



Denitrifying metabolism of the methylotrophic marine bacterium *Methylophaga nitratireducenticrescens* strain JAM1

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ABSTRACT

Background. *Methylophaga nitratireducenticrescens* strain JAM1 is a methylotrophic, marine bacterium that was isolated from a denitrification reactor treating a closed-circuit seawater aquarium. It can sustain growth under anoxic conditions by reducing nitrate (NO_3^-) to nitrite (NO_2^-). These physiological traits are attributed to gene clusters that encode two dissimilatory nitrate reductases (Nar). Strain JAM1 also contains gene clusters encoding two nitric oxide (NO) reductases and one nitrous oxide (N_2O) reductase, suggesting that NO and N_2O can be reduced by strain JAM1. Here we characterized further the denitrifying activities of *M. nitratireducenticrescens* JAM1.

Methods. Series of oxic and anoxic cultures of strain JAM1 were performed with N_2O , NO_3^- or sodium nitroprusside, and growth and N_2O , NO_3^- , NO_2^- and N_2 concentrations were measured. Ammonium (NH_4^+)-free cultures were also tested to assess the dynamics of N_2O , NO_3^- and NO_2^- . Isotopic labeling of N_2O was performed in $^{15}\text{NH}_4^+$ -amended cultures. Cultures with the JAM1 $\Delta narG1narG2$ double mutant were performed to assess the involvement of the Nar systems on N_2O production. Finally, RT-qPCR was used to measure the gene expression levels of the denitrification genes cytochrome *bc*-type nitric oxide reductase (*cnorB1* and *cnorB2*) and nitrous oxide reductase (*nosZ*), and also *nnrS* and *norR* that encode NO-sensitive regulators.

Results. Strain JAM1 can reduce NO to N_2O and N_2O to N_2 and can sustain growth under anoxic conditions by reducing N_2O as the sole electron acceptor. Although strain JAM1 lacks a gene encoding a dissimilatory NO_2^- reductase, NO_3^- -amended cultures produce N_2O , representing up to 6% of the N-input. NO_2^- was shown to be the key intermediate of this production process. Upregulation in the expression of *cnorB1*, *cnorB2*, *nnrS* and *norR* during the growth and the N_2O accumulation phases suggests NO production in strain JAM1 cultures.

Discussion. By showing that all the three denitrification reductases are active, this demonstrates that *M. nitratireducenticrescens* JAM1 is one of many bacteria species that maintain genes associated primarily with denitrification, but not necessarily related to the maintenance of the entire pathway. The reason to maintain such an incomplete pathway could be related to the specific role of strain JAM1 in the denitrifying biofilm of the denitrification reactor from which it originates. The production of N_2O in strain JAM1 did not involve Nar, contrary to what was demonstrated in *Escherichia coli*. *M. nitratireducenticrescens* JAM1 is the only reported *Methylophaga* species that has the

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capacity to grow under anoxic conditions by using NO_3^- and N_2O as sole electron acceptors for its growth. It is also one of a few marine methylotrophs that is studied at the physiological and genetic levels in relation to its capacity to perform denitrifying activities.

Subjects Environmental Sciences, Microbiology, Molecular Biology, Biogeochemistry

Keywords Denitrification, Marine bacterium, *Methylophaga*, Nitrous oxide, Nitrate

INTRODUCTION

The complete denitrification pathway describes the successive reduction of nitrate (NO_3^-) to nitrite (NO_2^-), nitric oxide (NO), nitrous oxide (N_2O), and nitrogen (N_2) (Van Spanning, Delgado & Richardson, 2005). This process is used by bacteria for respiration in environments with low oxygen concentrations and with NO_3^- as an electron acceptor. The process is driven by metalloenzymes NO_3^- reductase, NO_2^- reductase, NO reductase, and N_2O reductase (Einsle & Kroneck, 2004). As a facultative trait, denitrification occurs frequently across environments and is performed by bacteria of diverse origins (Zumft, 1997). However, numerous bacterial strains have been isolated with incomplete denitrification pathway, meaning that at least one reductase-encoding gene cluster is missing. As proposed by Zumft (Zumft, 1997), the four steps of reduction from NO_3^- to N_2 could be seen as a modular assemblage of four partly independent respiratory processes that respond to combinations of different external and internal signals. This could explain the vast diversity of bacteria with incomplete denitrification pathway that can sustain growth with one of the four nitrogen oxides as electron acceptor. Another purpose of the incomplete pathway is related to detoxification, as NO_2^- and NO are deleterious molecules (Kaspar, 1982; Poole, 2005; Schreiber et al., 2012; Simon & Klotz, 2013).

Methylophaga nitratireducentescens JAM1 is a marine methylotrophic gammaproteobacterium that was isolated from a naturally occurring multispecies biofilm that has developed in a methanol-fed, fluidized denitrification system that treated recirculating water of the marine aquarium in the Montreal Biodome (Auclair et al., 2010; Villeneuve et al., 2013). This biofilm is composed of at least 15 bacterial species and of numerous protozoans (Labbé et al., 2003; Laurin et al., 2008), among which *Methylophaga* spp. and *Hyphomicrobium* spp. compose more than 50% of the biofilm (Labbé et al., 2007). Along with the denitrifying bacterium *Hyphomicrobium nitrativorans* NL23, *M. nitratireducentescens* JAM1 was shown to be the representative of the *Methylophaga* population in the biofilm (Auclair et al., 2010).

M. nitratireducentescens JAM1 is considered as a nitrate respirer as it can grow under anoxic conditions through the reduction of NO_3^- to NO_2^- , which accumulates in the culture medium (Auclair et al., 2010). This trait is correlated with the presence of two gene clusters encoding dissimilatory nitrate reductases (*narGHJI*, referred as Nar1 and Nar2) in the genome of *M. nitratireducentescens* JAM1, which we showed that both contribute to NO_3^- reduction during strain JAM1 growth (Mauffrey, Martineau & Villemur, 2015). Anaerobic growth by strain JAM1 is a unique among *Methylophaga*

spp. that were described as strictly aerobic bacteria (Boden, 2012). Genome annotation revealed that strain JAM1 seems to maintain an incomplete denitrification pathway with the presence of gene clusters encoding two putative cytochrome *bc*-type complex NO reductase (cNor) (*cnor1* and *cnor2*) and one putative dissimilatory N₂O reductase, but lacks gene encoding a dissimilatory copper- (NirK) or cytochrome *cd1*-type (NirS) NO₂⁻ reductase. These gene clusters have been shown to be transcribed. However, the capacity of *M. nitratireducenticrescens* JAM1 to consume NO and N₂O has not been fully determined. In addition to these gene clusters, genes involved in the NO response such as *nnrS* and *norR* are present (Mauffrey, Martineau & Villemur, 2015) suggesting tight regulation of denitrification genes such as *cnorB* and *nosZ*. Finally, the genome has a gene cluster encoding assimilatory nitrate and NADH-dependent nitrite reductases.

In this study, we assessed further the denitrification activities of strain JAM1 in pure cultures by demonstrating the consumption of NO and N₂O by strain JAM1 in cultures amended with N₂O or sodium nitroprusside as NO provider. Through our investigation, we found that strain JAM1 cultured with NO₃⁻ generates a small amount of N₂O. We assessed whether nitrate and the nitrate reductases, nitrite and ammonium are directly involved in this N₂O production, and found that NO₂⁻ is a key intermediate of this production process. Finally, we showed that the N₂O accumulation/consumption cycle in NO₃⁻-amended cultures affects the expression of denitrification genes *cnorB* (*cnorB1* and *cnorB2*) and *nosZ*, and also *nnrS* and *norR*, which encode NO-sensitive regulators. These results suggest that NO is also generated in NO₃⁻-amended cultures.

MATERIALS AND METHODS

Bacterial growth conditions

M. nitratireducenticrescens JAM1 and the JAM1 $\Delta narG1 narG2$ double mutant were cultured in the American Type Culture Collection (ATCC, Manassas, VA, USA) *Methylophaga* medium 1403 (Villeneuve et al., 2013; Mauffrey, Martineau & Villemur, 2015). When required, NO₃⁻ (NaNO₃) or NO₂⁻ (NaNO₂) (Fisher Scientific Canada, Ottawa, ON, Canada) were added to the medium. Medium (40 or 60 mL) was dispensed into 720-mL bottles (680- or 660-mL head space) that were sealed with caps equipped with septum and which were then autoclaved. After autoclaving, the following filter-sterilized solutions were added to the bottles (40 mL volume): 120 μ L methanol (final concentration 0.3% [vol/vol]; 74.3 mM), 800 μ L solution T (per 100 mL: 0.7 g KH₂PO₄, 10 g NH₄Cl, 10 g Bis-Tris, 0.3 g ferric ammonium citrate (pH 8)), 400 μ L Wolf's mineral solution (pH 8) (ATCC), and 40 μ L vitamin B₁₂ (stock solution 0.1 mg/mL). The Wolf mineral solution is composed of (per liter) 0.5 g EDTA, 3.0 g MgSO₄·7H₂O, 0.5 g MnSO₄·H₂O, 1.0 g NaCl, 0.1 g FeSO₄·7H₂O, 0.1 g Co(NO₃)₂·6H₂O, 0.1 g CaCl₂ (anhydrous), 0.1 g ZnSO₄·7H₂O, 0.010 g CuSO₄·5H₂O, 0.010 g AlK(SO₄)₂ (anhydrous), 0.010 g H₃BO₃, 0.010 g Na₂MoO₄·2H₂O, 0.001 g Na₂SeO₃ (anhydrous), 0.010 g Na₂WO₄·2H₂O, and 0.020 g NiCl₂·6H₂O. The final concentration of ammonium (NH₄⁺) in the *Methylophaga* 1403 medium was measured as 21 mg-N vial⁻¹ (20.9 mg-N vial⁻¹ from NH₄Cl and 0.1 mg-N vial⁻¹ from ferric ammonium citrate). The amount of NO₃⁻ carried by the Wolf mineral solution (0.0038 mg-N vial⁻¹)

was deemed negligible. For the anoxic cultures, bottles were flushed with nitrogen gas (N_2 , purity >99.9%; Praxair, Mississauga, ON, Canada) or argon (purity 99.9%, Praxair) for 20 min prior to autoclaving. When necessary, N_2O (purity 99.9%, Praxair) and acetylene (10% [vol/vol] of headspace; Praxair) were injected into the headspace before autoclaving. Acetylene is an inhibitor of nitrous oxide reductase and has been extensively used in N_2O studies to observe N_2O production in cells (*Klemetsson et al., 1977*). Inoculums were made from fresh culture cultivated under oxic conditions without NO_3^- to reach an optical density (OD_{600}) of 0.025. Culture bottles were incubated at 30 °C in the dark. For oxic cultures, bottles were shaken at 150 rpm.

The capacity for strain JAM1 to reduce NO was tested with sodium nitroprusside (sodium nitroprusside hypochloride ([SNP]; purity $\geq 99.0\%$, Sigma-Aldrich, St. Louis, MO, USA) as the NO source. To avoid SNP toxicity, strain JAM1 was first cultured in *Methylophaga* 1403 medium under oxic conditions without NO_3^- for 24 h. The cells were then centrifuged (8,000 g 5 min) and dispersed into fresh medium supplemented with 2 mM, 5 mM, or no SNP. Culture medium with 5 mM SNP and no biomass was also used as a control. Cells were incubated under oxic conditions at 30 °C in the dark, and N_2O production was monitored. To investigate the potential role of NH_4^+ in N_2O production, NH_4Cl -free cultures were employed under oxic and anoxic conditions using solution T containing no NH_4Cl . Prior to inoculation, cells from start-up cultures were centrifuged and rinsed three times with saline solution to remove any residual traces of NH_4^+ .

Bacterial growth was monitored by spectrophotometry (OD_{600}). Bacterial flocs were dispersed with a Potter-Elvehjem homogenizer prior to measurement. Oxygen concentrations in the headspace were monitored in cultures under oxic conditions by gas chromatography using a temperature conductivity detector (7890B series GC Custom, SP1 option 7890-0504/0537; Agilent Technologies, Mississauga, ON, Canada). Although vials were capped in the oxic cultures, O_2 concentrations in the headspace (680 ml) did not significantly decrease ($T0\ h = 20.4 \pm 0.3\%$; $T100\ h = 19.7 \pm 0.9\%$).

^{15}N -labeling of N_2O

Strain JAM1 cultures were made with 22 mg-N vial⁻¹ $Na^{15}NO_3$ (Sigma-Aldrich) in NH_4Cl -free medium or with 22 mg-N vial⁻¹ $Na^{14}NO_3$ and 20.7 mg-N vial⁻¹ $^{15}NH_4Cl$ (Sigma-Aldrich). Both cultures were used under anoxic conditions, and 10% (vol/vol) acetylene was added to allow N_2O to accumulate. Cultures were made in triplicate. After 14 days of incubation, the headspace of each replicate was pooled, and 100 mL of the gaseous phase was sampled in Tedlar bags. N_2O -isotope measurements were performed at the Environmental Isotope Laboratory (Earth & Environmental Sciences; University of Waterloo, ON, Canada) via Trace Gas-GVI IsoPrime-Isotope Ratio Mass Spectrometry (TG-IRMS). $^{45}[N_2O]/^{44}[N_2O]$ and $^{46}[N_2O]/^{44}[N_2O]$ ratios were calculated according to the peak intensity measured for $^{46}[N_2O]$, $^{45}[N_2O]$ and $^{44}[N_2O]$. The $^{15}N/^{14}N$ isotopic ratio was derived from the previous results from Eq. (1).

$$R_s = \sum(^{15}N\ \text{vial}^{-1}) / \sum(^{14}N\ \text{vial}^{-1}) = [^{15}N45 + 2(^{15}N46)] / [2(^{14}N44) + ^{14}N45] \quad (1)$$

where R_s is the sample isotopic ratio. Calculated from the $^{45}[N_2O]/^{44}[N_2O]$ and $^{46}[N_2O]/^{44}[N_2O]$ isotopic ratios, $^{14}N45$ is the quantity of ^{14}N in $^{45}[N_2O]$, $^{15}N45$ is

the quantity of ^{15}N in $^{45}[\text{N}_2\text{O}]$, $^{14}\text{N}44$ is the quantity of ^{14}N in $^{44}[\text{N}_2\text{O}]$ and $^{15}\text{N}46$ is the quantity of ^{15}N in $^{46}[\text{N}_2\text{O}]$. We considered the isotope fractionation by denitrification enzymes as negligible in our calculations (delta values ranging from -10‰ to -40‰) (Snider, Schiff & Spoelstra, 2009).

Measurements of nitrogenous compounds

NO_3^- and NO_2^- concentrations were determined by ion chromatography using the 850 Professional IC (Metrohm, Herisau, Switzerland) with a Metrosep A Supp 5 analytical column (250 mm \times 4.0 mm).

N_2O and N_2 concentrations were determined by gas chromatography. Headspace samples (10 mL) were collected using a Pressure Lok gastight glass syringe (VICI Precision Sampling Inc., Baton Rouge, LA, USA) and were injected through the injection port of a gas chromatograph equipped with a thermal conductivity detector and electron-capture detector (7890B series GC Custom, SP1 option 7890-0504/0537; Agilent Technologies). The reproducibility of the N_2O was assessed before each set of measurements via the repeated analysis of certified N_2O standard gas with standard deviations $<5\%$. N_2O standards (500 ppmv and 250 ppmv) were created based on dilutions from the 10,000 ppmv N_2O stock standard. The 10,000 ppmv stock standard was obtained by injecting 1% pure N_2O (Praxair) into a 720 mL gastight bottle. The detection limit of the N_2O was set to <10 ppbv, corresponding to the 0.3 nmol/vial composition of our bioassays. No significant N_2O production patterns were observed through our blank experiments involving sterile media and empty glass bottles. The total quantity of N_2O in the culture bottle (aqueous phase and headspace) ($X_{\text{N}_2\text{O}}$ in $\mu\text{mole vial}^{-1}$) was calculated according to Eq. (2).

$$X_{\text{N}_2\text{O}} = [K_{\text{cpH}30\text{sw}} * A_{\text{N}_2\text{O}} * P * V_1]_{\text{aq}} + [A_{\text{N}_2\text{O}} * V_g / V_n]_{\text{gaz}} \quad (2)$$

where $A_{\text{N}_2\text{O}}$: the N_2O mixing ratio measured in the headspace ($\mu\text{mole}_{\text{N}_2\text{O}} \text{mole}^{-1}$); P : 1 atm; V_1 and V_g : volume of the aqueous (0.04 or 0.06 L vial^{-1}) and gaseous phases (0.68 or 0.66 L vial^{-1}), respectively; and V_n : molar volume (RT (gas constant): $0.08206 \text{ L atm K}^{-1} \text{ mol}^{-1} * 303 \text{ K} = 24.864 \text{ L mol}^{-1}$). $K_{\text{H}30\text{sw}}$ is the corrected Henry's constant for seawater at 30°C ($0.01809 \text{ mol L}^{-1} \text{ atm}^{-1}$) according to Weiss and Price (1980). $X_{\text{N}_2\text{O}}$ was then converted (Eq. (3)) into mg-N vial^{-1} for an easier calculation of mass balances using the other nitrogenous compounds:

$$X_{\text{N-N}_2\text{O}} = X_{\text{N}_2\text{O}} * [2\text{N}/\text{N}_2\text{O}] * [0.014 \text{ mg-N } \mu\text{mole}^{-1}]. \quad (3)$$

The reproducibility of the N_2 was assessed before each set of measurements via a repeated analysis of N_2 (purity $>99.99\%$, Praxair) diluted in a 720 mL gastight bottle (0 and 500 ppmv) flushed with argon (purity $>99.99\%$, Praxair). The total quantity of N_2 in the culture bottles was only considered for the headspace, as the quantity of dissolved N_2 in the aqueous phase was considered to be negligible in our experimental design based on Henry's constant ($0.0005 \text{ mol L}^{-1} \text{ atm}^{-1}$) and was thus calculated according to Eq. (4).

$$X_{\text{N-N}_2}(\text{mg-N } \text{vial}^{-1}) = [A_{\text{N}_2} * V_g / V_n]_{\text{gaz}} * [2\text{N}/\text{N}_2] * [0.014 \text{ mg-N } \mu\text{mole}^{-1}]. \quad (4)$$

RNA extraction

Anoxic cultures of strain JAM1 were created in an NH_4Cl -free 1403 medium supplemented with $22 \text{ mg-N vial}^{-1} \text{ NO}_3^-$. Cells were harvested at specific times, and RNA was immediately extracted using the PureLink RNA mini kit (Ambion Thermo Fisher Scientific, Burlington, ON, Canada). RNA extracts were treated twice with TurboDNase (Ambion), and RNA quality was verified by agarose gel electrophoresis. The absence of remaining DNA was checked via the end-point polymerase chain reaction (PCR) amplification of the 16S rRNA gene using RNA extracts as the template.

Gene expression

cDNAs samples were generated from the RNA using hexameric primers and the Reverse Transcription System developed by Promega (Madison, WI, USA) with $1 \mu\text{g}$ of RNA and quantified by spectrophotometry. Real-time quantitative PCR (qPCR) assays were performed using the Faststart SYBR Green Master (Roche Diagnostics, Laval, QC, Canada) according to the manufacturer's instructions. All reactions were performed in a Rotor-Gene 6000 real-time PCR thermocycler (Qiagen Inc. Toronto, ON, Canada), and each reaction contained 25 ng of cDNA and 300 nM of primers (Table 1). Genes tested included *cnorB1*, *cnorB2*, *nnrS*, *nosZ*, *norR* and *nr*, and the reference genes *dnaG*, *rpoD* and *rpoB* (Mauffrey, Martineau & Villemur, 2015) and the PCR began with an initial denaturation step of 10 min at $95 \text{ }^\circ\text{C}$ followed by 40 cycles of 10 s at $95 \text{ }^\circ\text{C}$, 15 s at $60 \text{ }^\circ\text{C}$, and 20 s at $72 \text{ }^\circ\text{C}$. To confirm the purity of the amplified products, a melting curve analysis was performed by increasing the temperature from $65 \text{ }^\circ\text{C}$ to $95 \text{ }^\circ\text{C}$ at increments of $1 \text{ }^\circ\text{C}$ per step with a pause of 5 s included between each step. All genes for each sample and standard were tested in a single run. The amplification efficiency level was tested for each set of primer pairs by qPCR using a dilution of strain JAM1 genomic DNA as the template. The amplification efficiencies for all primer pairs varied between 0.9 and 1.1 . The copy number of each gene was calculated according to standard curves using dilutions of strain JAM1 genomic DNA. The gene expression levels of the targeted genes were standardized with the three reference genes. The RNA extraction and qPCR were performed with three to four independent biological replicates. The significance of differential expression levels was tested for each phase against the pre-culture phase using One-way ANOVA tests with Tukey *post hoc* tests.

RESULTS

M. nitratireducens JAM1 grows on N_2O under anoxic conditions

Strain JAM1 was cultured under anoxic conditions with either NO_3^- in the medium or with N_2O injected in the headspace as the sole electron acceptor. Both types of culture received the same electron equivalent of NO_3^- or N_2O ($1.3 \text{ mmole vial}^{-1}$ or 18.2 and $36.4 \text{ mg-N vial}^{-1}$, respectively) according to:



Table 1 Primers used for RT-qPCR.

Primers	Target gene	Locus tag*	Sequence (5'–3')
cytochrome <i>bc</i> -type complex NO reductases			
cnorB1-510f	<i>cnorB1</i>	Q7A_0433	CCTGATCGGTTTGGCTCTC
cnorB1-635r			CCCATGATCAATTCCCAGAC
cnorB2-334f	<i>cnorB2</i>	Q7A_0487	GGCAACAAGCTATTGGAGCA
cnorB2-449r			GTGGTGGTAAAGCGACCAGA
N ₂ O reductase			
nosZ-826f	<i>nosZ</i>	Q7A_0459	GAGCGTGACTGGGTAGTCGT
nosZ-952r			GTGTCAACTCGCTCCCTTTG
NO-sensitive regulators			
nnrs-749f	<i>nnrS</i>	Q7A_1801	TGTTCCGCATTTACGCAATA
nnrs-848r			TAACCGATGTGCAAAGACCA
norR-265f	<i>norR</i>	Q7A_0435	CGGTTTGCTGCAGATAGTGA
norR-386r			CCCCAGGGCCTGTTATTTAT
Assimilatory nitrate reductase			
nr-1350f	<i>nr</i>	Q7A_2619	ATTCGGTACAGTCGGTTTGC
nr-1474r			TGTCTGGATTATTGCCACCA
Reference genes			
dnaG-774f	<i>dnaG</i>	Q7A_342	CATCCTGATCGTGGAAGGTT
dnaG-894r			GCTGCGAATCAACTGACGTA
rpob (3861F)	<i>rpob</i>	Q7A_2329	TGAGATGGAGGTTTGGGCAC
rpob (4006R)			GCATACCTGCATCCATCCGA
rpoD (10F)	<i>rpoD</i>	Q7A_343	CAGCAATCACGCGTTAAAGA
rpoD(153R)			ACCCAGGTCGCTGAACATAC

Notes.*from GenBank accession number [CP003390.3](https://www.ncbi.nlm.nih.gov/nuclot/CP003390.3).

In N₂O-amended cultures, N₂O decrease was apparent from the start and consumption continued for 48 h (Fig. 1A). The N₂O decrease paralleled strain JAM1 growth with almost complete N₂O consumption. The NO₃⁻-amended cultures showed complete NO₃⁻ consumption and equivalent NO₂⁻ accumulation after 24 h (Fig. 1B). However, slower growth than that recorded for the N₂O cultures was observed. Such growth kinetics could be related to the toxicity of NO₂⁻ that accumulated in the medium. Both types of culture reached equivalent biomass concentration (*t* test on the last 4-time points, *P* > 0.05).

***M. nitratireducenticrescens* JAM1 consumes N₂O under oxic conditions**

In a previous study, [Mauffrey, Martineau & Villemur \(2015\)](#) demonstrated that strain JAM1 can consume NO₃⁻ under oxic growth conditions with equivalent accumulation of NO₂⁻. We tested if this was also the case with N₂O. Culturing strain JAM1 under oxic conditions with N₂O (3.5 mg-N vial⁻¹) showed a complete N₂O consumption within 24 h (Fig. 2). Growth patterns illustrated in Fig. 2 were similar between oxic cultures amended with or without either N₂O or NO₃⁻. In the presence of O₂, cultures reached higher (4–5 times) biomass concentration than the anoxic cultures.

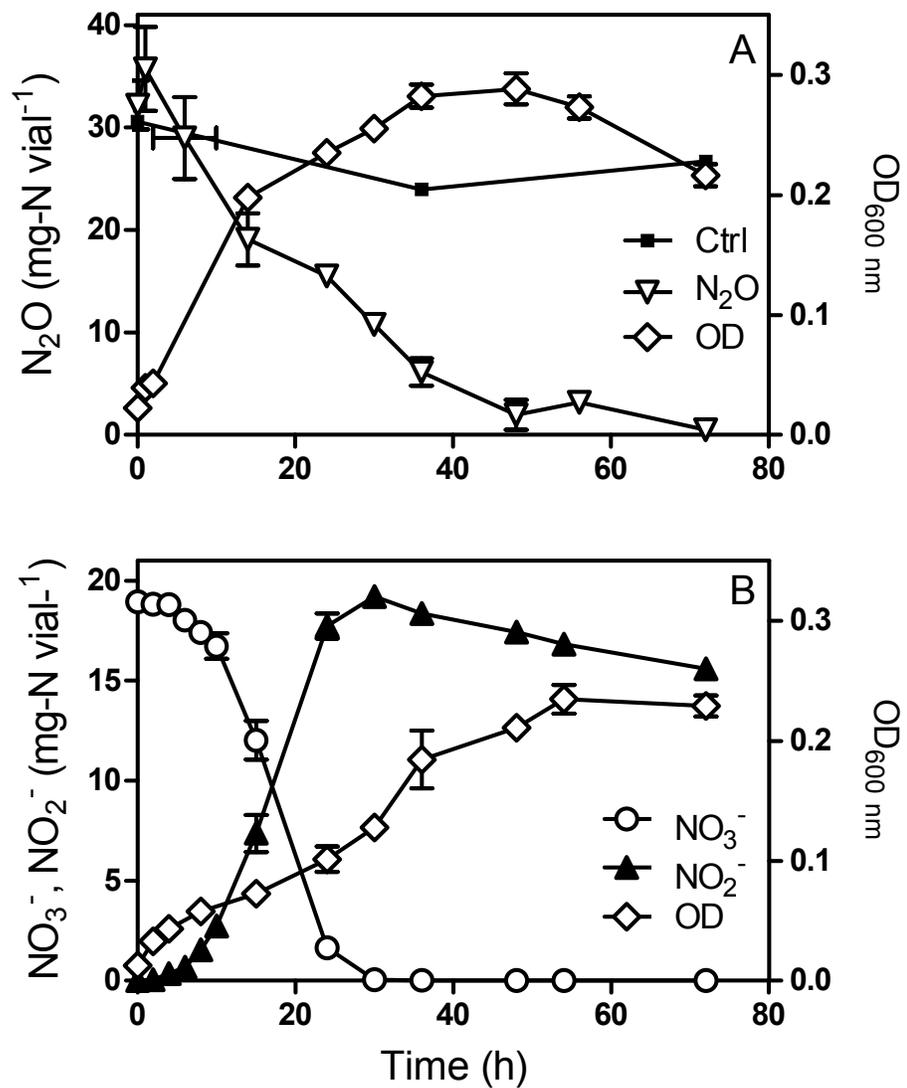


Figure 1 *Methylophaga nitritireducentescens* JAM1 growth with N₂O or NO₃⁻ as an electron acceptor. Strain JAM1 was cultured with 36.4 mg-N vial⁻¹ N₂O (A) or 18.2 mg-N vial⁻¹ NO₃⁻ (B) under anoxic conditions. N₂O, NO₃⁻ and NO₂⁻ concentrations and growth were measured over different time intervals. Control (A): N₂O injected in non-inoculated vials. To minimize oxygen contamination, sampling was performed using a glove bag inflated with nitrogen gas. Data represent mean values ± standard deviation (SD; n = 3).

Full-size [DOI: 10.7717/peerj.4098/fig-1](https://doi.org/10.7717/peerj.4098/fig-1)

N₂O production in NO₃⁻-amended cultures

During the first assays to test the capacity of strain JAM1 to reduce N₂O under anoxic conditions, cultures were performed with N₂O (3.5 mg-N vial⁻¹) but with the addition of NO₃⁻ (20 mg-N vial⁻¹) to make sure that growth would occur. Although N₂O was completely consumed within 24 h, a net production of N₂O was observed after 48 h. To further investigate this observation, strain JAM1 was cultured under anoxic conditions with NO₃⁻, and NO₃⁻, NO₂⁻ and N₂O were measured (Fig. 3A). Complete NO₃⁻ reduction

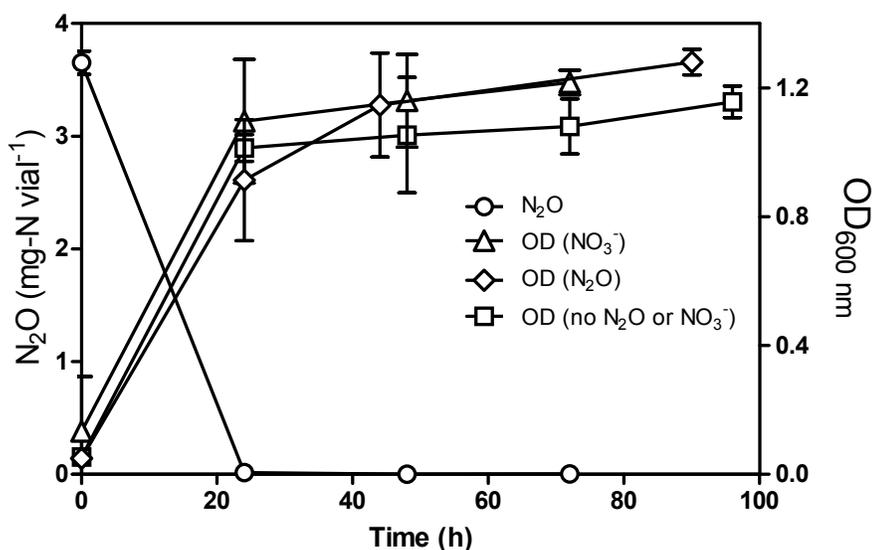


Figure 2 N₂O consumption by *Methylophaga nitratireducens* JAM1 under oxic conditions. Strain JAM1 was cultured with 3.5 mg-N vial⁻¹ N₂O, with 22 mg-N vial⁻¹ NO₃⁻ or without N₂O and NO₃⁻, and under oxic conditions. N₂O concentration was measured in N₂O-amended cultures only. Growth were measured over different time intervals. Data represent mean values ± SD ($n = 3$).

Full-size [DOI: 10.7717/peerj.4098/fig-2](https://doi.org/10.7717/peerj.4098/fig-2)

(19.3 ± 0.3 mg-N vial⁻¹) was performed within 55 h. The NO₂⁻ level reached 17.5 ± 0.2 mg-N vial⁻¹ over this period and decreased slowly to 15.9 ± 0.5 mg-N vial⁻¹. N₂O production initiated when NO₃⁻ was nearly reduced and reached 0.70 ± 0.21 mg-N vial⁻¹ after 55 h of incubation (Fig. 3A). N₂O was completely reduced after 127 h. In parallel, for cultures in which the headspace was flushed with argon, N₂ production was also measured. The corresponding results show an increase of N₂ in the headspace (Fig. 3A) by 1.14 ± 0.54 mg-N vial⁻¹ after 127 h, which represent $6.0 \pm 2.9\%$ of the N input. As the *Methylophaga* 1403 medium contains ferrous chloride ($216 \mu\text{mole vial}^{-1}$), N₂O production could originate from the abiotic reaction between NO₂⁻ that accumulated in the cultures and the ferrous ion (Klueglein et al., 2014). An abiotic control was performed with 18.2 mg-N vial⁻¹ NO₂⁻. N₂O was detected in the abiotic control after 20 h and reached 0.00172 ± 0.00012 mg-N vial⁻¹ after 114 h, which is 407 times lower than the N₂O concentration measured in the anoxic cultures. This results showed that the abiotic reaction generated negligible amount of N₂O.

Under oxic conditions, NO₃⁻ reduction (17.4 ± 2.1 mg-N vial⁻¹) was complete after 24 h with equivalent NO₂⁻ accumulation (17.1 ± 1.3 mg-N vial⁻¹). N₂O production started after complete NO₃⁻ reduction (Fig. 3B) and increased to reach 0.31 ± 0.32 mg-N vial⁻¹ after 96 h of incubation (1.7% of N input). Unlike trends observed for the anoxic cultures, N₂O concentration did not decrease in the oxic cultures. N₂O production and consumption could have reached an equilibrium and loss of nitrogen would occur by N₂ production.

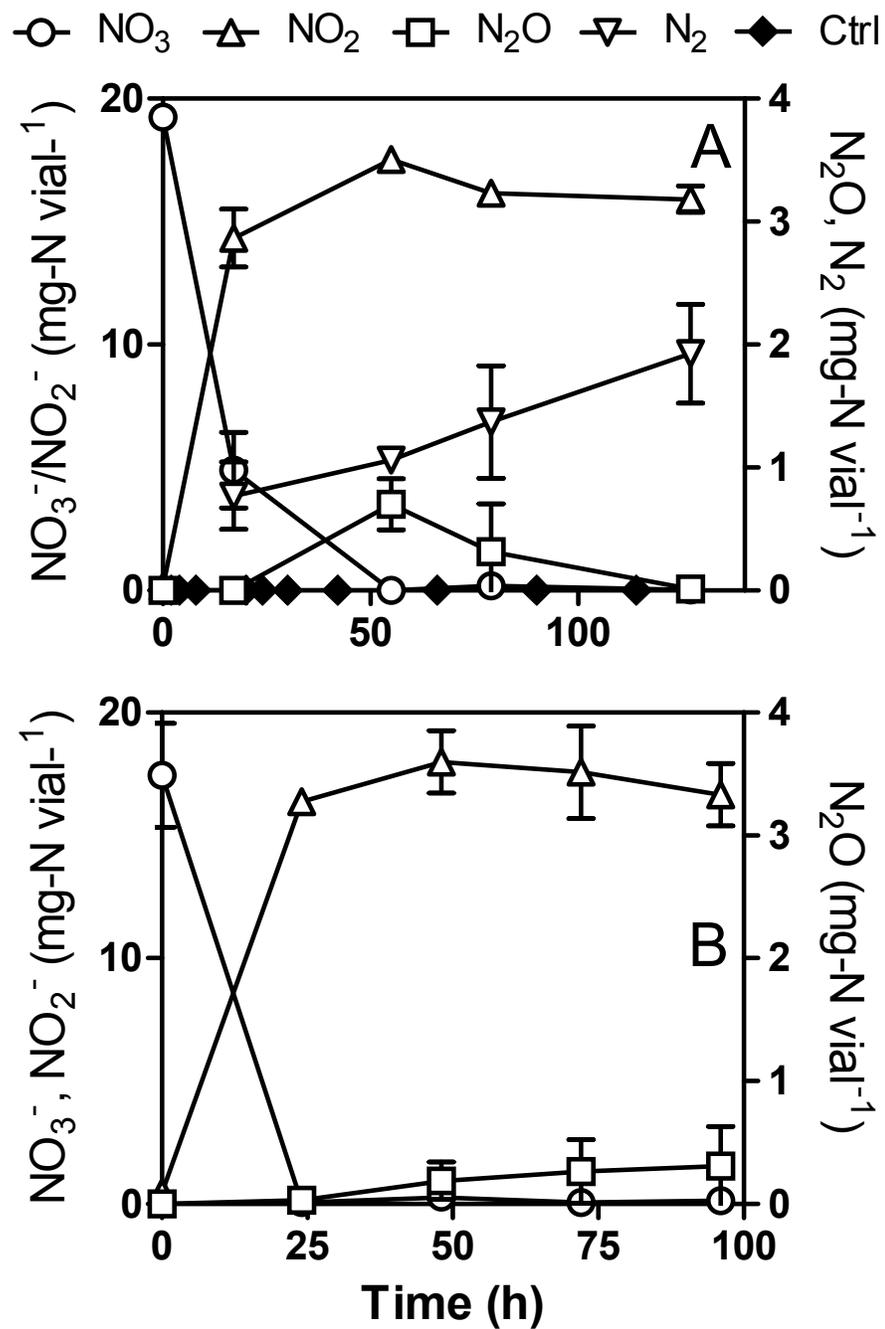


Figure 3 N_2O production by *Methylophaga nitratreducentescens* JAM1. Strain JAM1 was cultured under anoxic (A) or oxic (B) conditions with NO_3^- (22 mg-N vial $^{-1}$). NO_3^- , NO_2^- and N_2O concentrations were measured over different time intervals. N_2 concentration was measured in anoxic cultures that were flushed with argon. Ctrl: Abiotic control with 18.2 mg-N vial $^{-1}$ NO_2^- . Data represent mean values \pm SD ($n=3$).

Full-size DOI: 10.7717/peerj.4098/fig-3

NO_3^- , NO_2^- and N_2O dynamics in NH_4Cl -free cultures

The original 1403 medium recommended by the ATCC for culturing *Methylophaga* spp. contains 20.9 mg-N vial⁻¹ NH_4Cl and 0.1 mg-N vial⁻¹ ferric ammonium citrate (see ‘Material and Methods’). Based on the deduced nitrogen metabolic pathways from strain JAM1 genome (Fig. S1), N-assimilation into the biomass should proceed directly from NH_4^+ and minimal NO_3^- reduction to NH_4^+ would be occurring. For the next set of experiments, we aimed to determine the effect of removing NH_4Cl , which provides most of the NH_4^+ (99.5%), on the dynamics of NO_3^- , NO_2^- and N_2O . We hypothesized that forcing strain JAM1 to reroute some NO_3^- for N assimilation would affect denitrification and thus growth rates. Strain JAM1 was cultured with ca. 20 mg-N vial⁻¹ NO_3^- under anoxic or oxic conditions in NH_4Cl -free medium (Fig. 4). Growth pattern observed under anoxic conditions was similar between the regular and NH_4Cl -free cultures, as also the growth pattern under oxic conditions between the regular and NH_4Cl -free cultures.

Under anoxic NH_4Cl -free conditions, full NO_3^- reduction (19.1 ± 0.6 mg-N vial⁻¹) occurred within 48 h (Fig. 4A). The N_2O profile found was similar to that observed in regular cultures (Fig. 3A), though lower N_2O concentrations were detected during the accumulating phase. The NO_2^- level reached 18.5 ± 0.8 mg-N vial⁻¹ after 24 h and then slowly decreased to 12.8 ± 0.5 mg-N vial⁻¹ after 96 h. Cultures flushed with argon showed an increase of N_2 in the headspace (Fig. 4A) by 0.90 ± 0.28 mg-N vial⁻¹ after 127 h, which is similar to N_2 production in the regular culture medium. Nitrogen assimilation by the biomass could account for the difference in nitrogen mass balance (28.3%).

Unlike the cultures in regular medium (Fig. 3B), NO_3^- (21.3 ± 1.0 mg-N vial⁻¹) was not completely reduced under oxic NH_4Cl -free conditions, and it stopped after 24 h at 2.9 ± 2.7 mg-N vial⁻¹ (Fig. 4B). In conjunction with NO_3^- reduction, NO_2^- levels stopped accumulating at 13.0 ± 2.6 mg-N vial⁻¹ after 24 h. N_2O was observed after 48 h of incubation (Fig. 4B), after which it slowly accumulated and reached a concentration of 0.043 ± 0.048 mg-N vial⁻¹. This level is seven times lower than that of the regular culture medium (Fig. 3B).

To assess whether N_2O could have been generated through NH_4^+ , strain JAM1 was cultured under anoxic conditions with 22 mg-N vial⁻¹ NO_3^- , 20.7 mg-N vial⁻¹ $^{15}\text{NH}_4^+$, and acetylene to prevent the reduction of N_2O to N_2 . If NH_4^+ is involved in N_2O production, high proportion of labelled N_2O is expected. If NH_4^+ is not involved in N_2O production, we expected the production of labeled N_2O to be derived from $^{15}\text{NO}_3^-$ naturally present in NaNO_3^- at a natural $^{15}\text{N}/^{14}\text{N}$ isotopic ratio of 0.0036765. In the $^{15}\text{NH}_4^+$ -amended cultures, the $^{45}[\text{N}_2\text{O}]/^{44}[\text{N}_2\text{O}]$ and $^{46}[\text{N}_2\text{O}]/^{44}[\text{N}_2\text{O}]$ ratios measured were 0.008 and 0.0165, respectively, with an $^{15}\text{N}/^{14}\text{N}$ isotopic ratio of 0.020418. As a control, strain JAM1 cultured under anoxic conditions with $^{15}\text{NO}_3^-$ in NH_4Cl -free medium with acetylene showed, as was expected, all N_2O recovered in $^{46}[\text{N}_2\text{O}]$. Because low $^{15}\text{N}/^{14}\text{N}$ isotopic ratio were found in the $^{15}\text{NH}_4^+$ -amended cultures, our results suggest that N_2O do not proceed through NH_4^+ .

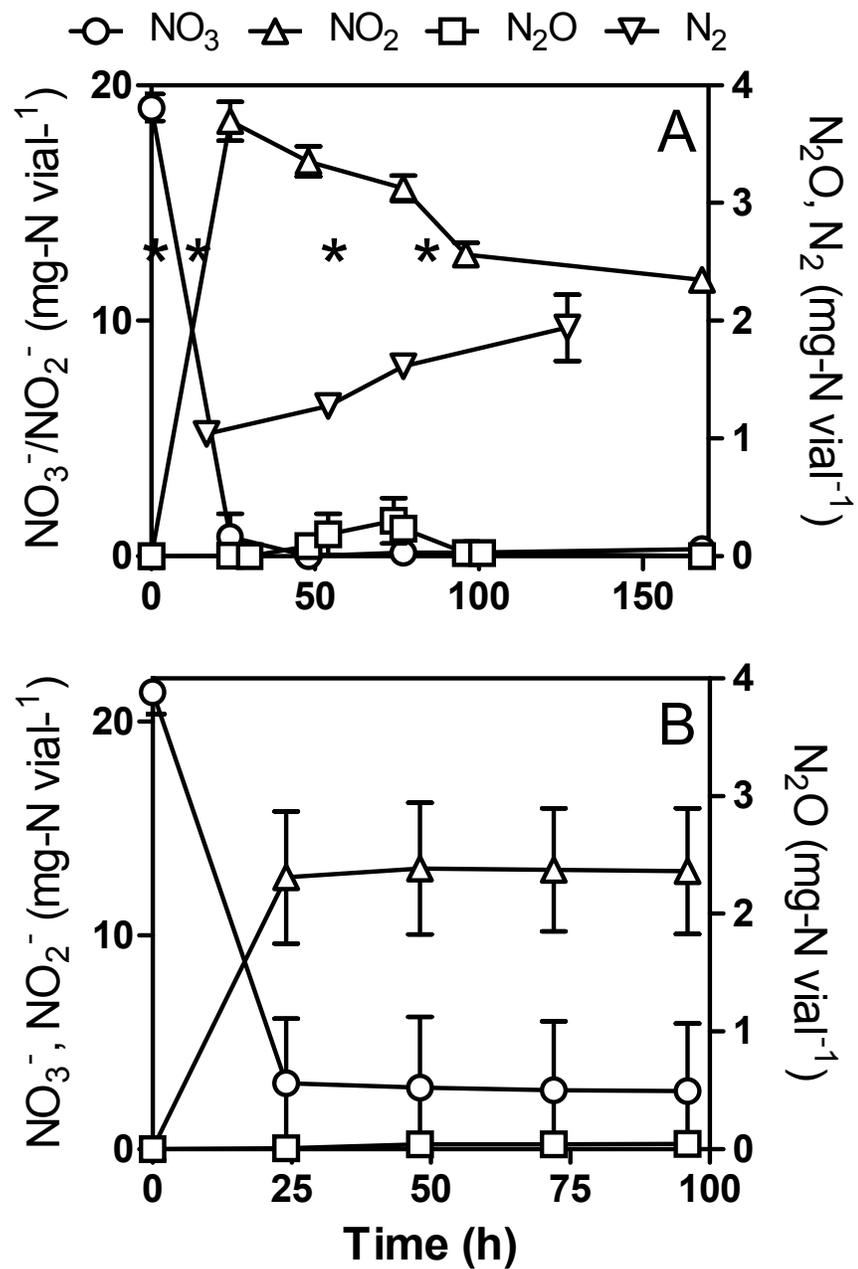


Figure 4 NO₃⁻, NO₂⁻ and N₂O dynamics by *Methylophaga nitratireducentiscrescens* JAM1 in NH₄Cl-free cultures. Strain JAM1 was cultured under anoxic (A) or oxic (B) conditions with NO₃⁻ (22 mg-N vial⁻¹) in NH₄Cl-free 1403 medium. NO₃⁻, NO₂⁻ and N₂O concentrations were measured over different time intervals. N₂ concentration was measured in anoxic cultures that were flushed with argon. The results are derived from triplicate cultures. In (A) asterisks denote the sampling times used for RNA extraction (see Fig. 6). Data represent mean values ± SD ($n = 3$).

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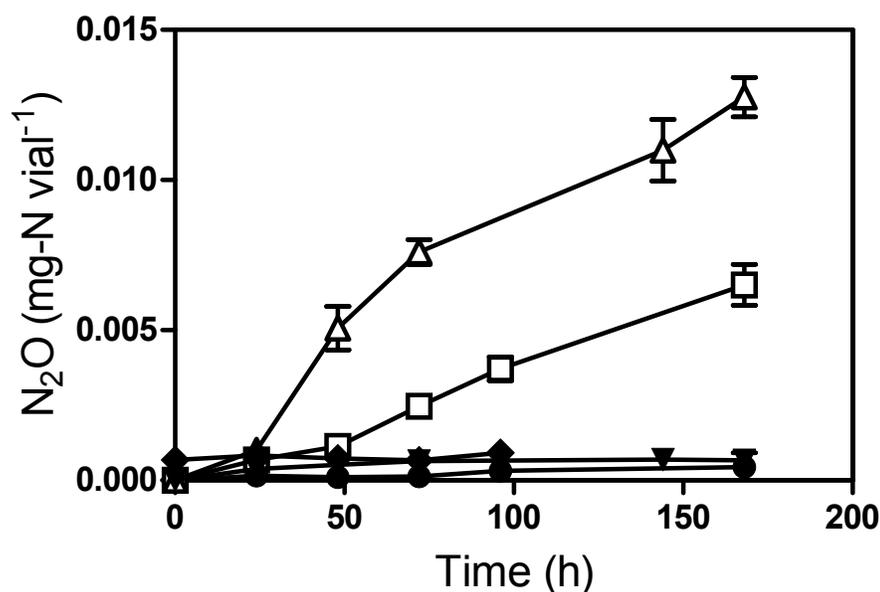


Figure 5 Reduction of NO to N₂O by *Methylophaga nitratireducens* JAM1. Strain JAM1 was cultured under oxic conditions without NO₃⁻ and with 2 mM (square), with 5 mM (triangle), or with no (circle) sodium nitroprusside (SNP). N₂O concentrations were measured over different time intervals. Controls with 5 mM SNP in non-inoculated culture medium (reverse triangle) and in culture medium inoculated with autoclaved biomass (diamond) were also performed. Data represent mean values ± SD (*n* = 3).

Full-size DOI: 10.7717/peerj.4098/fig-5

NO reduction by *M. nitratireducens* JAM1

To verify NO reduction by strain JAM1, N₂O generation was monitored in cultures without NO₃⁻ and supplemented with sodium nitroprusside hypochloride (SNP) used as an NO donor (Moore et al., 2004). Because N₂O is quickly reduced under anoxic conditions but accumulates under oxic conditions, these assays were performed under oxic conditions (Fig. 5). N₂O started to accumulate in both 2 mM and 5 mM SNP-supplemented media after 24 h of incubation, reaching $7.9 \pm 0.5 \mu\text{g-N vial}^{-1}$ and $14.5 \pm 0.4 \mu\text{g-N vial}^{-1}$, respectively, after 168 h. No N₂O production was observed in strain JAM1 cultures without SNP or in the controls with non-inoculated culture medium supplemented with SNP or inoculated with autoclaved biomass.

Role of Nar systems in NO/N₂O production

In the absence of NirK or NirS, N₂O could have been generated via NO by the Nar system (see ‘Discussion’). We used the JAM1 $\Delta narG1 narG2$ double mutant, which lacks functional Nar-type nitrate reductases and which cannot grow under anoxic conditions (Mauffrey, Martineau & Villemur, 2015). Strain JAM1 and the JAM1 $\Delta narG1 narG2$ were cultured with $16.8 \text{ mg-N vial}^{-1}$ NO₃⁻ under oxic conditions. The growth of strain JAM1 and the mutant was similar (Mauffrey, Martineau & Villemur, 2015). After 96 h of incubation, strain JAM1 completely reduced NO₃⁻ to NO₂⁻ and produced $0.14 \text{ mg-N vial}^{-1}$ of N₂O (Table 2). As

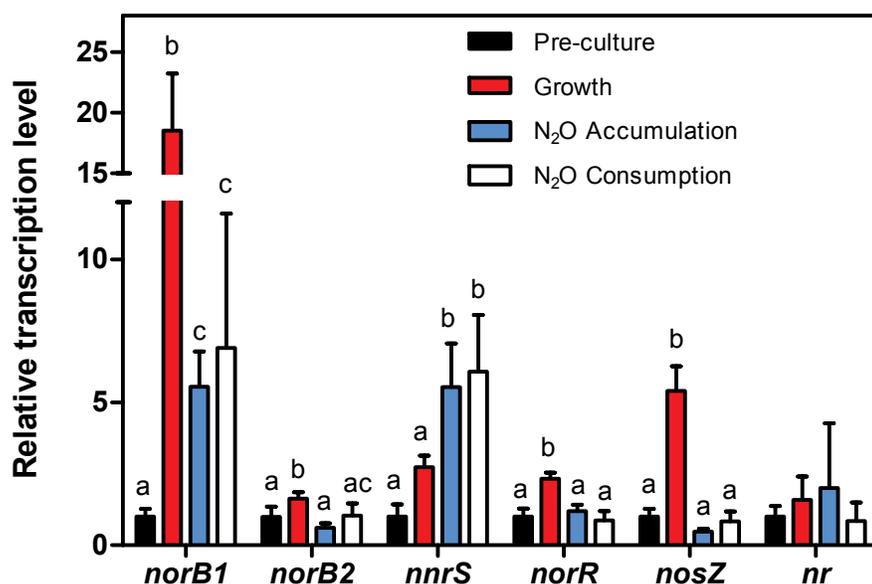


Figure 6 Relative transcript levels of *cnorB1*, *cnorB2*, *nnrS*, *nosZ*, *norR* and the assimilatory nitrate reductase (*nr*). Strain JAM1 was cultured under anoxic conditions in NH_4Cl -free 1403 medium with $22 \text{ mg-N vial}^{-1} \text{NO}_3^-$. Growth patterns were similar to those shown in Fig. 1B under the same conditions with regular 1403 medium. Samples were drawn from the pre-cultures (2–3 day old oxic cultures with no NO_3^-) and during the growth phase (NO_3^- reduction), N_2O accumulation phase, and N_2O consumption phase (see Fig. 4A), from which total RNA was extracted for RT-qPCR assays. Changes in the levels of *cnorB1*, *cnorB2*, *nnrS*, *nosZ*, *norR* and *nr* transcripts were calculated relative to their expression during the pre-culture phase (set to one, black column). One-way ANOVA tests with Tukey *post hoc* tests were performed within each phase. Columns represented by different letters are significantly different ($P < 0.05$). Data represent mean values \pm SD.

Full-size DOI: 10.7717/peerj.4098/fig-6

Table 2 Production of N_2O by strain JAM1 and the JAM1 $\Delta narG1 narG2$ double mutant. Concentrations of NO_3^- , NO_2^- and N_2O were measured after 96 h ($\text{OD}_{600\text{nm}} \sim 1.2$) of incubation in strain JAM1 and JAM1 $\Delta narG1 narG2$ cultured under oxic conditions with (A) $16.8 \text{ mg-N vial}^{-1} \text{NO}_3^-$ added at T0h or (B) $4.7 \text{ mg-N vial}^{-1} \text{NO}_2^-$ added at T24h. The results are derived from triplicate cultures. Data represent mean values (SD) ($n = 3$).

Strain	Conditions	NO_3^- (mg-N vial ⁻¹)	NO_2^- (mg-N vial ⁻¹)	N_2O (mg-N vial ⁻¹)
JAM1	A	0.17 ± 0.06	16.6 ± 0.7	0.14 ± 0.01
JAM1 $\Delta narG1 narG2$	A	17.1 ± 0.1	0.22 ± 0.22	0.004 ± 0.002
JAM1	B	0	4.25 ± 0.09	0.11 ± 0.03
JAM1 $\Delta narG1 narG2$	B	0	4.87 ± 0.39	0.18 ± 0.02

was expected, NO_3^- was not reduced, and NO_2^- was not produced by JAM1 $\Delta narG1 narG2$. Contrary to the wild type strain, the mutant did not produce N_2O .

The influence of NO_2^- was also tested. As the toxicity of NO_2^- has been attested from 0.36 mM ($0.2 \text{ mg-N vial}^{-1}$) (Auclair *et al.*, 2010), strain JAM1 and the mutant were cultured without NO_3^- under oxic conditions to allow for biomass growth. After 24 h, $4.7 \text{ mg-N vial}^{-1} \text{NO}_2^-$ was added to the cultures and was incubated for another 72 h. Strain JAM1 and the mutant produced $0.11 \text{ mg-N vial}^{-1}$ and $0.18 \text{ mg-N vial}^{-1}$ of N_2O , respectively,

reflecting N_2O concentrations produced by strain JAM1 under oxic conditions with NO_3^- (Table 2). Our results show that NO_2^- and not NO_3^- is directly involved in N_2O production, and the Nar systems are not involved in N_2O production via NO.

Relative expression levels of denitrification genes in *M. nitratireducentis* JAM1

We assessed whether variations in the expression levels of denitrification genes correlate with the N_2O accumulation and consumption cycles of strain JAM1 cultures. Strain JAM1 was cultured in NH_4Cl -free medium with $22 \text{ mg-N vial}^{-1} \text{ NO}_3^-$ under anoxic conditions. RNA was extracted from cells harvested over four different phases (Fig. 4): (1) at T0 for the pre-cultures (oxic cultures with no NO_3^-), (2) during the growth phase with NO_3^- reduction and no N_2O accumulation, (3) during the N_2O accumulation phase, and (4) during the N_2O consumption phase. The transcript levels of *cnorB1*, *cnorB2* and *nosZ*, which encode the catalytic subunits of the corresponding NO and N_2O reductases, and *nnrS* and *norR*, were measured by RT-qPCR. *nnrS* and *norR* encode NO-sensitive regulators and were used as an indicator of the presence of NO in the cultures. Because the assimilatory nitrate reductase is involved in the re-routing of NO_3^- to the biomass, RT-qPCR assays were also performed on the gene encoding this reductase (named here *nr*). The expression levels were calculated relative to the transcript levels measured during the preculture phase (set to one) (Fig. 6).

The relative *cnorB1* transcript levels showed an 18.5-fold increase during the growth phase. *cnorB1* expression was still upregulated during the N_2O accumulation and consumption phases (5.5 and 6.9-fold increases, respectively). The relative *cnorB2* transcript levels had a 1.6-fold increase during the growth phase. These levels returned nearly to the same levels of those in the preculture phase. Significant increases (5.5- and 6.0-fold) of the relative expression levels of *nnrS* were observed in the N_2O accumulation and consumption phases. *norR* was upregulated (2.3-fold increase) during the growth phase. The *nosZ* expression levels had a 5.4-fold increase during the growth phase relative to the preculture phase, and decreased to the preculture levels during the N_2O accumulation phase. No significant difference was observed in the relative transcript levels of the *nr* gene between all phases.

DISCUSSION

Our results show that *M. nitratireducentis* JAM1 can consume NO and N_2O via the mechanism of reduction of NO to N_2O and then to N_2 as predicted by the genome sequence (Fig. S1) (Villeneuve et al., 2013; Mauffrey, Martineau & Villemur, 2015). The N_2O -amended cultures yielded equivalent biomass results to those of the NO_3^- -amended cultures as predicted by the respiratory electron transport chains of the denitrification pathway (Simon, 2011). Therefore, in addition of reducing NO_3^- , strain JAM1 has another respiratory capacity under anoxic conditions by reducing N_2O for its growth.

Although denitrification is generally an anaerobic process, there are cases where it occurs under oxic conditions (Otani, Hasegawa & Hanaki, 2004). As observed with NO_3^- reduction, NO and N_2O reduction can occur under oxic conditions, reinforcing the

lack of a functional oxygen regulation response in strain JAM1 (Mauffrey, Martineau & Villemur, 2015). However, there is little benefit to this consumption, as N_2O or NO_3^- amended cultures have a similar growth pattern than the cultures with only oxygen as terminal electron acceptor. The methylamine-utilizing bacterium *Methylothermobacter mobilis* strain JLW8 also showed denitrifying activities under oxic conditions (Kalyuzhnaya et al., 2009). This freshwater bacterium has an incomplete denitrification pathway with gene clusters encoding a periplasmic NapA-type nitrate reductase, NirK and cNor. Although no growth were recorded in methylamine-amended culture supplemented with NO_3^- under oxic conditions, growth occurred in methanol-amended cultures with reduction of NO_3^- to N_2O . Contrary to *M. nitratireducenticrescens* JAM1, *Methylothermobacter mobilis* JLW8 cannot grow under anoxic conditions with NO_3^- (Kalyuzhnaya et al., 2006; Mustakhimov et al., 2013). Denitrification enzymes were showed to be active under anoxic conditions but oxygen is required for strain JLW8 growth (Kalyuzhnaya et al., 2009).

N_2O production was observed in NO_3^- -amended cultures either under oxic or anoxic conditions when NO_2^- was accumulating. This production represented up to 6% of N-input in the anoxic cultures, and NO_2^- was shown to be the key element of this production process. Because we showed that the NO reductase activities were carried out in strain JAM1 cultures, the N_2O could originate from NO production despite the absence of gene encoding NirS or NirK. Intermediate NO creates problems as this molecule is highly toxic to microorganisms, inducing nitrosative stress in cells (Poole, 2005). Reducing NO is a key step in denitrification and is closely regulated by various sensors and regulators. NnrS and NorR are involved in cell defense against nitrosative stress and are positively regulated by the presence of NO (Stern et al., 2013; Stern et al., 2012; Bartnikas et al., 2002). Therefore, the expression of *nnrS* and *norR* reflects NO concentrations in a medium and was used as a marker of NO presence. The upregulation of the expression of *norR* during the growth phase strongly suggest that NO is produced during this phase. This correlates with higher expressions of both *cnorB*, and *nosZ* during the growth phase, which can be regulated by NO-sensitive regulators such as NorR (Spiro, 2012). *nnrS* is upregulated during the N_2O accumulation and consumption phases, which suggests that NO is still generated during these phases. This upregulation can be linked to the decrease of the relative transcript levels of both *cnorB*, and of *nosZ* and *norR*, but also to the highest level of NO_2^- concentrations in the culture medium. It is therefore possible that NO is not reduced quickly enough in the cells by cNorB and strain JAM1 must rely on another mechanism, such as NnrS, to protect itself from NO toxicity. Stern et al. (2013) suggest that NnrS does not remove NO directly, but protects cells against the formation of iron-NO complexes, which are inhibitory to iron-sulfur cluster proteins. Moreover, Vaccaro et al. (2016) proposed that NnrS senses NO and signals to cytoplasmic transcription factors or Fe-S cluster repair proteins.

Other NO_3^- respiring bacteria that lack NirK or NirS have been shown to be N_2O producers (Bleakley & Tiedje, 1982; Smith & Zimmerman, 1981; Sun, Vos & Heylen, 2016). For instance, *Bacillus vireti* contains three denitrification reductases (Nar, qCu_ANor, N₂OR) and lacks, like *M. nitratireducenticrescens* JAM1, gene encoding NirK or NirS (Mania et al., 2014). This bacterium also produces NO and N_2O in anaerobic, NO_3^- -amended TSB cultures during NO_2^- accumulation. NO was shown to originate from chemical

decomposition of NO_2^- (Schreiber *et al.*, 2012) and from an unknown biotic reaction. In our study, the abiotic control of the *Methylophaga* 1403 medium amended with NO_3^- did not show significant N_2O production. Also, the abiotic reaction between the ferrous ion in the medium and NO_2^- generated negligible amount of N_2O . Furthermore, no N_2O was detected in this medium inoculated with autoclaved biomass (Fig. 5). These results rule out abiotic reactions as the source of the total amount of N_2O produced in strain JAM1 cultures. The possible biotic source of NO in absence of NirS or NirK has been studied in *Escherichia coli* (see review by Vine & Cole, 2011). There are supporting evidence that NO is generated in *E. coli* as a side product during NO_2^- reduction (i) by the cytoplasmic, NADH-dependent nitrite reductase (NirBD), (ii) by the nitrite reductase NrfAB, and (iii) by NarGHI. Vine, Purewal & Cole (2011) showed, with mutants defective in these reductases, that NarGHI is the major enzyme responsible of NO production. However, a small production of NO was still occurring in *narG* mutant, suggesting the involvement of another molybdoprotein. In *M. nitratireducentescens* JAM1, the double-knockout mutant JAM1 Δ *narG1narG2*, which lacks the two dissimilatory NO_3^- reductases, was still able to produce N_2O under oxic conditions at the same level of the wild type when NO_2^- was added to the cultures. These results suggest the two Nar systems are not involved in NO production. The genome of strain JAM1 did not reveal gene encoding NrfAB, but contain a gene cluster encoding a cytoplasmic, NADH-dependent nitrite reductase (CP003390.3; Q7A_2620 and Q7A_2621), which may be the source of NO (Fig. S1). In the latter case, this could be verified by generating a knockout mutant of this gene.

The significance of maintaining an incomplete pathway by *M. nitratireducentescens* JAM1 is unclear and may depend upon the original habitat and environment, here the denitrifying biofilm. While *M. nitratireducentescens* JAM1 serves as an important actor among the microbial community of the marine biofilm in performing optimal denitrifying activities (Labbé *et al.*, 2003; Labbé, Parent & Villemur, 2003), it was thought to participate uniquely in the reduction of NO_3^- to NO_2^- . It was previously proposed that NO_2^- reduction to N_2 is carried out by *Hyphomicrobium nitrativorans* NL23, the second most represented bacterium in the biofilm (Labbé *et al.*, 2007; Auclair, Parent & Villemur, 2012). Its capacity to reduce NO and N_2O and to grow on N_2O suggests that *M. nitratireducentescens* JAM1 may participate in the reduction of NO and N_2O during denitrification in the biofilm. Although our culture assays were performed with high levels of NO_3^- (37 mM), which rarely exceeds a value of 0.7 mM in natural environments (Yeats, 1990), similar levels can be reached in closed-circuit systems like the seawater aquarium tank located in the Montreal Biodome, where NO_3^- levels reached up to 14 mM (Parent & Morin, 2000). Rissanen *et al.* (2016) also observed the combination of *Methylophaga* spp. and *Hyphomicrobium* spp. in the fluidized-bed type denitrification reactors treating the recirculating seawater of the public fish aquarium SEA LIFE at Helsinki, Finland. Although this study provided no indication of the denitrification pathway in these *Methylophaga* and *Hyphomicrobium*, it reinforces the importance of the natural combination of these two genera in marine denitrification environment.

CONCLUSIONS

M. nitratireducenticrescens JAM1 is one of few isolated marine methylotrophic bacterial strains to exhibit anaerobic respiratory capacities by reducing NO_3^- to NO_2^- and, as reported here, by reducing N_2O to N_2 . It can also generate N_2O via NO by an unknown biotic system. Very few marine denitrifying bacteria have been isolated from recirculating marine systems (Borges *et al.*, 2008; Foessel, Drake & Schramm, 2011; Zheng *et al.*, 2012; Zheng *et al.*, 2011). No previous studies have generated genetic information related gene arrangement or expression on these bacteria. Based on substantial data accumulated on the genome, gene arrangement and gene expression of denitrification and on methylotrophy, *M. nitratireducenticrescens* JAM1 can serve as a model for studying such activities in marine environments. Finally, our results enable a better understanding of the ecophysiological role of *M. nitratireducenticrescens* JAM1 in the original biofilm developed in the denitrification reactor of a closed-circuit marine aquarium.

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ADDITIONAL INFORMATION AND DECLARATIONS

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Competing Interests

The authors declare there are no competing interests.

Author Contributions

- Florian Mauffrey conceived and designed the experiments, performed the experiments, analyzed the data, wrote the paper, prepared figures and/or tables, reviewed drafts of the paper.
- Alexandra Cucaita performed the experiments, analyzed the data.
- Philippe Constant conceived and designed the experiments, analyzed the data, contributed reagents/materials/analysis tools, reviewed drafts of the paper.
- Richard Villemur conceived and designed the experiments, analyzed the data, contributed reagents/materials/analysis tools, wrote the paper, prepared figures and/or tables, reviewed drafts of the paper.

Data Availability

The following information was supplied regarding data availability:

The raw data was provided as a [Supplemental File](#).

Supplemental Information

Supplemental information for this article can be found online at <http://dx.doi.org/10.7717/peerj.4098#supplemental-information>.

REFERENCES

- Auclair J, Lepine F, Parent S, Villemur R. 2010.** Dissimilatory reduction of nitrate in seawater by a *Methylophaga* strain containing two highly divergent *narG* sequences. *The ISME Journal* **4**:1302–1313 DOI [10.1038/ismej.2010.47](https://doi.org/10.1038/ismej.2010.47).
- Auclair J, Parent S, Villemur R. 2012.** Functional diversity in the denitrifying biofilm of the methanol-fed marine denitrification system at the Montreal Biodome. *Microbial Ecology* **63**:726–735 DOI [10.1007/s00248-011-9960-2](https://doi.org/10.1007/s00248-011-9960-2).
- Bartnikas TB, Wang YS, Bobo T, Veselov A, Scholes CP, Shapleigh JP. 2002.** Characterization of a member of the NnrR regulon in *Rhodobacter sphaeroides* 2.4.3 encoding a haem-copper protein. *Microbiology* **148**:825–833.
- Bleakley BH, Tiedje JM. 1982.** Nitrous oxide production by organisms other than nitrifiers or denitrifiers. *Applied and Environmental Microbiology* **44**:1342–1348.
- Boden R. 2012.** Emended description of the genus *Methylophaga* Janvier et al., 1985. *International Journal of Systematic and Evolutionary Microbiology* **62**:1644–1646 DOI [10.1099/ijs.0.033639-0](https://doi.org/10.1099/ijs.0.033639-0).
- Borges MT, Sousa A, De Marco P, Matos A, Hönigová P, Castro PML. 2008.** Aerobic and anoxic growth and nitrate removal capacity of a marine denitrifying bacterium isolated from a recirculation aquaculture system. *Microbial Ecology* **55**:107–118 DOI [10.1007/s00248-007-9256-8](https://doi.org/10.1007/s00248-007-9256-8).
- Einsle O, Kroneck PM. 2004.** Structural basis of denitrification. *Biological Chemistry* **385**:875–883 DOI [10.1515/BC.2004.115](https://doi.org/10.1515/BC.2004.115).
- Foesel BU, Drake HL, Schramm A. 2011.** *Defluviimonas denitrificans* gen. nov., sp. nov., and *Pararhodobacter aggregans* gen. nov., sp. nov., non-phototrophic *Rhodobacteraceae* from the biofilter of a marine aquaculture. *Systematic and Applied Microbiology* **34**:498–502 DOI [10.1016/j.syapm.2011.08.006](https://doi.org/10.1016/j.syapm.2011.08.006).
- Kalyuzhnaya MG, Bowerman S, Lara JC, Lidstrom ME, Chistoserdova L. 2006.** *Methylotenera mobilis* gen. nov., sp. nov., an obligately methylamine-utilizing bacterium within the family *Methylophilaceae*. *International Journal of Systematic and Evolutionary Microbiology* **56**:2819–2823 DOI [10.1099/ijs.0.64191-0](https://doi.org/10.1099/ijs.0.64191-0).
- Kalyuzhnaya MG, Martens-Habbena W, Wang TS, Hackett M, Stolyar SM, Stahl DA, Lidstrom ME, Chistoserdova L. 2009.** *Methylophilaceae* link methanol oxidation to denitrification in freshwater lake sediment as suggested by stable isotope probing and pure culture analysis. *Environmental Microbiology Reports* **1**:385–392 DOI [10.1111/j.1758-2229.2009.00046.x](https://doi.org/10.1111/j.1758-2229.2009.00046.x).

- Kaspar HF. 1982.** Nitrite reduction to nitrous-oxide by *Propionibacteria*—Detoxication Mechanism. *Archives of Microbiology* **133**:126–130 DOI [10.1007/BF00413525](https://doi.org/10.1007/BF00413525).
- Klemedtsson L, Svensson BH, Lindberg T, Rosswall T. 1977.** Use of acetylene inhibition of nitrous-oxide reductase in quantifying denitrification in soils. *Swedish Journal of Agricultural Research* **7**:179–185.
- Klueglein N, Zeitvogel F, Stierhof YD, Floetenmeyer M, Konhauser KO, Kappler A, Obst M. 2014.** Potential role of nitrite for abiotic Fe(II) oxidation and cell encrustation during nitrate reduction by denitrifying bacteria. *Applied and Environmental Microbiology* **80**:1051–1061 DOI [10.1128/AEM.03277-13](https://doi.org/10.1128/AEM.03277-13).
- Labbé N, Juteau P, Parent S, Villemur R. 2003.** Bacterial diversity in a marine methanol-fed denitrification reactor at the Montreal Biodome, Canada. *Microbial Ecology* **46**:12–21 DOI [10.1007/s00248-002-1056-6](https://doi.org/10.1007/s00248-002-1056-6).
- Labbé N, Laurin V, Juteau P, Parent S, Villemur R. 2007.** Microbiological community structure of the biofilm of a methanol-fed, marine denitrification system, and identification of the methanol-utilizing microorganisms. *Microbial Ecology* **53**:621–630 DOI [10.1007/s00248-006-9168-z](https://doi.org/10.1007/s00248-006-9168-z).
- Labbé N, Parent S, Villemur R. 2003.** Addition of trace metals increases denitrification rate in closed marine systems. *Water Research* **37**:914–920 DOI [10.1016/S0043-1354\(02\)00383-4](https://doi.org/10.1016/S0043-1354(02)00383-4).
- Laurin V, Labbé N, Parent S, Juteau P, Villemur R. 2008.** Microeukaryote diversity in a marine methanol-fed fluidized denitrification system. *Microbial Ecology* **56**:637–648 DOI [10.1007/s00248-008-9383-x](https://doi.org/10.1007/s00248-008-9383-x).
- Mania D, Heylen K, Van Spanning RJM, Frostegard A. 2014.** The nitrate-ammonifying and *nosZ*-carrying bacterium *Bacillus vireti* is a potent source and sink for nitric and nitrous oxide under high nitrate conditions. *Environmental Microbiology* **16**:3196–3210 DOI [10.1111/1462-2920.12478](https://doi.org/10.1111/1462-2920.12478).
- Mauffrey F, Martineau C, Villemur R. 2015.** Importance of the two dissimilatory (Nar) nitrate reductases in the growth and nitrate reduction of the methylotrophic marine bacterium *Methylophaga nitratreducenticrescens* JAM1. *Frontiers in Microbiology* **6**:1475 DOI [10.3389/fmicb.2015.01475](https://doi.org/10.3389/fmicb.2015.01475).
- Moore CM, Nakano MM, Wang T, Ye RW, Helmann JD. 2004.** Response of *Bacillus subtilis* to nitric oxide and the nitrosating agent sodium nitroprusside. *Journal of Bacteriology* **186**:4655–4664 DOI [10.1128/JB.186.14.4655-4664.2004](https://doi.org/10.1128/JB.186.14.4655-4664.2004).
- Mustakhimov I, Kalyuzhnaya MG, Lidstrom ME, Chistoserdova L. 2013.** Insights into denitrification in *Methylothenera mobilis* from denitrification pathway and methanol metabolism mutants. *Journal of Bacteriology* **195**:2207–2211 DOI [10.1128/JB.00069-13](https://doi.org/10.1128/JB.00069-13).
- Otani Y, Hasegawa K, Hanaki K. 2004.** Comparison of aerobic denitrifying activity among three cultural species with various carbon sources. *Water Science and Technology* **50**:15–22.
- Parent S, Morin A. 2000.** N budget as water quality management tool in closed aquatic mesocosms. *Water Research* **34**:1846–1856 DOI [10.1016/S0043-1354\(99\)00343-7](https://doi.org/10.1016/S0043-1354(99)00343-7).

- Poole R. 2005.** Nitric oxide and nitrosative stress tolerance in bacteria. *Biochemical Society Transactions* **33**:176–180 DOI [10.1042/BST0330176](https://doi.org/10.1042/BST0330176).
- Rissanen AJ, Ojala A, Dernjatin M, Jaakkola J, Tiirola M. 2016.** *Methylophaga* and *Hyphomicrobium* can be used as target genera in monitoring saline water methanol-utilizing denitrification. *Journal of Industrial Microbiology & Biotechnology* **43**:1647–1657 DOI [10.1007/s10295-016-1839-2](https://doi.org/10.1007/s10295-016-1839-2).
- Schreiber F, Wunderlin P, Udert KM, Wells GF. 2012.** Nitric oxide and nitrous oxide turnover in natural and engineered microbial communities: biological pathways, chemical reactions, and novel technologies. *Frontiers in Microbiology* **3**:372 DOI [10.3389/fmicb.2012.00372](https://doi.org/10.3389/fmicb.2012.00372).
- Simon J. 2011.** Organization of respiratory electron transport chains in nitrate-reducing and nitrifying bacteria. In: Moir JWB, ed. *Nitrogen cycling in bacteria*. Norfolk: Molecular Analysis Caister Academic Press, 39–58.
- Simon J, Klotz MG. 2013.** Diversity and evolution of bioenergetic systems involved in microbial nitrogen compound transformations. *BBA-Bioenergetics* **1827**:114–135 DOI [10.1016/j.bbabi.2012.07.005](https://doi.org/10.1016/j.bbabi.2012.07.005).
- Smith MS, Zimmerman K. 1981.** Nitrous-oxide production by non-denitrifying soil nitrate reducers. *Soil Science Society of America Journal* **45**:865–871 DOI [10.2136/sssaj1981.03615995004500050008x](https://doi.org/10.2136/sssaj1981.03615995004500050008x).
- Snider DM, Schiff SL, Spoelstra J. 2009.** $^{15}\text{N}/^{14}\text{N}$ and $^{18}\text{O}/^{16}\text{O}$ stable isotope ratios of nitrous oxide produced during denitrification in temperate forest soils. *Geochim Cosmochim A* **73**:877–888 DOI [10.1016/j.gca.2008.11.004](https://doi.org/10.1016/j.gca.2008.11.004).
- Spiro S. 2012.** Nitrous oxide production and consumption: regulation of gene expression by gas-sensitive transcription factors. *Philosophical Transactions of the Royal Society of London. Series B, Biological Sciences* **367**:1213–1225 DOI [10.1098/rstb.2011.0309](https://doi.org/10.1098/rstb.2011.0309).
- Stern AM, Hay AJ, Liu Z, Desland FA, Zhang J, Zhong ZT, Zhu J. 2012.** The NorR regulon is critical for *Vibrio cholerae* resistance to nitric oxide and sustained colonization of the intestines. *Mbio* **3**(2):300013–12 DOI [10.1128/mBio.00013-12](https://doi.org/10.1128/mBio.00013-12).
- Stern AM, Liu BB, Bakken LR, Shapleigh JP, Zhu J. 2013.** A novel protein protects bacterial iron-dependent metabolism from nitric oxide. *Journal of Bacteriology* **195**:4702–4708 DOI [10.1128/JB.00836-13](https://doi.org/10.1128/JB.00836-13).
- Sun YH, De Vos P, Heylen K. 2016.** Nitrous oxide emission by the non-denitrifying, nitrate ammonifier *Bacillus Licheniformis*. *BMC Genomics* **17**:68 DOI [10.1186/s12864-016-2382-2](https://doi.org/10.1186/s12864-016-2382-2).
- Vaccaro BJ, Thorgersen MP, Lancaster WA, Price MN, Wetmore KM, Poole FL, Deutschbauer A, Arkin AP, Adams MWW. 2016.** Determining roles of accessory genes in denitrification by mutant fitness analyses. *Applied and Environmental Microbiology* **82**:51–61 DOI [10.1128/AEM.02602-15](https://doi.org/10.1128/AEM.02602-15).
- Van Spanning RJM, Delgado MJ, Richardson DJ. 2005.** The nitrogen cycle: denitrification and its relationship to N_2 fixation. In: Werner D, Newton WE, eds. *Nitrogen fixation in agriculture, forestry, ecology, and the environment*. Dordrecht: Springer, 277–342.

- Villeneuve C, Martineau C, Mauffrey F, Villemur R. 2013.** *Methylophaga nitratreducenticrescens* sp. nov., and *Methylophaga frappieri* sp. nov., isolated from the biofilm of the methanol-fed denitrification system treating the seawater at the Montreal Biodome. *International Journal of Systematic and Evolutionary Microbiology* **63**:2216–2222 DOI [10.1099/ijs.0.044545-0](https://doi.org/10.1099/ijs.0.044545-0).
- Vine CE, Cole JA. 2011.** Unresolved sources, sinks, and pathways for the recovery of enteric bacteria from nitrosative stress. *FEMS Microbiology Letters* **325**:99–107 DOI [10.1111/j.1574-6968.2011.02425.x](https://doi.org/10.1111/j.1574-6968.2011.02425.x).
- Vine CE, Purewal SK, Cole JA. 2011.** NsrR-dependent method for detecting nitric oxide accumulation in the *Escherichia coli* cytoplasm and enzymes involved in NO production. *FEMS Microbiology Letters* **325**:108–114 DOI [10.1111/j.1574-6968.2011.02385.x](https://doi.org/10.1111/j.1574-6968.2011.02385.x).
- Yeats PA. 1990.** Reactivity and transport of nutrients and metals in the St. Lawrence Estuary. In: El-Sabh M, Silverberg N, eds. *Coastal and Estuarine Studies*. vol 39. New York: Springer-Verlag, 153–169.
- Zheng HY, Liu Y, Gao XY, Ai GM, Miao LL, Liu ZP. 2012.** Characterization of a marine origin aerobic nitrifying-denitrifying bacterium. *Journal of Bioscience and Bioengineering* **114**:33–37 DOI [10.1016/j.jbiosc.2012.02.025](https://doi.org/10.1016/j.jbiosc.2012.02.025).
- Zheng HY, Liu Y, Sun GD, Gao XY, Zhang QL, Liu ZP. 2011.** Denitrification characteristics of a marine origin psychrophilic aerobic denitrifying bacterium. *Journal of Environmental Sciences (China)* **23**:1888–1893 DOI [10.1016/S1001-0742\(10\)60615-8](https://doi.org/10.1016/S1001-0742(10)60615-8).
- Zumft WG. 1997.** Cell biology and molecular basis of denitrification. *Microbiology and Molecular Biology Reviews* **61**:533–616.