorter time was challenging. This period is also an optimum time
237 for wetland mapping in Canada due to the high value of wetland phenological information
238 (reflected in the range of spectral signatures for different classes), and the availability of more
239 cloud-free Sentinel-2 imagery at this time. A total of 72,046 Sentinel-2 images (with cloud-cover
240 less than 20%) from the summers of 2017-2019 were queried from the GEE data catalog. It should
241 be noted that in this study, we only used the four multispectral bands with 10m resolution to
242 produce a high-resolution (10m) wetland inventory map. Compared to our previous study, we
243 added an optical feature, Green Chlorophyll Vegetation Index (GCVI), to our analysis to
244 investigate the capability of different vegetation indices extracted from Sentinel-2 imagery. Other
245 pre-processing steps to prepare multi-spectral features for classification were explained in detail
246 in our previous work (Mahdianpari et al., 2020). Figure 4 demonstrates the spatial distribution of
247 all available Sentinel-2 observations.

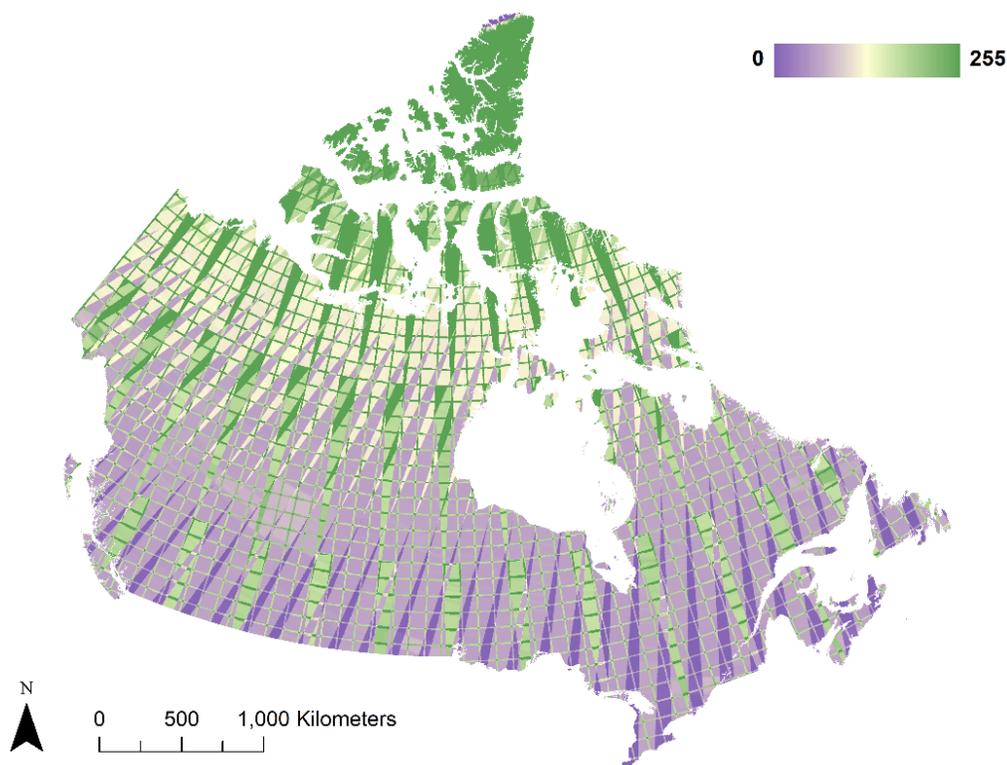


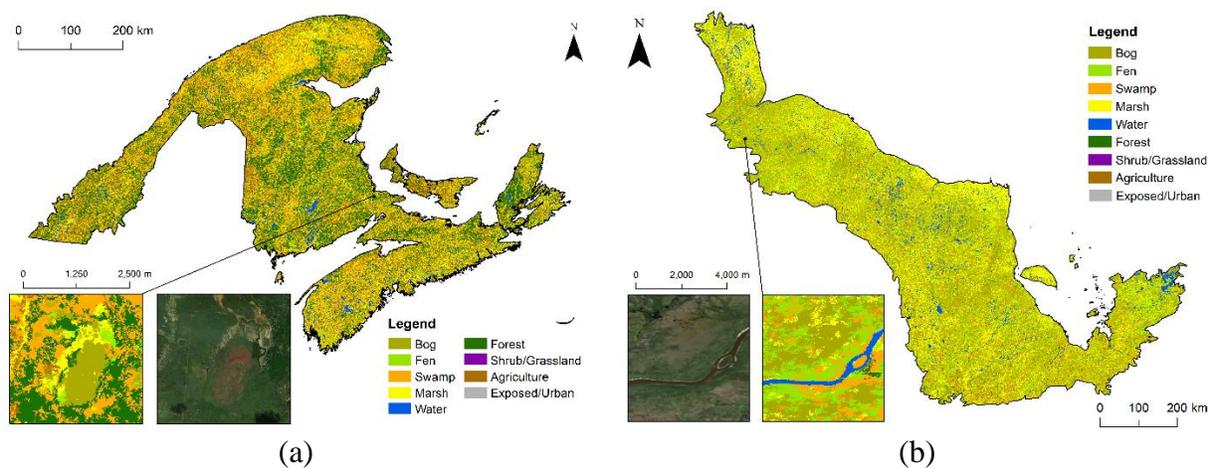
Figure 4: The spatial distribution of all available Sentinel-2 observations during the summers of 2017-2019 in Canada. The color bar represents the number of collected images.

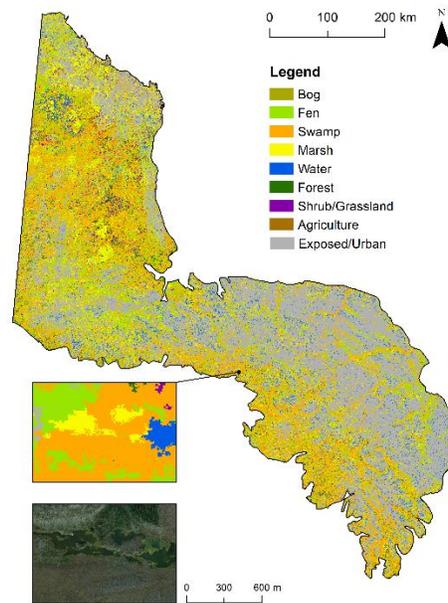
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249 In this study, an object-based classification scheme consisting of a simple non-iterative clustering
250 method, and the Random Forest algorithms were used. This classification framework is similar to
251 our previous work (Mahdianpari et al., 2020); however, we applied the classification models
252 within each ecozone rather than each province. This is because there is more commonality between
253 wetland vegetation classes, in terms of climate, landform, human activities, wildlife, soil, and
254 vegetation, within an ecozone, compared to within each provincial borders (Statistics Canada,
255 2018). In the first generation of the Canadian wetland inventory map, there was a lack of training
256 data in some ecozones, making this study impossible at that time, as training data is a major
257 bottleneck in the machine learning algorithms. The processing time for training RF models in
258 different ecozones is presented in Figure 1.

259 3. Results and discussion

260 Three examples of classified wetland ecozone maps, located in eastern, central, and western
261 Canada are presented in this section. Figure 5 demonstrates the wetland inventory map of the AM,
262 HP, and TC.





(c)

Figure 5: Classified maps of the (a) Atlantic Maritime, (b) Hudson Plains, and (c) Taiga Cordillera ecozones. [SH2]

263 Figure 5(a) shows the results of the AM classification. The most common wetlands in this area are
 264 swamp and marsh, followed by the peatlands (bog and fen). The spatial extent of wetlands, and
 265 dominance of the swamp class here, is consistent with a previous assessment of this ecozone,
 266 which states that treed wetlands are the most common type of wetland in the AM (ESTR
 267 Secretariat, 2014). However, our results likely over-estimate the extent of swamp wetlands, due in
 268 part to the limited number of training data and the difficulty in separating swamp wetlands from
 269 treed uplands (Jahncke et al., 2018). Peatlands tend to be limited to the south-east and centre of
 270 the ecozone. The most common non-wetland land cover in the AM is forest. Human-related land
 271 cover is mostly present along some of the edges of the ecozone.

272 Figure 5(b) illustrates the results for the HP, which by far, has the broadest wetland coverage
 273 relative to the results of all other ecozones. This is also in line with previous assessments of HP
 274 and reflects its reputation as the largest wetland complex in Canada, and the third-largest wetland

275 complex in the world (Abraham and McKinnon, 2011). The most dominant wetland types here
276 are bog and marsh, followed by fen, while the least dominant is the swamp. Most of the marsh is
277 located along the coast to the north and north-west. This ecozone is known to have extensive
278 coastal marshes, including tidal flats and salt marshes in this area (Abraham and McKinnon, 2011).
279 Bog and fen wetlands are also known to commonly occur in this ecozone and make up a large
280 portion of the wetland complex. Here, bog and fen occur across much of the ecozone, though they
281 are mostly concentrated through the centre. Non-wetland land cover types are mostly absent.

282 Figure 5(c) demonstrates the results for the TC, wherein the most common wetland is the swamp,
283 followed by the fen. Bog and marsh are much less common. It appears that there is likely an
284 overestimation of wetland cover in this area if we consider previous descriptions of TC, which
285 note the limited coverage of wetlands in this area (Ecosystem Classification Group, 2010). The
286 over-estimation of wetlands, particularly swamp, is likely a result of misclassification of the forests
287 and shrubby tundra in this region. Additionally, as discussed in section 2.2, there was no wetland
288 training data available in this ecozone, and as a result, it was classified in tandem with the Boc
289 ecozone. This lack of training data is reflected by the overall accuracy for this ecozone, which is
290 the lowest (along with the Boc) overall accuracy of all ecozones (see Table 4.). [SH3]The most
291 common upland classes are exposed areas, capturing the mountains along the north.

292 Table 4 shows the overall accuracy, Kappa, producer's, and user's accuracies for all ecozones. The
293 ecozone with the highest overall accuracy is the Prairies, located mainly within southern
294 Saskatchewan. Note that there was no bog data available within the Prairies ecozone, and most of
295 this area is dominated by non-wetland agricultural land (Ahern et al., 2013). As previously
296 mentioned, the ecozones with the lowest accuracies are the Boreal and Taiga Cordillera, at 76%
297 accuracies. The reasoning for this is discussed in more detail in section 2.2. However, to

298 summarize, the overall accuracy is likely a result of the lack of training data available for the Taiga
 299 Cordillera and the subsequent need to classify both the Taiga Cordillera and the Boreal Cordillera
 300 (an adjacent ecozone) at the same time, using the dataset only present within the Boreal Cordillera.
 301 Note that outside of the Taiga and Boreal Cordillera, all other ecozones were relatively well
 302 classified, with the overall accuracies higher than 80%, a majority of which (eight ecozones) are
 303 above 85%.

304 **Table 4.** Accuracy assessment indices determined for each ecozone.

Ecozone	Bog		Fen		Swamp		Marsh		Water		Upland		<i>OA</i>	<i>Kappa</i>
	<i>UA</i>	<i>PA</i>	<i>UA</i>	<i>PA</i>	<i>UA</i>	<i>PA</i>	<i>UA</i>	<i>PA</i>	<i>UA</i>	<i>PA</i>	<i>UA</i>	<i>PA</i>		
AM	0.85	0.90	0.88	0.84	0.87	0.85	0.88	0.86	0.93	0.93	0.90	0.83	0.88	0.87
Boc/TC	0.55	0.54	0.71	0.72	0.73	0.71	0.65	0.66	0.93	0.93	0.75	0.68	0.76	0.73
BP	0.94	0.75	0.80	0.90	0.86	0.84	0.78	0.84	0.94	0.94	0.89	0.84	0.87	0.86
BSE	0.83	0.92	0.81	0.70	0.83	0.81	0.86	0.76	0.93	0.94	0.87	0.83	0.86	0.84
SW	0.84	0.90	0.87	0.87	0.88	0.87	0.89	0.83	0.93	0.91	0.90	0.80	0.87	0.86
HP	0.85	0.88	0.87	0.86	0.88	0.86	0.90	0.91	0.94	0.94	0.91	0.76	0.88	0.87
MP	0.86	0.91	0.86	0.86	0.87	0.86	0.89	0.80	0.92	0.94	0.89	0.85	0.88	0.87
MC	na	na	0.94	0.65	0.77	0.63	0.77	0.63	0.94	0.94	0.79	0.77	0.85	0.83
NE	na	na	0.69	0.77	0.75	0.80	0.82	0.84	0.94	0.94	0.77	0.83	0.89	0.87
PM	0.73	0.82	0.91	0.90	0.85	0.59	0.83	0.80	0.93	0.94	0.71	0.74	0.84	0.82
Pr	na	na	0.91	0.90	0.85	0.87	0.89	0.88	0.93	0.94	0.91	0.90	0.91	0.90
TP	0.81	0.78	0.76	0.75	0.71	0.55	0.68	0.78	0.94	0.94	0.81	0.75	0.82	0.79
TS	0.74	0.72	0.62	0.64	0.54	0.39	0.70	0.66	0.94	0.94	0.76	0.76	0.84	0.79

305
 306 Figure 6 illustrates the second generation of the Canada-wide wetland inventory map at a spatial
 307 resolution of 10m using the object-based RF classification.
 308

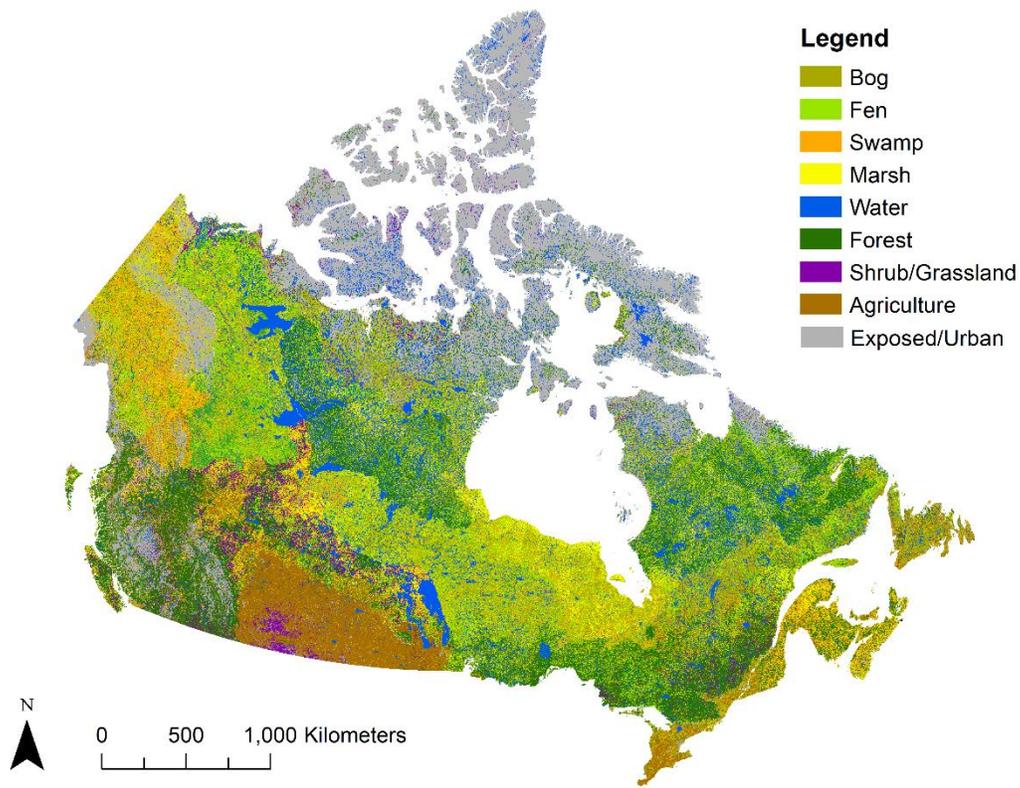


Figure 6: The second generation of Canada-wide wetland inventory map.

309
 310 According to our results, peatlands (bog and fen) are the most common wetland class in Canada,
 311 which is reflective of Canada's reputation of having extensive peatland wetlands (Mahdianpari et
 312 al., 2020). The dominance of peatlands is mostly the result of Canada's general climate, which
 313 facilitates the build-up of peat (higher precipitation than evaporation). Peatlands appear to be
 314 distributed mainly across the centre portion of Canada, from Newfoundland and Labrador to the
 315 Yukon. The ecozones that contain the highest amount of peatland include the BS, HP, MP, TP,
 316 and TS, which have been reported previously as being the major peatland-containing ecozones in
 317 Canada (Webster et al., 2018). Peatlands occur less frequently in southern Canada, where forest
 318 and anthropogenic land cover seem to dominate. Marsh wetlands are the least common of all
 319 wetland classes, with the most significant coverage by-far occurring in the HP ecozone, where

320 there are known expansive coastal marshes and tidal flats (Abraham and McKinnon, 2011). The
321 ecozones with the least marsh are in the MP and Pr ecozones, of which the landscapes have been
322 highly modified as a result of human activity, in particular, agriculture.

323 Swamp wetlands are also estimated as being a typical wetland; however, this must be interpreted
324 in relation to the known difficulty related to remotely-classifying swamp wetlands and
325 differentiating this class from the upland forest (Jahncke et al., 2018). Here, swamp appears to be
326 over-classified versus the other wetland types. However, results may be improved by increasing
327 upland forest training data, using higher resolution imagery as well as L-band for better swamp
328 forest separation, or incorporating high-resolution topographic information. However, this is not
329 always a simple solution at such large scales. Additionally, many of the swamp wetlands occur
330 along streams and rivers, and as a result, the training data polygons for these wetlands are not
331 always optimally shaped (long and thin) for use at medium spatial resolutions. Compared to the
332 first generation results (Mahdianpari et al., 2020), swamp appears to be much more common. This
333 increase may be attributed to a general increase in available wetland training data versus the first
334 generation, particularly in the Maritime Provinces. The difficulties in mapping treed wetlands,
335 such as swamp, using remote sensing has been discussed in similar studies (Jahncke et al., 2018),
336 and is of even greater difficulty when using 10m resolution imagery, or when topographical data
337 cannot be applied as is often the case with large-scale studies such as this. Notably, ecozones with
338 the greatest swamp coverage include the Boc and TC (Figure 5(c)), which, as discussed previously,
339 were the ecozones with the lowest training data and overall accuracy (Table 4).

340 One of the significant advantages of the RF classifier is its capability to determine the importance
341 of input features (i.e., variable ranking). This is beneficial when a large number of input features
342 are incorporated into the classification scheme. The RF variable ranking has been recently added

343 to GEE as an output of the random forest classifier. Figure 7 demonstrates the most important
 344 features, by ecozones.

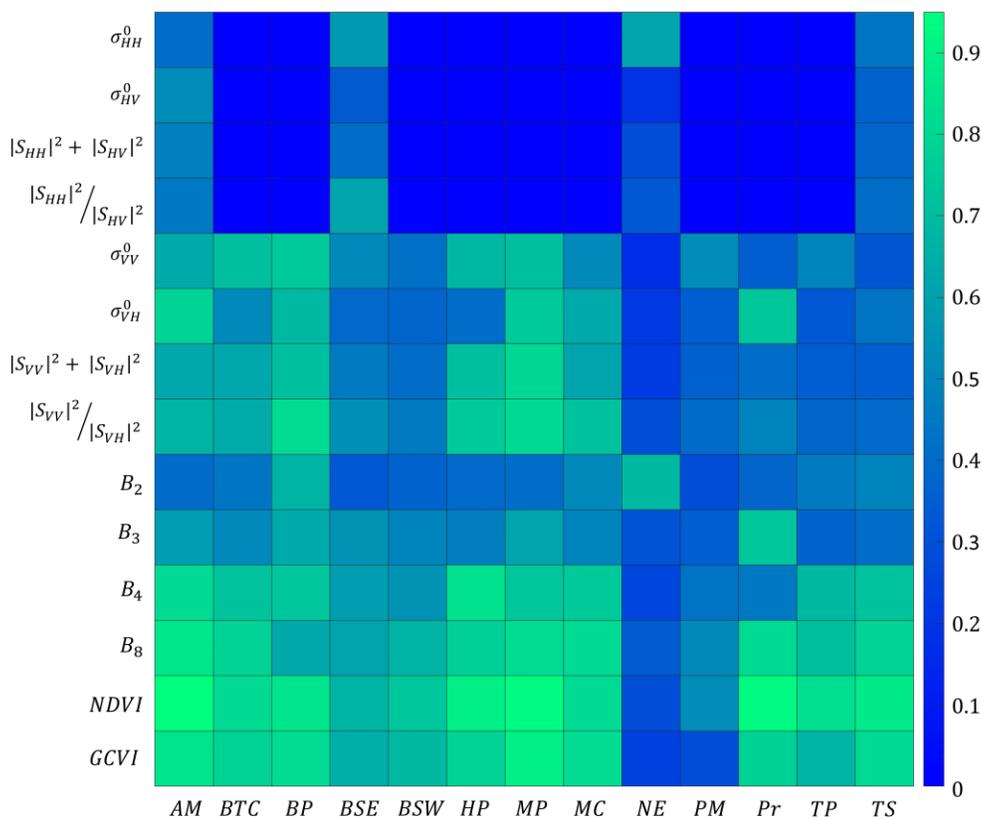


Figure 7. Normalized variable importance returned by random forest models trained on each ecozone.

345 Overall, the extracted features from optical data are more helpful for achieving higher accuracies,
 346 compared to SAR features. NDVI is the most important feature in many ecozones, particularly in
 347 ecozones with dominant agricultural activities (e.g., AM and Pr). GCVI and B_8 (Near-infrared) are
 348 also important features in several ecozones. This is expected, as forests, wetlands, and agricultural
 349 fields are dominant land cover classes throughout most of Canada's ecozones. Although B_2 is the
 350 least important optical features in most ecozones, it shows greater importance in the NE ecozone,
 351 given the presence of several small and big water bodies across this ecozone. Notably, there was
 352 a lack of dual-polarized HH-HV data in most of Canada ecozones. These features are illustrated
 353

354 with dark blue in Figure 7 in those regions. Similar to NDVI, albeit with a lower rank,
355 $\frac{|S_{VV}|^2}{|S_{VH}|^2}$ was identified as an important feature for ecozones with dominant agricultural fields (e.g.,
356 AM). This is expected, as σ_{VV}^0 observations are appropriate for discriminating herbaceous wetland
357 classes, and dominant scattering mechanisms of vegetation are volume scattering, and they have
358 the strongest responses in the cross-polarized signal (σ_{VH}^0). Span or total power, extracted from
359 dual-polarized VV-VH data, and σ_{VH}^0 are also among the useful SAR features in many ecozones.

360
361 It is often very challenging in the study like this to source a large amount of quality data from such
362 a wide variety of organizations, collaborators, institutions, and more. The present study would be
363 impossible without this data. In this study, we have managed to produce a Canada-wide wetland
364 map with very high overall accuracies. It is important to note, however, that in the case of collected
365 data such as this^[SH4], there will naturally be differences in the methods which were used to collect
366 and produce the data, the purposes for which the data was collected (many not for originally
367 produced for application in imagery classification), the years these data were collected and so on.
368 These issues are entirely expected in studies such as these. ^[SH5]Referring to section 2.2, there are
369 large differences in the amount and characteristics of data available across and within individual
370 ecozones. For example, some datasets may have more spectrally homogenous polygons than
371 others, depending on their original purpose. Additionally, the distribution of the datasets does not
372 always adequately represent the entirety of the ecozone area. All of this will have impacts on the
373 quality of the final classifications and must be considered when interpreting the results. While
374 effort was made to standardize across datasets, such as removing inappropriately sized polygons,
375 and removing any obviously out-dated polygons, much more dedicated work is needed to modify

376 and make these datasets as cohesive as possible, which was beyond the time and resources
377 available to this study, and is an on-going process.

378 Nevertheless, these datasets may act as a substantial jumping-off point for the development of a
379 Canada-wide wetland dataset suitable for applications in remote sensing. The significant effort
380 would need to be dedicated to carefully examine all available wetland data, modifying their
381 boundaries to produce more homogenous polygons, removing out-dated or inaccurate polygons,
382 and perhaps further dividing the bog, fen, swamp, and marsh polygons into sub-classes based on
383 broad vegetation characteristics (treed fen, shrub swamp, emergent marsh etc.), which would also
384 contribute to improving the homogeneity of the polygons. [SH6] This, however, is made more difficult
385 given the transient nature of wetland boundaries over the years, seasons, and even days.
386 Incorporation of some hydrological and topographical data may improve the overall classification
387 as well, particularly that of the swamp. Additionally, greater amounts of non-wetland land cover
388 would contribute to a better overall-quality remote-sensing centered wetland dataset.

389 In addition to reference data collection, it is recommended to evaluate land cover change at local,
390 regional-, or national-scales on a periodic basis, given the inherently dynamic nature of wetlands.
391 Change detection based on multi-temporal satellite imagery provides a unique opportunity to
392 monitor these changes in a cost- and time-efficient manner.

393 **4. Conclusions**

394 Wetland mapping and monitoring, especially at large scales, is challenging due to the
395 inaccessibility and diversity of wetlands, fuzziness of wetland's boundaries, as well as the cost and
396 time requirement for field data collection. Nevertheless, recent advances in remote sensing tools,
397 such as the availability of high-resolution open-access satellite imagery as well as powerful cloud

398 computing resources, alleviate these issues to the feasible extent, offering unprecedented
399 opportunities for monitoring these important natural resources using cost and time-efficient
400 methods. By leveraging the state-of-the-art remote sensing techniques, this study produced the
401 second generation of 10 m wetland inventory map of Canada using the RF classifier and data
402 collected from dual-polarimetry Sentinel-1 SAR and multi-spectral Sentinel-2 optical Earth
403 observations on the GEE cloud computing platform.

404 Compared to the first generation of this product, RF models were trained for each ecozone rather
405 than each province or territory, which increased wetland classification accuracy. This
406 improvement is a result of more commonality between wetland vegetation classes within an
407 ecozone compared to the provincial administration borders. Furthermore, significant effort has
408 been devoted to the data collection to prepare structured, cleaned, and consistent training data for
409 each ecozone, which included data acquisition, labeling, and improvement of existing data.
410 Because a data gap was identified in the Northern ecozones, high-resolution optical data from
411 Worldview-2 and Pleiades were used to delineate wetland training data in those regions. Using
412 this well distributed training data, the whole country was mapped with an overall accuracy
413 approaching 86%, representing an improvement of 7% compared to the first generation. Accuracy
414 varied from 76% to 91% in different ecozones, depending on available resources. Overall, the
415 results of the RF variable ranking demonstrate the greater importance of the optical features
416 compared to the SAR features in all ecozones. NDVI is found the most important optical feature,
417 followed by GCVI and NIR band. Among the SAR features, $\frac{|S_{VV}|^2}{|S_{VH}|^2}$ and σ_{VH}^0 illustrate the greater
418 contribution to the overall accuracy relative to others. Nevertheless, there was a lack of dual-
419 polarized HH-HV data in many ecozones. Thus, these results can not compare the capability of
420 extracted features from HH-HV and VV-VH data with each other.

421 Future works can investigate the effect of incorporating additional high-quality satellite imagery
422 collected by advanced SAR missions, such as L-band ALOS-2, L- and S- bands NASA-ISRO
423 Synthetic Aperture Radar (NISAR), or Hybrid Compact Polarimetry (HCP) data from
424 RADARSAT Constellation Mission (RCM) satellites. It is expected that adding these valuable
425 data will improve the classification accuracy considerably.

426 **Acknowledgment**

427 **References**

- 428 Abraham, K.F., McKinnon, L.M., 2011. Hudson Plains Ecozone+ evidence for key findings summary.
429 Canadian Biodiversity: Ecosystem Status and Trends 2010, Evidence for Key Findings Summary
430 Report No. 2. Canadian Councils of Resource Ministers. Ottawa, ON 98.
- 431 Agriculture and Agri-food Canada, 2018. ISO 19131 Annual Crop Inventory – Data Product Specifications.
432 Agriculture and Agri-food Canada 27.
- 433 Ahern, F.J., Frisk, J., Latifovic, R., Pouliot, D., 2013. Monitoring ecosystems remotely: a selection of trends
434 measured from satellite observations of Canada. Canadian Councils of Resource Ministers.
- 435 Alberta Environment and Sustainable Resource Development, 2015. Alberta wetland classification system.
- 436 Aschbacher, J., Milagro-Pérez, M.P., 2012. The European Earth monitoring (GMES) program: Status and
437 perspectives. *Remote Sensing of Environment* 120, 3–8.
- 438 Brisco, B., Li, K., Tedford, B., Charbonneau, F., Yun, S., Murnaghan, K., 2013. Compact polarimetry
439 assessment for rice and wetland mapping. *International journal of remote sensing* 34, 1949–1964.
- 440 Chen, B., Xiao, X., Li, X., Pan, L., Doughty, R., Ma, J., Dong, J., Qin, Y., Zhao, B., Wu, Z., 2017. A
441 mangrove forest map of China in 2015: analysis of time series Landsat 7/8 and Sentinel-1A imagery
442 in Google Earth Engine cloud computing platform. *ISPRS Journal of Photogrammetry and Remote
443 Sensing* 131, 104–120.
- 444 Chen, H., Lu, X., Wang, G., 2010. Wetland definitions: Creation, evolution and application. *Wetland
445 Science* 8, 299–304.
- 446 DeLancey, E.R., Simms, J.F., Mahdianpari, M., Brisco, B., Mahoney, C., Kariyeva, J., 2020. Comparing
447 Deep Learning and Shallow Learning for Large-Scale Wetland Classification in Alberta, Canada.
448 *Remote Sensing* 12, 2.
- 449 Dingle Robertson, L., King, D.J., Davies, C., 2015. Object-based image analysis of optical and radar
450 variables for wetland evaluation. *International Journal of Remote Sensing* 36, 5811–5841.
451 <https://doi.org/10.1080/01431161.2015.1109727>
- 452 Ducks Unlimited Canada, 2014. Boreal wetland classes in the Boreal Plains Ecozone of Canada: field guide.
453 Ducks Unlimited Canada, Stonewall, Man.
- 454 Ecosystem Classification Group, 2010. Ecological Regions of the Northwest Territories – Cordillera.
455 Department of Environment and Natural Resources, Government of the Northwest Territories,
456 Yellowknife, NT, Canada 245.
- 457 Environment and Climate Change Canada, 2016. Canadian Environmental Sustainability Indicators: Extent
458 of Canada’s Wetlands. Consulted on April 4, 2020, 12.
- 459 Erwin, K.L., 2009. Wetlands and global climate change: the role of wetland restoration in a changing world.
460 *Wetlands Ecol Manage* 17, 71–84. <https://doi.org/10.1007/s11273-008-9119-1>

461 ESTR Secretariat, 2014. Atlantic Maritime Ecozone+ evidence for key findings summary. Canadian
462 Biodiversity: Ecosystem Status and Trends 2010, Evidence for Key Findings Summary. Report No.
463 3. Canadian Councils of Resource Ministers. Ottawa, ON 100.

464 Federal, Provincial, and Territorial Governments of Canada, 2010. Canadian biodiversity: ecosystem status
465 and trends 2010. Canadian Councils of Resource Ministers. Ottawa, ON. 142.

466 Fournier, R.A., Grenier, M., Lavoie, A., Hélie, R., 2007. Towards a strategy to implement the Canadian
467 Wetland Inventory using satellite remote sensing. *Canadian Journal of Remote Sensing* 33, 16.

468 Gerbeaux, P., Finlayson, C.M., van Dam, A.A., 2016. Wetland Classification: Overview, in: Finlayson, C.
469 Max, Everard, M., Irvine, K., McInnes, R.J., Middleton, B.A., van Dam, Anne A., Davidson, N.C.
470 (Eds.), *The Wetland Book: I: Structure and Function, Management and Methods*. Springer
471 Netherlands, Dordrecht, pp. 1–8. https://doi.org/10.1007/978-94-007-6172-8_329-1

472 Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., Moore, R., 2017a. Google Earth Engine:
473 Planetary-scale geospatial analysis for everyone. *Remote Sensing of Environment* 202, 18–27.

474 Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., Moore, R., 2017b. Google Earth Engine:
475 Planetary-scale geospatial analysis for everyone. *Remote Sensing of Environment* 202, 18–27.
476 <https://doi.org/10.1016/j.rse.2017.06.031>

477 Hu, S., Niu, Z., Chen, Y., 2017. Global Wetland Datasets: a Review. *Wetlands* 37, 807–817.
478 <https://doi.org/10.1007/s13157-017-0927-z>

479 Jahncke, R., Leblon, B., Bush, P., LaRocque, A., 2018. Mapping wetlands in Nova Scotia with multi-beam
480 RADARSAT-2 Polarimetric SAR, optical satellite imagery, and Lidar data. *International Journal*
481 *of Applied Earth Observation and Geoinformation* 68, 139–156.
482 <https://doi.org/10.1016/j.jag.2018.01.012>

483 Liu, P., 2015. A survey of remote-sensing big data. *frontiers in Environmental Science* 3, 45.

484 Mahdianpari, M., Salehi, B., Mohammadimanesh, F., Brisco, B., Homayouni, S., Gill, E., DeLancey, E.R.,
485 Bourgeau-Chavez, L., 2020. Big Data for a Big Country: The First Generation of Canadian Wetland
486 Inventory Map at a Spatial Resolution of 10-m Using Sentinel-1 and Sentinel-2 Data on the Google
487 Earth Engine Cloud Computing Platform: Mégadonnées pour un grand pays: La première carte
488 d’inventaire des zones humides du Canada à une résolution de 10 m à l’aide des données Sentinel-
489 1 et Sentinel-2 sur la plate-forme informatique en nuage de Google Earth Engine™. *Canadian*
490 *Journal of Remote Sensing* 1–19.

491 Mahdianpari, M., Salehi, B., Mohammadimanesh, F., Homayouni, S., Gill, E., 2018. The First Wetland
492 Inventory Map of Newfoundland at a Spatial Resolution of 10 m Using Sentinel-1 and Sentinel-2
493 Data on the Google Earth Engine Cloud Computing Platform. *Remote Sensing* 11, 43.
494 <https://doi.org/10.3390/rs11010043>

495 Mahdianpari, M., Salehi, B., Mohammadimanesh, F., Motagh, M., 2017. Random forest wetland
496 classification using ALOS-2 L-band, RADARSAT-2 C-band, and TerraSAR-X imagery. *ISPRS*
497 *Journal of Photogrammetry and Remote Sensing* 130, 13–31.

498 Millard, K., Richardson, M., 2015. On the Importance of Training Data Sample Selection in Random Forest
499 Image Classification: A Case Study in Peatland Ecosystem Mapping. *Remote Sensing* 7, 8489–
500 8515. <https://doi.org/10.3390/rs70708489>

501 Mohammadimanesh, F., Salehi, B., Mahdianpari, M., Brisco, B., Motagh, M., 2018a. Wetland Water Level
502 Monitoring Using Interferometric Synthetic Aperture Radar (InSAR): A Review. *Canadian Journal*
503 *of Remote Sensing* 1–16. <https://doi.org/10.1080/07038992.2018.1477680>

504 Mohammadimanesh, F., Salehi, B., Mahdianpari, M., Motagh, M., 2018b. A New Hierarchical Object-
505 Based Classification Algorithm for Wetland Mapping in Newfoundland, Canada, in: *IGARSS 2018*
506 *- 2018 IEEE International Geoscience and Remote Sensing Symposium*. Presented at the *IGARSS*
507 *2018 - 2018 IEEE International Geoscience and Remote Sensing Symposium*, IEEE, Valencia, pp.
508 9233–9236. <https://doi.org/10.1109/IGARSS.2018.8517844>

509 Mohammadimanesh, F., Salehi, B., Mahdianpari, M., Motagh, M., Brisco, B., 2018c. An efficient feature
510 optimization for wetland mapping by synergistic use of SAR intensity, interferometry, and

511 polarimetry data. *International Journal of Applied Earth Observation and Geoinformation* 73, 450–
512 462.

513 Mui, A., He, Y., Weng, Q., 2015. An object-based approach to delineate wetlands across landscapes of
514 varied disturbance with high spatial resolution satellite imagery. *ISPRS Journal of Photogrammetry
515 and Remote Sensing* 109, 30–46.

516 National Wetlands Working Group, 1997. The Canadian wetland classification system. Wetlands Research
517 Branch, University of Waterloo, Waterloo, Ont.

518 Rezaee, M., Mahdianpari, M., Zhang, Y., Salehi, B., 2018. Deep convolutional neural network for complex
519 wetland classification using optical remote sensing imagery. *IEEE Journal of Selected Topics in
520 Applied Earth Observations and Remote Sensing* 11, 3030–3039.

521 Shelestov, A., Lavreniuk, M., Kussul, N., Novikov, A., Skakun, S., 2017. Exploring Google earth engine
522 platform for Big Data Processing: Classification of multi-temporal satellite imagery for crop
523 mapping. *Frontiers in Earth Science* 5, 17.

524 Sidhu, N., Pebesma, E., Câmara, G., 2018. Using Google Earth Engine to detect land cover change:
525 Singapore as a use case. *European Journal of Remote Sensing* 51, 486–500.
526 <https://doi.org/10.1080/22797254.2018.1451782>

527 Smith, C.A.S., Meikle, J.C., Roots, C.F., 2004. Ecoregions of the Yukon Territory: Biophysical properties
528 of Yukon landscapes. Agriculture and Agri-Food Canada, PARC Technical Bulletin No. 04-01,
529 Summerland, British Columbia 313.

530 Statistics Canada, 2018. Ecological Land Classification, 2017. Statistics Canada, Ottawa.

531 van der Kamp, G., Hayashi, M., Bedard-Haughn, A., Pennock, D., 2016. Prairie Pothole Wetlands –
532 Suggestions for Practical and Objective Definitions and Terminology. *Wetlands* 36, 229–235.
533 <https://doi.org/10.1007/s13157-016-0809-9>

534 Webster, K.L., Bhatti, J.S., Thompson, D.K., Nelson, S.A., Shaw, C.H., Bona, K.A., Hayne, S.L., Kurz,
535 W.A., 2018. Spatially-integrated estimates of net ecosystem exchange and methane fluxes from
536 Canadian peatlands. *Carbon Balance and Management* 13, 21.

537 White, L., Millard, K., Banks, S., Richardson, M., Pasher, J., Duffe, J., 2017. Moving to the RADARSAT
538 constellation mission: Comparing synthesized compact polarimetry and dual polarimetry data with
539 fully polarimetric RADARSAT-2 data for image classification of peatlands. *Remote Sensing* 9,
540 573.

541 Zhou, B., Okin, G.S., Zhang, J., 2020. Leveraging Google Earth Engine (GEE) and machine learning
542 algorithms to incorporate in situ measurement from different times for rangelands monitoring.
543 *Remote Sensing of Environment* 236, 111521. <https://doi.org/10.1016/j.rse.2019.111521>
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